Supporting Information

Fischer et al. 10.1073/pnas.0900110106

SI Methods

Remotely Sensed Data. Images from the French SPOT were used to create a 10-m resolution map of tree cover. We used 31 natural color orthomosaics of 2.5-m resolution and six 10-m, 4-band multispectral images from 2004 and 2005. Each was individually classified in ENVI (ITT Visual Information Solutions) 4.3 to 15 to 20 classes by using a standard unsupervised Isodata algorithm. Using ArcGIS (ESRI), the classes for each image were then visually aggregated into 2 classes: woody and nonwoody vegetation. We mosaicked the 37 overlapping, 2-class layers to a seamless 10-m resolution layer, prioritizing input layers based on their local performance. The 10-m data effectively delineated the crowns of many individual large trees but often underpredicted tree cover in sparsely vegetated areas with small tree crowns. The 2.5-m data had more predictive power in areas with scattered trees, but they occasionally classed shadows and high-vigor crops as trees. The final mosaic captured scattered trees and woodlands in the lower slope areas very well (≈80% of the landscape, including all land dominated by farming). Performance was low only in areas of complex topography (where farming was not a dominant land use). A visual indication of the performance of the classification is given in Fig. S1. A formal assessment of the classification showed a highly significant relationship between the number of trees present on the ground and remotely sensed percent tree cover (for details, see data analysis section in *Methods*, Fig. 1*B*, and Table S1).

Tree Identification. Some species were grouped because (*i*) they were difficult to distinguish in the field and/or may hybridize (1), or (*ii*) they were uncommon in our sites. Tree species that were grouped were White Box (*Eucalyptus albens*; common) and Grey Box (*Eucalyptus microcarpa*; very uncommon; combined total individuals measured, 362); Blakely's Red Gum (*Eucalyptus blakelyi*; very common), Dwyer's Red Gum (*Eucalyptus dwyeri*; uncommon), and River Red Gum (*Eucalyptus camaldulensis*; very uncommon; combined total, 818); 2 species of Long-leaved Box (*Eucalyptus goniocalyx* and *Eucalyptus nortonii*; combined total, 449); and "other gums" (*Eucalyptus mannifera* and *Eucalyptus rossii*; combined total, 215). Additional species encountered, in decreasing order of abundance, were Red Stringybark (*Eucalyptus macrorhyncha*; 689 trees), Yellow Box (*Eucalyptus melliodora*; 581 trees), Red Box (*Eucalyptus polyanthemos*; 384 trees), Apple Box (*Eucalyptus bridgesiana*; 129 trees), and Red Ironbark (*Eucalyptus sideroxylon*; 68 trees).

Diameter Measurements. The diameter at breast height of trees was measured at 130 cm above ground. For trees with multiple stems, the diameter of each stem was measured. Before analysis, first, tree diameters were adjusted to combine measurements from single-stemmed and multistemmed trees. To account for multiple stems, the basal area at breast height of each tree was calculated. We then calculated the diameter of an equivalent single-stemmed tree with the same basal area as:

$$DBH_{\text{equiv}} = 2\sqrt{AREA_{\text{basal}}/\pi}$$

Second, tree diameters were standardized to account for different tree species growing to different diameters. Ideally, we would have estimated the age of each individual tree from its diameter, but such relationships are highly uncertain and are published only for 1 species, Yellow Box (2). A key problem is that the growth of eucalypt diameters is approximately linear until the age of 100 years, but after that it slows down (2). Assuming that different trees grow to different diameters but that the shape of their growth curves is similar as they age, we standardized diameters as follows. For each species, we identified all trees with a basal area above the 95th percentile observed for that species. We calculated the mean basal area of those trees, which acted as a proxy for the basal area of a representative "very old tree" of a given species. We converted this basal area into a diameter by using the formula above. We then scaled the diameter of each tree by dividing it by the estimated diameter of a very old tree of the same species. By definition, more than 95% of the resulting unit-free index values were between 0 and 1. For our graphs, we multiplied these unit-free index values by 165 cm, which was the estimated diameter for a very old Yellow Box tree. In summary, our procedure scaled all tree diameters to one reference species (Yellow Box) to allow effective comparison of tree diameters (and relative ages) between different sites.

We acknowledge that the same species can grow at different rates at different sites (3). However, we believe this to be a relatively minor problem for 3 reasons. First, woodlands are relatively open by definition, so that suppressed growth due to high stem density is less likely than in denser forest systems. Second, published differences in growth rates for Yellow Box between different sites were relatively minor (2), compared with much more pronounced between-species differences. Third, major differences in site conditions typically express themselves via species turnover, and we did standardize diameters for differences between species (see above).

Estimation of Tree Density. In paddock sites and scattered tree sites, all individual trees were identified and measured. The total count of trees was a direct measurement of tree density at the site.

In woodland sites, a sample of 64 trees was measured by using a distance sampling protocol. Eight "random points" were arbitrarily distributed throughout each site before field work and without reference to satellite imagery that would have revealed the spatial distribution of trees. In the field, 8 sectors of 45° each were delineated around each random point, starting at magnetic north. Within each sector, the closest tree was identified, and the distance to this tree from the random point was measured. Assuming random dispersal of stems throughout the site, the distance measurements were used to calculate the number of trees in the site, following the formula for an unbiased estimate of point-centered density in Pollard (4). We consider that random dispersal of trees was a reasonable assumption in woodland sites, because unlike in some Northern Hemisphere ecosystems (3), regeneration usually is not highly clumped in woodland sites.

At each site, based on the approximate total number of trees, the proportions of trees sampled in different diameter classes (in 20-cm intervals) and of different species were used to estimate the actual number of trees representing each different species and diameter class.

Tree Community Composition. We calculated a "tree species profile" for each site to summarize the mix of species present at each site in a single number for use as a covariate in statistical analyses. We constructed a matrix of sites by tree species, where each cell contained the estimated integer number of individuals belonging to a particular tree species at a particular site. We applied simple correspondence analysis to this matrix (5). The first principal axis had a canonical correlation of 0.76, suggesting it was a reasonable univariate summary of the underlying multivariate data. Correspondence analysis sorted species along an ecologically meaningful gradient, represented by Apple Box and Yellow Box at the one end (typical foothill species) and Red Ironbark, Red Box, and Red Stringybark at the other end (typical ridgetop species). Each site's row score was interpreted as its tree species profile.

Soil Chemistry. At each site, the same 8 random points used for tree measurements (see estimation of tree density, above) were used for topsoil sampling in early 2008. Around each random point, we located 4 canopy gaps ≈10 m from the random point and >10 m from one another. At each of the resulting 32 locations, we obtained 2-cm diameter cores to a depth of 7 cm. We obtained samples from canopy gaps whenever possible because soil nutrients are typically enriched underneath the canopy of trees (6). Therefore, locating samples without reference to canopy cover would have systematically inflated the recorded nutrient levels of woodland sites compared with the more open paddock or scattered tree sites, thus confounding the effects of tree cover with the effects of topsoil nutrients. The 32 samples from each site were mixed into a single site-level sample. Soils were air-dried for several weeks before analysis. A range of attributes were quantified, but we focus here on available phosphorus and total nitrogen. Available phosphorus was quantified by using the Colwell method (ref. 7, p. 64). Total nitrogen was quantified by using wet oxidation (the Kjeldahl method), using the Technicon Autonalyzer II, industrial method no. 329-74W/B. We acknowledge that it would have been informative to analyze soil samples for their nitrate content, rather than simply focusing on total nitrogen. Nitrate may exert a stronger influence on vegetation dynamics than total nitrogen per se (8). However, we could not quantify soil nitrate because (*i*) we were unable to obtain samples over a short enough time period that would avoid seasonal changes in nitrate levels (9), and (*ii*) given our large study area and the lack of access to reliable refrigeration, soil nitrate levels may have altered during transport and initial storage in the field.

1. Costermans L (2000) *Native Trees and Shrubs of South-eastern Australia* (Reed New Holland, Sydney).
2. Banks JCG (1997) in *The Coming of Age–Forest Age & Heritage Values*, ed Dargavel J (Environment Australia, Canberra, Australia), pp 17–28.
3. Pulido FJ, Diaz M, de Trucios SJH (2001) Size structure and regeneration of Spanish holm oak *Quercus ilex* forests and dehesas: Effects of agroforestry use on their long-term sustainability. *For Ecol Manage* 146:1–13.
4. Pollard JH (1971) On distance estimators of density in randomly distributed forests. *Biometrics* 27:991–1002.
5. Greenacre MJ (1993) *Correspondence Analysis in Practice* (Academic, London).
6. Wilson BR, Growns I, Lemon J (2007) Scattered native trees and soil patterns in grazing land on the Northern Tablelands of New South Wales, Australia. *Aust J Soil Res* 45:199–205.
7. Rayment GE, Higginson RR (1992) *Australian Laboratory Handbook of Soil and Water Chemical Methods* (Inkata Press, Melbourne).
8. Prober, SM, Lunt ID, Morgan JW (2008) in *New Models for Ecosystem Dynamics and Restoration*, eds Hobbs RJ, Suding K (Island Press, Washington, DC), pp. 156–168.
9. Prober SM, Thiele KR, Lunt ID, Koen TB (2005) Restoring ecological function in temperate grassy woodlands: Manipulating soil nutrients, exotic annuals and native perennial grasses through carbon supplements and spring burns. *J Appl Ecol* 42:1073–1085.

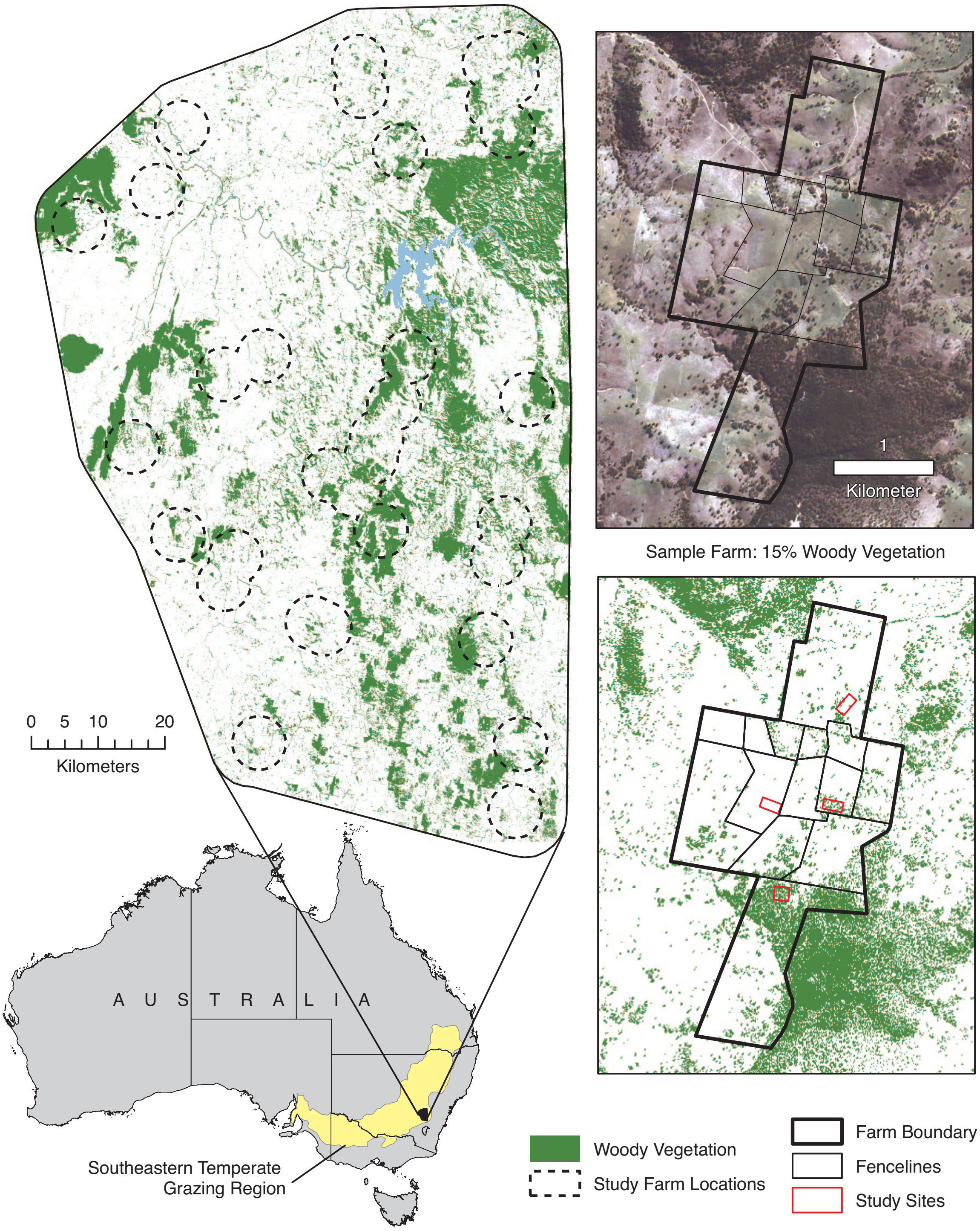

Fig. S1. Location of southeastern Australia's temperate grazing region, with our study area in the Upper Lachlan Catchment of New South Wales highlighted. The temperate grazing region is based on the Southern Temperate Beef Industry Region (adapted from www.anra.gov.au/). The approximate location of the 33 farms surveyed is shown on top of our regional classification of tree cover. To illustrate the quality of the classification, both SPOT imagery and our classified map of tree cover are shown for 1 farm. Superimposed are paddock boundaries (black) and site perimeters (red).

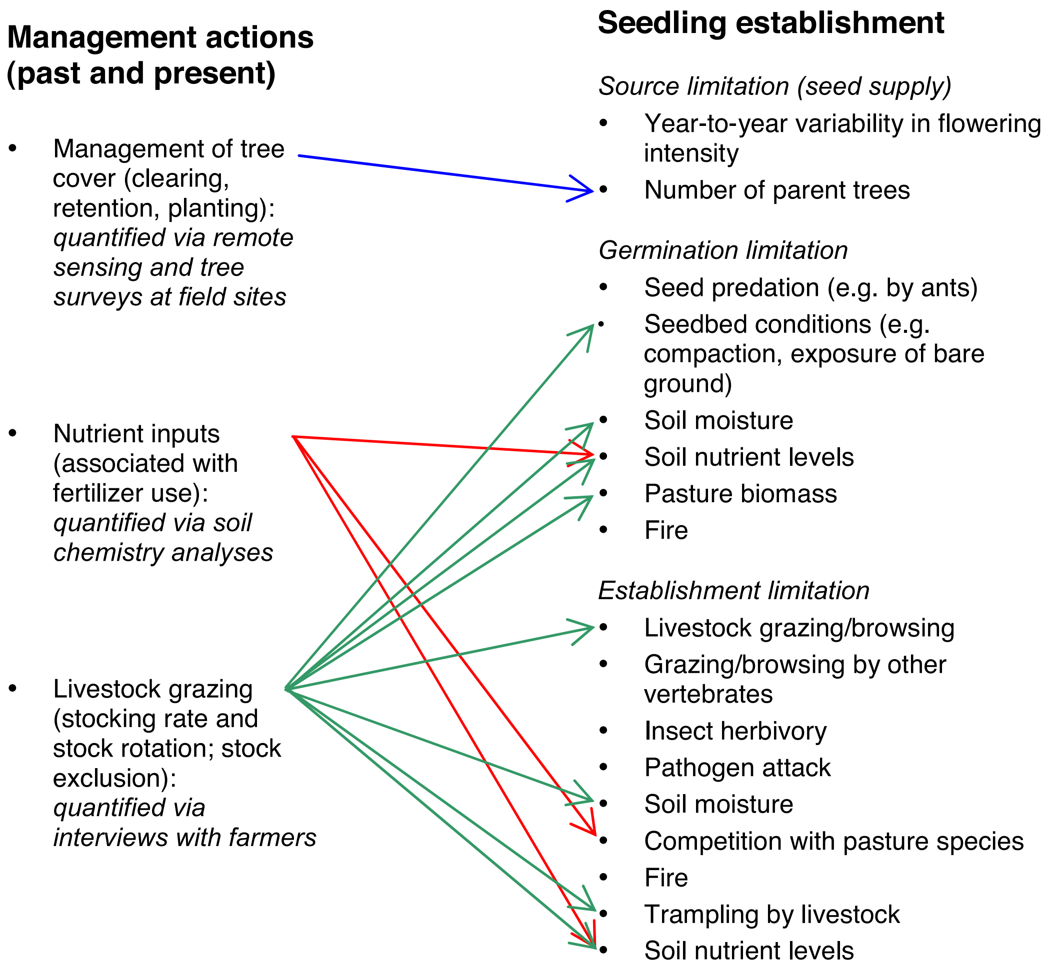

Fig. S2. Framework summarizing key variables affecting eucalypt regeneration [adapted from Acácio V, Holmgren M, Jansen PA, Schrotter O (2007) *Ecosystems* 10:1220–1230, and Vesk PA, Dorrough JW (2006) *Aust J Bot* 54:509–519] . The left column shows management actions offering potential leverage points for changing patterns of tree regeneration on the regional scale. Variables related to these actions were of particular interest in the design of the study, whereas variables less amenable to management action (e.g., year-to-year variability in flowering intensity) or operating at finer scales (e.g., seed predation by ants) were not of primary interest. The application of different grazing regimes can affect a range of variables related to germination and seedling establishment.

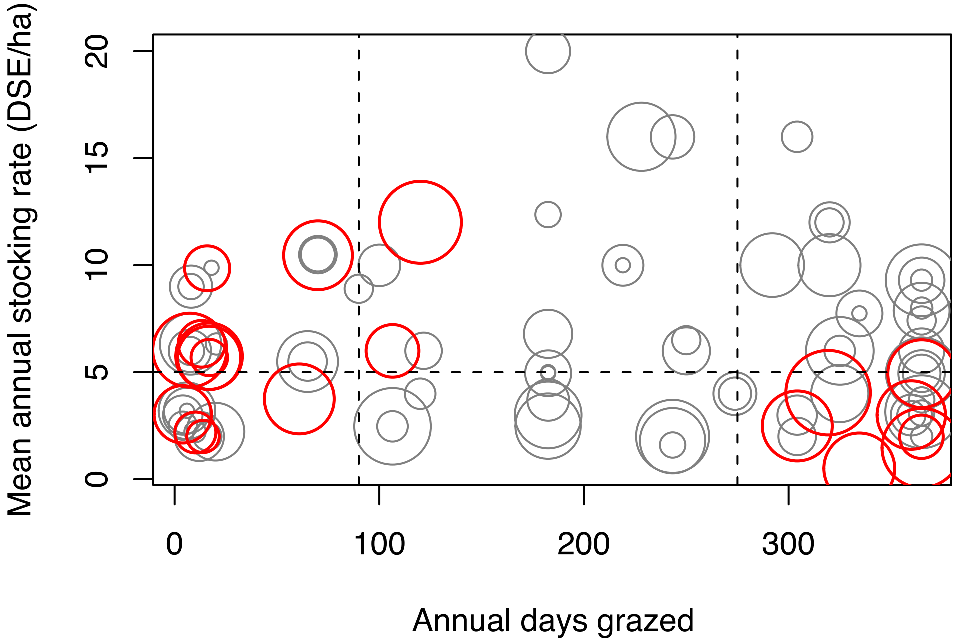

Fig. S3. Scatterplot showing the livestock grazing regime applied for the last 6 years at the grazed sites in the study. The size of the plotting symbol increases with the number of trees at the site. Stocking rates above 5 DSE per hectare were considered ''high'' (as opposed to ''low''). Rotation regimes were fast rotation for up to 90 days of grazing per year, continuous grazing for more than 275 days per year, and slow rotation for intermediate grazing regimes. Sites plotted in red had seedlings, whereas those plotted in gray had none. In addition to the sites shown here, there were 17 ungrazed woodland sites, 15 of which had seedlings. Note that the presence of seedlings was unknown when sites were first established.

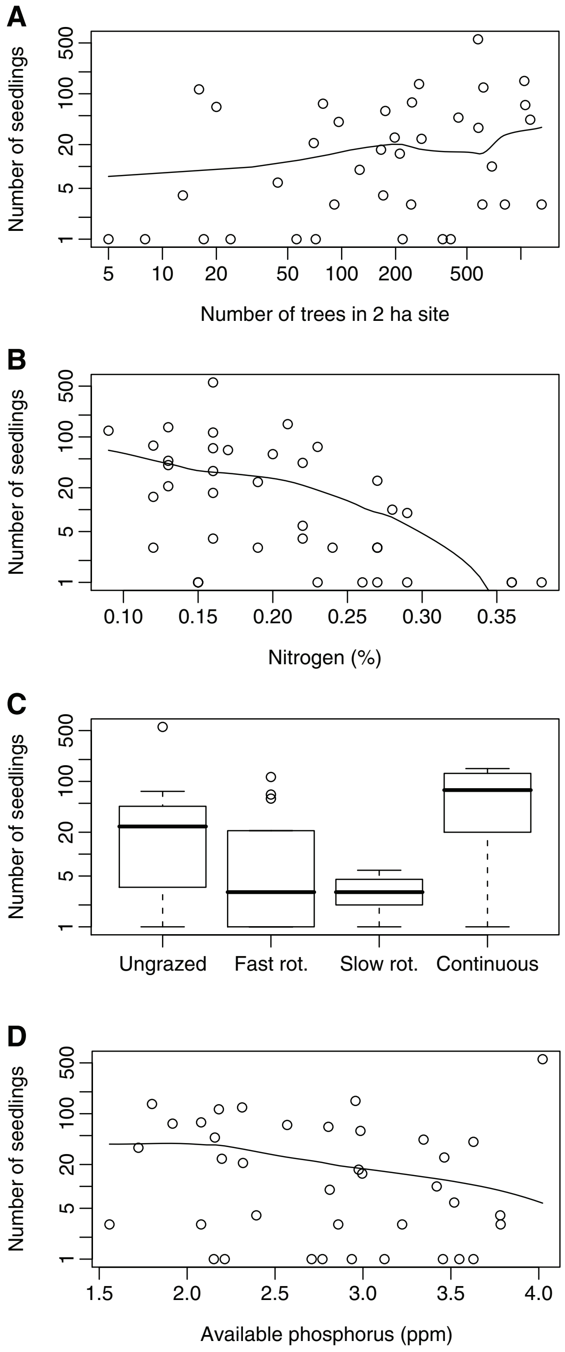

Fig. S4. Graphical display of the 38 sites with more than 1 seedling, in relation to those variables significantly explaining the presence of seedlings, or minimum tree diameter. (*A*) Relationship with tree density; (*B*) relationship with soil nitrogen; (*C*) relationship with grazing regime; (*D*) relationship with soil phosphorus. Smooth lines are fitted for illustration purposes only; formal significance tests were not conducted because data were limited.

Table S1. Results from linear regression of the number of trees at a site as a function of remotely sensed tree cover within 2 ha in percent

Variable	Estimate	Standard error	*P*
Intercept	0.570	0.226	
Tree cover	0.745	0.217	<0.001
Tree cover squared	0.102	0.0439	0.021

Both response and explanatory variables were log-transformed prior to analysis (after the addition of one to avoid zero values). The model reported here includes both primary survey sites and validation sites. The same variables were significant when only primary survey sites were included. The model explained 81% of variability in the response.

Table S2. Results from linear mixed model of the minimum diameter of trees at a site as a function of the number of trees at a site and available soil phosphorus

Variable	Estimate	Standard error	*P*
Intercept	−0.806	0.527	
Number of trees	−0.666	0.0493	<0.001
Available phosphorus	0.313	0.143	0.031

The minimum diameter of trees was modeled as a unit-free index (see *Methods*), with most values ranging between 0 and 1. Both explanatory variables were log-transformed prior to analysis because they were highly skewed. Farm was fitted as a random effect. The variance component associated with the random effect was 0.13, and the residual variance component was 0.80. The model reported here includes both primary survey sites and validation sites. The same variables were significant when only primary survey sites were included.

Table S3. Results from generalized linear mixed model of the presence of seedlings at a site as a function of total soil nitrogen, livestock rotation regime, and number of trees at the site

Variable	Estimate	Standard error	*P*
Intercept	0.4923	0.5420	
Number of trees	0.7393	0.1997	<0.001
Total soil nitrogen	−8.294	3.337	0.013
Rotation regime (baseline = fast rotation)		0.9959	0.004
Continuous grazing	−2.200		
Slow rotation	−2.485		
Ungrazed	0.258		

Logit link function and binomial error distribution were used. The number of trees at a site was log-transformed prior to analysis because it was highly skewed. Farm was fitted as a random effect. The variance component associated with the random effect was 0.82, and the residual variance component was fixed at 1. The mean standard error of differences between the treatments is shown for rotation regime. The model reported here includes both primary survey sites and validation sites. The same variables were significant when only primary survey sites were included.

Table S4. Conservatively estimated list of the birds, mammals, and reptiles in the study area that use trees or tree-derived features (such as logs or litter) as habitat

Scientific name	Common name	Key habitat feature used	Likely user of scattered trees
Birds			
Dromaius novaehollandiae	Emu	Trees for shelter	Yes
Chenonetta jubata	Australian Wood Duck	Tree hollows	Yes
Tadorna tadornoides	Australian Shelduck	Tree hollows	Yes
Anas superciliosa	Pacific Black Duck	Stumps and tree hollows	Yes
Anas gracilis	Grey Teal	Tree hollows	Yes
Accipiter cirrhocephalus	Collared Sparrowhawk	Trees for nesting	Yes
Accipiter fasciatus	Brown Goshawk	Trees for nesting	Yes
Hieraaetus morphnoides	Little Eagle	Trees for nesting	Yes
Aquila audax	Wedge-tailed Eagle	Trees for nesting	Yes
Falco berigora	Brown Falcon	Trees for nesting	Yes
Falco cenchroides	Nankeen Kestrel	Trees for nesting	Yes
Falco longipennis	Australian Hobby	Trees for nesting	Yes
Geopelia striata	Peaceful Dove	Trees for nesting	No
Phaps chalcoptera	Common Bronzewing	Trees and stumps for nesting	Yes
Ocyphaps lophotes	Crested Pigeon	Trees for nesting	Yes
Callocephalon fimbriatum	Gang-gang Cockatoo	Tree hollows	No
Cacatua roseicapilla	Galah	Tree hollows	Yes
Cacatua sanguinea	Little Corella	Tree hollows	Yes
Cacatua galerita	Sulphur-crested Cockatoo	Tree hollows	Yes
Glossopsitta pusilla	Little Lorikeet	Tree hollows	Yes
Polytelis swainsonii	Superb Parrot	Tree hollows	Yes
Nymphicus hollandicus	Cockatiel	Tree hollows	Yes
Platycercus elegans	Crimson Rosella	Tree hollows	Yes
Platycercus eximius	Eastern Rosella	Tree hollows	Yes
Psephotus haematonotus	Red-rumped Parrot	Tree hollows and stumps	Yes
Neophema pulchella	Turquoise Parrot	Stumps	Yes
Cacomantis pallidus	Pallid Cuckoo	Trees for nesting	Yes
Cacomantis flabelliformis	Fan-tailed Cuckoo	Trees for nesting	No
Chrysococcyx basalis	Horsfield's Bronze-cuckoo	Trees for nesting	Yes
Chrysococcyx lucidus	Shining Bronze-cuckoo	Trees for nesting	No
Ninox novaeseelandiae	Southern Boobook	Tree hollows	Yes
Dacelo novaeguineae	Laughing Kookaburra	Tree hollows	Yes
Todiramphus sanctus	Sacred Kingfisher	Tree hollows	Yes
Merops ornatus	Rainbow Bee-eater		Yes
Eurystomus orientalis	Dollarbird	Tree hollows	Yes
Cormobates leucophaeus	White-throated Treecreeper	Tree hollows	Yes
Climacteris picumnus	Brown Treecreeper	Tree hollows and stumps	Yes
Malurus cyaneus	Superb Fairy-wren	Dense foliage	No
Pardalotus punctatus	Spotted Pardalote	Canopy	No
Pardalotus striatus	Striated Pardalote	Tree hollows	Yes
Sericornis frontalis	White-browed Scrubwren	Dense foliage	No
Chthonicola sagittata	Speckled Warbler	Trees for nesting	No
Gerygone fusca	Western Gerygone	Trees for nesting	Yes
Gerygone olivacea	White-throated Gerygone	Trees for nesting	Yes
Acanthiza pusilla	Brown Thornbill	Trees for nesting	No
Acanthiza reguloides	Buff-rumped Thornbill	Trees for nesting	No
Acanthiza chrysorrhoa	Yellow-rumped Thornbill	Trees for nesting	Yes
Acanthiza lineata	Striated Thornbill	Trees for nesting	No
Acanthiza nana	Yellow Thornbill	Trees for nesting	No
Smicrornis brevirostris	Weebill	Trees for nesting	Yes
Aphelocephala leucopsis	Southern Whiteface	Tree hollows and stumps	Yes
Anthochaera carunculata	Red Wattlebird	Trees for nesting	Yes
Philemon citreogularis	Little Friarbird	Trees for nesting	Yes
Philemon corniculatus	Noisy Friarbird	Trees for nesting	Yes
Acanthagenys rufogularis	Spiny-cheeked Honeyeater	Trees for nesting	Yes
Plectorhyncha lanceolata	Striped Honeyeater	Trees for nesting	Yes
Entomyzon cyanotis	Blue-faced Honeyeater	Trees for nesting	Yes
Manorina melanocephala	Noisy Miner	Trees for nesting	Yes
Lichenostomus chrysops	Yellow-faced Honeyeater	Trees for nesting	Yes
Lichenostomus leucotis	White-eared Honeyeater	Trees for nesting	Yes
Lichenostomus fuscus	Fuscous Honeyeater	Trees for nesting	Yes
Lichenostomus penicillatus	White-plumed Honeyeater	Trees for nesting	Yes

Scientific name	Common name	Key habitat feature used	Likely user of scattered trees
Melithreptus brevirostris	Brown-headed Honeyeater	Trees for nesting	No
Acanthorhynchus tenuirostris	Eastern Spinebill	Trees for nesting	Yes
Microeca fascinans	Jacky Winter	Trees for nesting	Yes
Petroica multicolor	Scarlet Robin	Trees for nesting	No
Petroica goodenovii	Red-capped Robin	Trees for nesting	No
Eopsaltria australis	Eastern Yellow Robin	Trees for nesting	No
Melanodryas cucullata	Hooded Robin	Trees for nesting	No
Pomatostomus superciliosus	White-browed Babbler	Trees for nesting	No
Pomatostomus temporalis	Grey-crowned Babbler	Trees for nesting	Yes
Daphoenositta chrysoptera	Varied Sittella	Trees for nesting	No
Falcunculus frontatus	Crested Shrike-tit	Trees for nesting	No
Pachycephala pectoralis	Golden Whistler	Trees for nesting	No
Pachycephala rufiventris	Rufous Whistler	Trees for nesting	Yes
Colluricincla harmonica	Grey Shrike-thrush	Trees for nesting	Yes
Myiagra rubecula	Leaden Flycatcher	Trees for nesting	Yes
Myiagra inquieta	Restless Flycatcher	Tress for nesting	Yes
Rhipidura leucophrys	Willie Wagtail	Trees for nesting	Yes
Rhipidura fuliginosa	Grey Fantail	Trees for nesting	Yes
Coracina novaehollandiae	Black-faced Cuckoo-shrike	Trees for nesting	Yes
Lalage sueurii	White-winged Triller	Trees for nesting	Yes
Oriolus sagittatus	Olive-backed Oriole	Trees for nesting	Yes
Artamus superciliosus	White-browed Woodswallow	Trees and stumps	Yes
Artamus personatus	Masked Woodswallow	Trees and stumps	Yes
Artamus cinereus	Black-faced Woodswallow	Trees and stumps	Yes
Artamus cyanopterus	Dusky Woodswallow	Trees and stumps	Yes
Cracticus torquatus	Grey Butcherbird	Trees for nesting	Yes
Cracticus nigrogularis	Pied Butcherbird	Trees for nesting	Yes
Grallina cyanoleuca	Magpie-lark	Trees for nesting	Yes
Gymnorhina tibicen	Australian Magpie	Trees for nesting	Yes
Strepera graculina	Pied Currawong	Trees for nesting	Yes
Corvus coronoides	Australian Raven	Trees for nesting	Yes
Corvus mellori	Little Raven	Trees for nesting	Yes
Struthidea cinerea	Apostlebird	Trees for nesting	Yes
Corcorax melanorhamphos	White-winged Chough	Trees for nesting	Yes
Anthus novaeseelandiae	Richard’s Pipit		Yes
Neochmia temporalis	Red-browed Finch	Trees for nesting	Yes
Taeniopygia guttata	Zebra Finch	Trees for nesting	Yes
Stagonopleura guttata	Diamond Firetail	Trees for nesting	No
Dicaeum hirundinaceum	Mistletoebird	Trees for nesting	Yes
Hirundo neoxena	Welcome Swallow	Trees for nesting	Yes
Cheramoeca leucosternus	White-backed Swallow		Yes
Hirundo nigricans	Tree Martin	Trees for nesting	Yes
Hirundo ariel	Fairy Martin	Trees for nesting	Yes
Cincloramphus cruralis	Brown Songlark	Trees for perching	Yes
Cincloramphus mathewsi	Rufous Songlark	Trees for perching	Yes
Zosterops lateralis	Silvereye	Trees for feeding	Yes
Mammals			
Tachyglossus aculeatus	Short-beaked Echidna	Logs and rotting wood	Yes
Antechinus flavipes	Yellow-footed Antechinus	Tree hollows and logs	No
Antechinus stuartii	Brown Antechinus	Tree hollows	No
Sminthopsis murina	Common Dunnart	Logs	No
Trichosurus vulpecula	Common Brushtail Possum	Tree hollows and logs	Yes
Petaurus breviceps	Sugar Glider	Tree hollows	Yes
Petaurus norfolcensis	Squirrel Glider	Tree hollows	Yes
Pseudocheirus peregrinus	Common Ringtail Possum	Tree hollows	Yes
Macropus giganteus	Eastern Grey Kangaroo	Trees for shade/shelter	Yes
Macropus robustus	Common Wallaroo	Trees for shade/shelter	Yes
Macropus rufogriseus	Red-necked Wallaby	Trees for shade/shelter	No
Wallabia bicolor	Swamp Wallaby	Trees for shade/shelter	Yes
Pteropus scapulatus	Little Red Flying-Fox	Roosts in trees	Yes
Saccolaimus flaviventris	Yellow-bellied Sheathtail bat	Tree hollows	Yes
Tadarida australis	White-stiped Freetail Bat	Tree hollows	Yes
Mormopterus planiceps	Little Mastiff-bat	Tree hollows	Yes

Scientific name	Common name	Key habitat feature used	Likely user of scattered trees
Mormopterus spp.	Other species	Tree hollows	Yes
Chalinolobus gouldii	Gould's Wattled Bat	Tree hollows	Yes
Chalinolobus morio	Chocolate Wattled Bat	Tree hollows	Yes
Vespadalus darlingtoni	Large Forest Bat	Tree hollows	Yes
Vespadalus regulus	Southern Forest Bat	Tree hollows	Yes
Vespadalus vulturnus	Little Forest Bat	Tree hollows	Yes
Scotorepens balstoni	Inland Broad-nosed Bat	Tree hollows	Yes
Nyctophilus geoffroyi	Lesser Long-eared Bat	Tree hollows	Yes
Nyctophilus gouldi	Gould's Long-eared Bat	Tree hollows and under bark	Yes
Rattus fuscipes	Bush Rat	Logs	No
Reptiles			
Christinus marmoratus	Marbled Gecko	Loose bark, logs	Yes
Diplodactylus vittatus	Eastern Stone Gecko	Logs, litter	Yes
Delma inornata	Olive Legless Lizard	Logs	Yes
Lialis burtonis	Burton's Snake-lizard	Logs, litter	Yes
Acritoscincus platynotum	Red-throated Skink	Logs, litter	Yes
Carlia tetradactyla	Southern Rainbow Skink	Logs, litter	Yes
Ctenotus robustus	Eastern Striped Skink	Logs and litter	Yes
Ctenotus taeniolatus	Copper-tailed Skink	Logs	Yes
Egernia cunninghami	Cunningham's Skink	Logs	Yes
Egernia saxatilis	Black Rock Skink	Logs	No
Egernia striolata	Tree Skink	Tree hollows, under bark, logs	Yes
Egernia whitii	White's Skink	Logs	No
Hemiergis decresiensis	Three-toed Skink	Logs	No
Lampropholis delicata	Garden Skink	Litter	Yes
Lampropholis guichenoti	Grass Skink	Litter	Yes
Morethia boulengeri	Boulenger's Skink	Logs and under bark	Yes
Tiliqua rugosa	Shingleback	Logs and Litter	Yes
Tiliqua scincoides	Common Blue-tongue	Logs	Yes
Amphibolurus muricatus	Jacky Dragon	Logs and fallen timber	Yes
Pogona barbata	Common Bearded Dragon	Logs, stumps	Yes
Varanus varius	Lace Monitor	Often aboreal	Yes
Ramphotyphlops spp.	Blind snakes	Logs	Yes
Morelia spilota	Carpet Python	Often aboreal	Yes
Parasuta dwyeri	Dwyer's Snake	Logs	Yes
Pseudechis porphyriacus	Red-bellied Black Snake	Logs	Yes
Pseudonaja textilis	Eastern Brown Snake	Logs	Yes
Vermicella annulata	Bandy-bandy	Logs	Yes

Likely use of scattered trees is also indicated. Data are based on range maps, publicly accessible databases, field guides, and local knowledge.