

 $\overline{a}$ 

<sup>∗</sup> Corresponding author phone: +541-737-9194: fax: +541-737-0497: e-mail: Staci.Simonich@orst.edu

1 Lake west of the Continental Divide. Hoh Lake (Olympic) is a hanging cirque lake (1384 masl) 2 orientated on the north side of Mt. Olympus. PJ Lake (Olympic) is a hanging cirque lake (1433 masl) 3 orientated on the northeast side of Mt. Olympus, approximately 27 km from PJ Lake. Golden Lake 4 (Rainier) is high-altitude perched lake orientated on the west side of Mt. Rainier. LP19 (Rainier) is an 5 unnamed high-altitude perched lake also orientated on the west side of Mt Rainier, approximately 7.5 6 km south of Golden Lake. Snyder Lake (Glacier) is a perched cirque lake (1600 masl) located east of 7 the Continental Divide. Oldman Lake (Glacier), a cirque lake (2026 masl), is located approximately 30 8 km from Snyder Lake west of the Continental Divide. Wonder Lake (Denali) and McLeod Lake 9 (Denali) are a piedmont lakes located approximately 55 km north of Mt. McKinley. Lake Matcharak 10 (Gates) is located in the arctic tundra along the Noatak River south of the central Brooks Range. Burial 11 Lake (Noatak) is located in the arctic and resides in the Foothills of the North Brooks Range 12 approximately 144 km west of Lake Matcharak.

13

14 Snowpack Collection and Analysis. Snow sampling was conducted as described previously (1,2). 15 Briefly, season snowpack (50 kg) was collected at the end of the snow accumulation season (March or 16 April) into polytetrafluoroethylene (PTFE) bags from a snow pit. Samples were shipped overnight on 17 dry ice back to the laboratory. Snow samples were shielded from light and were allowed to melt in 18 sealed bags at 22°C overnight prior to extraction using a modified hydrophobic/hydrophilic Speedisk 19 (*1*). Prior to extraction, 1 mL of methanol spiked with isotopically labeled deuterated polycyclic 20 aromatic hydrocarbons (d-PAH) surrogates-ethyl acetate (EA) solution was apportioned amongst the 21 PTFE bags to correct for PAH loss over the entire analytical method. PAHs were isolated and matrix 22 interferences were removed using gel permeation chromatography, followed by silica gel 23 chromatography (*1*). Extracts were concentrated ( $\sim$ 200  $\mu$ L) and spiked with an isotopically labeled d-24 PAH (internal standard-EA solution. Gas chromatographic mass spectrometry with electron impact 25 ionization (GC/EI-MS) with selective ion monitoring was used for the separation, detection, and 26 quantification of PAHs (*1*).

Lichen Collection and Analysis. In brief, one lichen species was collected within a  $\sim 0.01 \text{ km}^2$ 2 3 sampling area at each site. Lichens were collected from Snyder Lake catchment (*Platismatia glauca*) 4 and Oldman Lake Catchment (*Letharia vulpine*). The lichen samples were collected upon availability at 5 each site by hand and stored in 2-L metalized polyester bags (Kapak Corp, Minneapolis, MN). Once 6 full, the bags were sealed with laboratory tape and double bagged in zipper-locking plastic bags. The 7 samples were stored and shipped overnight to the laboratory in coolers with ice and then stored at -40°C 8 until analysis. The samples were ground using a Büchi Mixer B-400 (Büchi-Mixer B-400, Flawil, 9 Switzerland), packed in 100 mL Accelerated Solvent Extractor (ASE) cells (Dionex, Sunnyvale, CA), 10 spiked with d-PAH surrogates solution, and extracted using pressurized liquid extraction. Lichen 11 samples were extracted twice using an ASE 300 and dichloromethane (100 °C, 1500 psi, 1 cycles of 5 12 minutes, 75% flush volume). Extract cleanup steps, isotopically-labeled surrogate and internal standard 13 spikes, separation, detection, and quantification (GC/EI-MS) were the same as described above.

14

15 Sediment Collection, Dating, and Analysis. Sediment sampling was conducted as described 16 previously (*3*). Briefly, at least two sediment cores were collected from the deepest area of the lake 17 using an Uwitec gravimetric corer and 78 mm polycarbonate core tube. Sediment cores were extruded 18 vertically and sectioned in the field, 0.5 cm intervals for first 10 cm and 1 cm intervals thereafter. 19 Intervals were stored separately in pre-cleaned glass jars. Sediment samples were kept on blue ice and 20 shipped overnight to the laboratory. Sediment cores were identified for SOC analysis based on visual 21 inspection of the sediment core, followed by  $^{210}$ Pb dating. Intervals from within a sediment core were 22 identified based on  $^{210}$ Pb dating and contaminant use profiles. An aliquot from each sediment core 23 interval was analyzed for percent moisture, total organic carbon (TOC), and  $^{210}Pb$ ,  $^{137}Cs$ ,  $^{226}Ra$ , and 24<sup>241</sup>Am activity. <sup>210</sup>Pb, <sup>137</sup>Cs, <sup>226</sup>Ra, and <sup>241</sup>Am were analyzed by direct gamma assay at the Liverpool 25 University Environmental Radioactivity Laboratory, using an Ortec HPGe GWL series well-type coaxial 26 low background intrinsic germanium detector (4). Radiodating of 210Pb and 226Ra was determined by

1 the gamma emissions at 46.5 keV and 295 keV, respectively. Complete details of these analyses are 2 provided elsewhere  $(3,4)$ . The constant rate of supply model, along with measured <sup>210</sup>Pb activity, was 3 used to determine the sedimentation rate and average date of each interval (*4-6*). The total and 4 supported <sup>210</sup>Pb activity for each core, along with the <sup>137</sup>Cs concentrations versus depth are provided in 5 Figure S2 and S15.

6 The extraction, isolation, and quantification of PAHs in sediment is described elsewhere (*3*). In 7 brief, wet sediment (10-40 g ww) was dried with sodium sulfate and extracted using pressurized liquid 8 extraction. Prior to extraction, samples were spiked with an isotopically labeled recovery surrogates to 9 correct for PAH loss over the analytical method. Extracts were purified using the gel permeation 10 chromatography and silica gel chromatography (*3*). Extract cleanup steps, isotopically-labeled surrogate 11 and internal standard spikes, separation, detection, quantification (GC/EI-MS), recoveries, and WACAP 12 quality assurance objectives have been previously described (3,7). The sediment extracts were analyzed 13 for target PAHs by GC/MS, using electron impact (EI) ionization with selective ion monitoring as 14 described in by Usenko et al (*3*).

15 Sodium sulfate was used as the laboratory blank and was carried through the entire analytical 16 method (extraction, cleanup, and concentrating), starting at the grinding step, the laboratory blank was 17 spiked with the same quantity of isotopically labeled surrogate and internal standards as samples (*3*). 18 PAH concentrations in sediment were surrogate recovery (concentration calculated relative to surrogate) 19 and laboratory blank corrected.

20

21 Sediment Focusing Correction. Focusing factors (FF) were calculated for each sediment core in order 22 to correct for sediment focusing. Focusing factors are used to quantify sedimentation enhancement or 23 reduction at the coring site. The <sup>210</sup>Pb inventory was derived by plotting the unsupported <sup>210</sup>Pb against 24 the mass sedimentation accumulation rate  $(8)$  and the <sup>210</sup>Pb atmospheric fallout was modeled from ice 25 cores, soil samples, and atmospheric collectors near sampling sites  $(9-13)$ . <sup>210</sup>Pb atmospheric fallout 26 may vary over short time periods and in mountainous terrain, where precipitation varies greatly with

elevation and orientation. However, over the time frame of these sediment cores ( $\leq 150$  years), the <sup>210</sup>Pb 2 atmospheric fallout is considered to be fairly constant in a regional context (4,14). The lake sediment 3 core FFs ranged from 0.78 to 4.55 (Table S1). In order to compare the spatial and temporal trends of 4 PAH deposition among different sediment cores, all sediment PAH concentrations were multiplied by 5 the lake sedimentation rate and normalized to the FF. This converted all PAH concentrations in 6 sediment ( $\mu$ g g<sup>-1</sup> dry wt) to focus-corrected PAH fluxes ( $\mu$ g m<sup>-2</sup> y<sup>-1</sup>).

- 7
- 8





6 7





 $\frac{9}{10}$ 

Figure S2: Fallout radionuclides in Pear Lake core showing (a) total and supported <sup>210</sup>Pb, (b) unsupported <sup>210</sup>Pb, (c)  $137$ Cs concentrations versus depth. (c) Cs concentrations versus depth.





- 
- 
- 
- 
- 



13 Figure S5: Fallout radionuclides in Mills Lake core showing (a) total and supported <sup>210</sup>Pb, (b) unsupported  $^{210}Pb$ , (c)  $^{137}Cs$  concentrations and  $^{137}Cs^{210}Pb$  activity ratios versus depth.

- 
- 
- 
- 



13 Figure S7: Fallout radionuclides in Wonder Lake core showing (a) total and supported <sup>210</sup>Pb, (b) unsupported  $^{210}Pb$ , (c)  $^{137}Cs$  and  $^{241}Am$  concentrations versus depth.

- 
- 
- 
- 
- 





 

 $\begin{array}{c} 5 \\ 6 \\ 7 \end{array}$ 

 

13 Figure S9: Fallout radionuclides in Matcharak Lake core showing (a) total and supported <sup>210</sup>Pb, (b) unsupported  $^{210}Pb$ , (c)  $^{137}Cs$  concentrations versus depth.

- 
- 
- 
- 
- 



**Figure S10:** Fallout radionuclides in PJ Lake core showing (a) total and supported <sup>210</sup>Pb, (b) unsupported <sup>210</sup>Pb, (c)  $^{137}Cs$  concentrations versus depth.





12 Figure S11: Fallout radionuclides in Hoh Lake core showing (a) total and supported <sup>210</sup>Pb, (b) unsupported  $^{210}Pb$ , (c)  $^{137}Cs$  concentrations versus depth.

- 
- 
- 
- 
- 





- 
- 
- 
- 
- 
- 







16 Figure S15: Fallout radionuclides in Snyder Lake core showing (a) total and supported <sup>210</sup>Pb, (b) unsupported 17 <sup>210</sup>Pb, (c) <sup>137</sup>Cs and <sup>241</sup>Am concentrations versus depth.

- 
- 
- 





Lichen Concentration Snowpack Concentration  $\circ$  $\ddot{}$ 

national parks.

8

9

10





Figure S17: PAH profiles measured in the 2003 and 2004 seasonal snowpack for all WACAP lake catchments. Bars represent the average percent of the total PAH concentration. Error bars represent the percent relative standard deviation. Total PAH concentration for each lake catchment is provided in parentheses.



- 12
- 13

**3** Figure S18. Natural log focus-corrected flux (ng  $m^{-2} y^{-1}$ ) profiles of PAHs in WACAP sediment cores. 4 Doubling times (t<sub>2</sub>) and half-lives (t<sub>1/2</sub>) are given where least squares regressions were statistically significant ( $p$ <0.05). statistically significant  $(p<0.05)$ .



**Figure S19. A)** Map of Glacier National Park, MT. Brown line indicates park boundary and yellow<br>8 lines indicate lake catchment. **B)** Lake bathymetry maps for Snyder Lake (west) and Oldman Lake 8 lines indicate lake catchment. **B**) Lake bathymetry maps for Snyder Lake (west) and Oldman Lake (east). Stars indicate coring site location with in lake. (east). Stars indicate coring site location with in lake.



Figure S20: Isolines of fluoride (ppb) measured in foliage samples taken along transects from aluminum smelter in 1970 from reference  $(15)$ .

- 2 5 6 7 8 9
- 10



4 Figure S21: A) IcdP/(IcdP+BeP) and B) IcdP/(IcdP+BghiP) fraction ration (average  $\pm$  standard deviation) calculated from seasonal snowpack, lichen, and pre 1955 and surficial sedime 5 deviation) calculated from seasonal snowpack, lichen, and pre 1955 and surficial sediment cores intervals for lake catchments in Glacier National Park.  $a$  reference (16),  $b$  reference (17),  $c$  $7 \treference (18)$ , and  $d$  reference (19).  $*$  indicates concentration was below detection limits.



2 Figure S22: Ratios of Ind/(Ind+BeP), IcdP/(IcdPd+BghiP), and FLA/(FLA+PYR) calculated from all 2003 and 2004 season snowpack samples. Dashed lines  $(--)$  indicate the PAH ratio of a 2003 and 2004 season snowpack samples. Dashed lines  $(--)$  indicate the PAH ratio of a specific PAH source. Solid lines  $(--)$  indicate the PAH ratio from an aluminum smelter is specific PAH source. Solid lines  $\left(\frac{1}{6}\right)$  indicate the PAH ratio from an aluminum smelter using<br>Söderberg technology. <sup>a</sup> reference (16), <sup>b</sup> reference (19), <sup>c</sup> reference (17), <sup>d</sup> reference (18), and <sup>e</sup> 5 Söderberg technology.  $a^a$  reference (16),  $b^b$  reference (19),  $c^c$  reference (17),  $d^d$  reference (18), and  $c^e$ 6 reference (20).

S20



Indeno(1,2,3-cd)pyrene Ln(Flux) ( $\mu$ g m<sup>-2</sup> y<sup>-1</sup>)

1 2 **Figure S23:** A) Focus-corrected flux ( $\mu$ g m<sup>-2</sup> y<sup>-1</sup>) profiles of retene in Oldman Lake (east) and Snyder 4 Lake (west) sediment core. **B**) Natural log focus-corrected flux ( $\mu$ g m<sup>-2</sup> y<sup>-1</sup>) profiles of retene in Oldman Lake and Snyder Lake sediment core. C) Natural log focus-corrected flux ( $\mu$ g m<sup>-2</sup> y<sup>-</sup>  $\frac{5}{6}$ 6 <sup>1</sup>) profiles of retene versus ΣPAHs in Oldman Lake and Snyder Lake sediment core. **D**) 7 Natural log focus-corrected flux ( $\mu$ g m<sup>-2</sup> y<sup>-1</sup>) profiles of retene versus indeno(1,2,3-cd)pyrene in 8 Oldman Lake and Snyder Lake sediment core. 9



## 3  $\frac{3}{4}$  $\frac{5}{6}$

2

6 Figure S24: Individual PAH flux profiles in all WACAP lake sediment cores and intervals were analyzed by PCA using S-Plus 7.0 (Insightful, Seattle, WA). Results from the PCA show 7 analyzed by PCA using S-Plus 7.0 (Insightful, Seattle, WA). Results from the PCA showed that the total variance accounted for by PC1 and PC2. The circle on the right indicates a cluster that 8 the total variance accounted for by PC1 and PC2. The circle on the right indicates a cluster that<br>9 contains all Snyder Lake sediment intervals that dated after 1955. The oval on the left indicates contains all Snyder Lake sediment intervals that dated after 1955. The oval on the left indicates 10 a second cluster that contains all sediment intervals from the other WACAP catchments 11 including a sediment interval from Snyder Lake dated at 1893. The arrows indicate the direction 12 and magnitude of the individual PAH loadings.

- 13
- 14 15
- 16
- 17
- 18
- 19
- 20 21
- 22
- 23
- 24
- 25
- 26

1 Table S1: Physical and chemical limnological characteristics of each lake sites. \* indicates the data was obtained from the parameter-elevation regressions on independent slopes model (PRISM) 2 obtained from the parameter-elevation regressions on independent slopes model (PRISM)<br>3 (average annual total precipitation from 1971-2000, 800×800 m). Sampling site's physical 3 (average annual total precipitation from 1971-2000, 800×800 m). Sampling site's physical characteristics were determined from USGS 15' topographic quadrangles, mapping datum 4 characteristics were determined from USGS 15' topographic quadrangles, mapping datum WGS84, and on-site measurements. WGS84, and on-site measurements.



- 18 19
- 20
- 21
- 22
- 
- 23

Table S1 (Continued): Physical and Chemical Limnological Characteristics of Lake sites. \* indicates 4 the data was obtained from the parameter-elevation regressions on independent slopes model<br>5 (average annual total precipitation from 1971-2000, 800×800 m). Sampling site's physical (average annual total precipitation from 1971-2000,  $800 \times 800$  m). Sampling site's physical 6 characteristics were determined from USGS 15' topographic quadrangles, mapping datum 7 WGS84, and on-site measurements.



Glacier National Park **Denali National Park** Denali National Park Denali National Park



Gates of the Arctic National Park and Noatak National Preserve





11

1

 $\frac{2}{3}$ 

8 9

- 12
- 13

## 14 Reference

- 15 (1) Usenko, S.; Hageman, K. J.; Schmedding, D. W.; Wilson, G. R.; Simonich, S. L. Trace Analysis 16 of Semi-Volatile Organic Compounds in Large Volume Samples of Snow, Lake Water, and
- 17 Groundwater. *Environ. Sci. & Technol.* 2005, *39*, 6006-6015.
- 18 (2) Hageman, K. J.; Simonich, S. L.; Campbell, D. H.; Wilson, G. R.; Landers, D. H. Atmospheric
- 19 Deposition of Current-Use and Historic-Use Pesticides in Snow at National Parks in the Western United 20 States. *Environ. Sci. & Technol.* 2006, *40*, 3174-3180.
- 21 (3) Usenko, S.; Landers, D. H.; Appleby, P. G.; Simonich, S. L. Current and Historical Deposition of
- 22 PBDEs, Pesticides, PCBs, and PAHs to Rocky Mountain National Park, USA. *Environ. Sci. & Technol.*
- 23 2007, *41*, 7235-7241.
- 1 (4) Appleby, P. G.; Nolan, P. J.; Gifford, D. W.; Godfrey, M. J.; Oldfield, F.; Anderson, N. J.;
- 2 Battarbee, R. W. <sup>210</sup>Pb Dating by low background gamma counting. *Hydrobiologia* 1986, 141, 21-27.
- 3 (5) Appleby, P. G.; Piliposian, G. T. Radiometric dating of sediment records from mountain lakes in
- 4 the Tatra Mountains. *Biologia* 2006, *61*, S51-S64. 5 (6) Appleby, P. G. Chronostratigraphic Techniques in Recent Sediments. In *Tracking Environmental*  6 *Change Using Lake Sediments*; Last, W. M., Smol, J. P., Eds.; Kluwer Academic Publishers, 2001; Vol.
- 
- $\begin{array}{cc} 7 & 1. \\ 8 & (7) \end{array}$ 8 (7) Landers, D. H.; Simonich, S. L.; Jaffe, D. A.; Geiser, L.; Campbell, D. H.; Schwindt, A.;
- 9 Schreck, C.; Kent, M.; Hafner, W. D.; Taylor, G. E.; Hageman, K. J.; Blett, T.; Erway, M. M.; Pierson,
- 10 S.; Christie, S. An Assessment of Airborne Contaminants in National Parks of the Western United
- 11 States. EPA/600/R-03/035. [http://www2.nature.nps.gov/air/Studies/air\\_toxics/wacap.cfm.](http://www2.nature.nps.gov/air/Studies/air_toxics/wacap.cfm) 2007.
- 12 (8) Zhu, L. Y.; Hites, R. A. Brominated Flame Retardants in Sediment Cores from Lake Michigan 13 and Erie. *Environ. Sci. & Technol.* 2005, *39*, 3488-3494.
- (9) Granstein, W. C.; Turekian, K. K.  $^{210}Pb$  and  $^{137}Cs$  in Air and Soils Measure the Rate and Vertical 15 Profile of Aerosol Scavenging. *J. Geophys. Res.* 1986, *91*, 14,355-314,366.
- 16 (10) Carpenter, R.; Peterson, M. L.; Bennett, J. T.; Somayajulu, B. L. K. Mixing and cycling of
- 17 uranium, thorium and 210 Pb in Puget Sound sediments. *Geochim. Cosmochim. Acta* 1984, *48*, 1949- 18 1963.
- 19 (11) Monaghan, M. C. Lead 210 in Surface Air and Soils From California: Implications for the
- 20 Behavior of Trace Constituents in the Planetary Boundary Laker. *J. Geophys. Res.* 1989, *94*, 6449-6456.
- 21 (12) Monaghan, M. C.; Holdsworth, G. The origin of non-sea-salt sulphate in the Mount Logan ice
- 22 core. *Nature* 1990, *343*, 245 248.
- 23 (13) Nevissi, A. E. Measurement of 210Pb Atmospheric Flux in the Pacific Northwest. *Health Phys.* 24 1985, *48*, 169-174.
- 25 (14) Appleby, P. G. Radiometric dating of sediment cores. WACAPs Project Year 3 Report. 2005.
- 26 (15) Peterson, D. L.; Sullivan, T.; Eilers, J. M.; Brace, S.; Horner, D.; Savig, K.; Morse, D.
- 27 Assessment of Air Quality and Air Pollutant Impacts in National Parks of the Rocky Mountains and
- 28 Northern Great Plains. *U.S. Department of the Interior National Park Service.* 1998,
- 29 *<http://www2.nature.nps.gov/air/Pubs/>*.
- 30 (16) Sanderson, E. G.; Farant, J. P. Atmospheric size distribution of PAHs: Evidence of a high-
- 31 volume sampling artifact. *Environ. Sci. & Technol.* 2005, *39*, 7631-7637.
- 32 (17) Schauer, J. J.; Kleeman, M. J.; Cass, G. R.; Simoneit, B. R. T. Measurement of emissions from
- 33 air pollution sources. 3. C-1-C- 29 organic compounds from fireplace combustion of wood. *Environ. Sci.*  34 *& Technol.* 2001, *35*, 1716-1728.
- 35 (18) Schauer, J. J.; Kleeman, M. J.; Cass, G. R.; Simoneit, B. R. T. Measurement of emissions from
- 36 air pollution sources. 5. C-1-C-32 organic compounds from gasoline-powered motor vehicles. *Environ.*  37 *Sci. & Technol.* 2002, *36*, 1169-1180.
- 38 (19) Yunker, M. B.; Macdonald, R. W.; Vingarzan, R.; Mitchell, R. H.; Goyette, D.; Sylvestre, S.
- 39 PAHs in the Fraser River basin: a critical appraisal of PAH ratios as indicators of PAH source and
- 40 composition. *Org. Geochem.* 2002, *33*, 489-515.
- 41 (20) Booth, P.; Gribben, K. A review of the formation, environmental fate, and forensic methods for
- 42 PAHs from aluminum smelting processes. *Environ. Forensics* 2005, *6*, 133-142.
- 43
- 44