Supplemental Information

SPARROW Model Background

SPARROW models are calibrated by estimating model coefficients that minimize the difference between the predicted and observed mean annual stream loads at a large number of fixed monitoring stations that are widely distributed throughout and representative of a hydrologic region. To maximize the explanatory power of the models, the mean annual stream loads must be compiled from as many sites as possible and often including sites where data are not collected contemporaneously. To account for differences in record length and sample size between stations, the water-quality records for the stations are detrended to reflect conditions during a single base year (Preston *et al*., 2009). The base year (2002 for our analysis) is selected to provide the greatest overlap between water-quality and streamflow records and is usually consistent with the temporal availability of the explanatory data used in the model. The detrended, mean annual stream loads are also normalized to long-term, mean hydrologic conditions to account for short-term spatial variability in precipitation across a region. This means that the measured stream loads used to calibrate a SPARROW model can be interpreted as the mean annual stream load that would have occurred in a specified base year if the mean annual streamflow for that year was equal to the long-term mean annual streamflow for the period of record.

Stream Reach Attributes

The SPARROW model includes a variable that allows the user to simulate the diversion of streamflow in a hydrologic network (the "frac value", which is the fraction of streamflow that is delivered from one reach to the reach immediately downstream). We employed three methods for estimating frac values. The first, and preferred method, was to compare the measured, longterm, mean annual streamflow between gages located upstream and downstream from known irrigation diversions and to calculate the ratio between the downstream and upstream values. When the upstream and downstream gages were separated by more than one reach, the frac value was scaled equally over the intervening reaches. When only one gage was available with longterm, mean annual streamflow and it was located above known irrigation diversions for which estimates could be made of the long-term mean annual discharge, the frac value was calculated by subtracting the total discharge for the diversions from the value for the upstream gage and dividing this value by the value for the upstream gage. The last, and least preferred method, was to compare the estimated long-term, mean streamflow for sequential reaches on the RF1 network (Brakebill *et al*., 2011). A frac value less than 1.0 was assigned to any reach in which there was a decrease in estimated streamflow between that reach and the reach immediately upstream. The approach we used to account for irrigation diversions in the PNW caused nutrient stream load to be removed from the modeled stream network, but did not return any of that stream load back to the network because we did not have the data needed to estimate this. Most of the water used for irrigation is taken up by crops or is lost through evaporation, and some leaches into deep groundwater. The remaining water, along with nutrients originally in the water and nutrients collected as it passes over agricultural land, is returned directly to streams through irrigation return drains or through shallow groundwater. Although our estimates of the frac values likely did not provided a full accounting of the hydrologic budget for the PNW, the values were based on the best available information and led to a more accurate model than otherwise would have been possible without including them.

Table S1 and Figure S1 provide details on the 303(d) listing in the PNW related to nutrient enrichment.

Table S1. Compilation of Stream Reaches Placed on State 303(d) Lists in the United States Pacific Northwest Because of Impairment Related to Nutrient Enrichment.

Figure S1. Prevalence of 303(d) Listings Because of Impairment Related to Nutrient Enrichment in the United States Pacific Northwest.

Watershed Attributes

The watershed attributes used in the PNW SPARROW models are shown in Figures S2 – S14.

Figure S2. Dominant Land Use in the United States Pacific Northwest (2001).

Figure S3. Point Source Discharges of Total Nitrogen in the United States Pacific Northwest (2002 estimates).

Figure S4. Point Source Discharges of Total Phosphorus in the United States Pacific Northwest (2002 estimates).

Figure S5. Mean Annual Wet Deposition of Inorganic Nitrogen in the United States Pacific Northwest.

Figure S6. Nitrogen Input from Farm Fertilizer in the United States Pacific Northwest (2002 estimates).

Figure S7. Phosphorus Input from Farm Fertilizer in the United States Pacific Northwest (2002 estimates).

Figure S8. Nitrogen Input from Manure from Livestock Production in the United States Pacific Northwest (2002 estimates).

Figure S9. Phosphorus Input from Manure from Livestock Production in the United States Pacific Northwest (2002 estimates).

Figure S10. Mean Soil Permeability in the United States Pacific Northwest.

Figure S11 Percentage of Land Area Consisting of Hydrologic Landscape Region 20 in the United States Pacific Northwest.

Figure S12. Effective Mean Annual Precipitation in the United States Pacific Northwest.

Figure S13. Basal Density of Red Alder Trees in the United Stated Pacific Northwest (2001 estimates).

Figure S14. Estimated Concentration of Phosphorus in Surficial Geologic Material in the United States Pacific Northwest.

The dominant land use was determined using data from the National Land Cover Database (NLCD (Homer *et al*., 2004). The area of forest land in each watershed was estimated by summing the area of land classified as deciduous, evergreen, and mixed forests in the NLCD. The area of developed land in each watershed was estimated by summing the area of land classified as low, medium, and highly developed in the NLCD. Developed land classified as open space in the NLCD was not included in our estimate because this category included roads in nondeveloped watersheds. In addition, the area of developed land in watersheds where the population density was less than 10 people per $km²$ was set to zero.

The total red alder basal area within each SPARROW watershed was estimated for 2001. The red alder species models obtained from the Landscape, Ecology, Modeling, Mapping and Analysis (LEMMA) project (LEMMA, 2008 for Oregon; Mathew Gregory, Oregon State University, written commun., September 8, 2008 for Washington)were raster files with 30 meter cells. Estimates of red alder basal area density were extracted from each of the 5 raster files and

merged into a single raster file, from which the total red alder basal area in each SPARROW watershed was estimated.

The phosphorus sampling performed by the National Geochemical Survey (USGS, 2004) was limited to select streams in the PNW. Therefore, a value for each SPARROW watershed needed to be estimated by using landscape data that were spatially continuous across the region. This was done by combining data describing surface geology (Schruben *et al*., 1997) and USEPA level 4 ecoregion (Plate 1) (USEPA, 2008). The different surficial geologic materials were presumed to have different phosphorus contents (because of different mineral compositions) and the different USEPA level 4 ecoregions were presumed to have different potential for weathering of surficial geologic material (because of differences in [geology,](http://en.wikipedia.org/wiki/Geology) [physiography,](http://en.wikipedia.org/wiki/Physiography) vegetation, climate, [hydrology,](http://en.wikipedia.org/wiki/Hydrology) terrestrial and aquatic [fauna,](http://en.wikipedia.org/wiki/Fauna) and soil). We developed a layer of "geo-eco regions," calculated an average concentration of phosphorus for each geo-eco region from the available bed sediment data (expressed in parts per million), and calculated a value for each SPARROW watershed by using a weighted average for each geo-eco region in that watershed.

Plate 1. U.S. EPA Level 3 and Level 4 Ecoregions in the United States Pacific Northwest.

Calibration Results

Figures S15 – S22 compare predicted stream loads to measured stream loads, predicted stream loads to model residuals, and predicted yields to model residuals.

Figure S15. Measured Stream Load Relative to Predicted Stream Load for the Total Nitrogen SPARROW Model Developed for the United States Pacific Northwest.

Figure S16. Residual Stream Load Relative to Stream Predicted Load to for the Total Nitrogen SPARROW Model Developed for the United States Pacific Northwest.

Figure S17. Residual Stream Load Relative to Predicted Yield for the Total Nitrogen SPARROW Model Developed for the United States Pacific Northwest.

Figure S18. Probability Plot of Standardized Residual Stream Load for the Total Nitrogen SPARROW Model Developed for the United States Pacific Northwest.

Figure S19. Measured Stream Load Relative to Predicted Stream Load for the Total Phosphorous SPARROW Model Developed for the United States Pacific Northwest.

Figure S20. Residual Stream Load Relative to Predicted Stream Load for the Total Phosphorous SPARROW Model Developed for the United States Pacific Northwest.

Figure S21. Residual Stream Load Relative to Predicted Yield for the Total Phosphorous SPARROW Model Developed for the United States Pacific Northwest.

Figure S22. Probability Plot of Standardized Residual Stream Load for the Total Phosphorous SPARROW Model Developed for the United States Pacific Northwest.

Prediction Results

Table S3 provides details for each of the regional and national studies that we compared to our predicted yields for watersheds dominated by different land cover types.

Table S2. Details for Regional and National Nutrient Yield Estimates.

Tables S3 – S5 show the SPARROW estimates of nutrient enrichment and a summary of nutrient-related 303(d) listings for each HUC8 watershed in the PNW.

Table S3. Estimation of Nutrient Enrichment and Summary of State 303(d) Listings Because of Impairment Related to Nutrient Enrichment within the Columbia River Basin.

Table S4. Estimation of Nutrient Enrichment and Summary of State 303(d) Listings Because of Impairment Related to Nutrient Enrichment within the Snake River Basin.

Table S5. Estimation of Nutrient Enrichment and Summary of State 303(d) Listings Because of Impairment Related to Nutrient Enrichment within the Puget Sound Basin, the Coast Ranges of Washington and Oregon Draining Directly to the Pacific Ocean, and the Closed Basins of Oregon.

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Notes:

na: not applicable, since there were no 303(d) listed streams for the state in the modeling region.

km: kilometers

¹California Environmental Protection Agency, 2002

²Idaho Department of Environmental Quality, 2002

³Montana Department of Environmental Quality, 2002

⁴Nevada Division of Environmental Protection, 2004

⁵Oregon Department of Environmental Quality, 2002

⁶Utah Department of Environmental Quality, 2006

⁷Washington Department of Ecology, 2002

⁸Wyoming Department of Environmental Quality, 2002

Notes:

TN: total nitrogen.

TP: total phosphorus.

NO3-N: nitrate, as nitrogen

N: nitrogen

NO3/NO2-N: nitrate plus nitrage, as nitrogen.

Phosphate-P: phosphate, as phosphorus.

SPARROW: Spatially Referenced Regressions on Watershed Attributes model.

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Figure S21. Residual Stream Load Relative to Predicted Yield for the Total Phosphorous SPARROW Model Developed for the United States Pacific Northwest.

natural log of predicted yield (kg/sqkm-yr)

Figure S22. Probability Plot of Standardized Residual Stream Load for the Total Phosphorous SPARROW Model Developed for the United States Pacific Northwest.

normal quantile

U. S. Environmental Protection Agency Level 3 and Level 4 Ecoregion Boundaries in the United States Pacific Northwest

Ecoregions—PLATE 1

bonate-Rich Mountains

Alpine

ntains hrublands