

# **A randomised cross-over cohort study of exposure to emissions from a road tunnel ventilation stack**



# **TITLE PAGE**

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A randomised cross-over cohort study of exposure to emissions from a road tunnel ventilation stack

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## **ABSTRACT**

## **Background & objective**

Road tunnels are increasingly important components of urban infrastructure. However knowledge of their health impact on surrounding communities is limited. Our objective was to estimate the short-term respiratory health effects of exposure to emissions from a road tunnel ventilation stack.

# **Methods**

ation stack.<br>
In a randomised cross-over cohort study in 36 volunteers who under<br>
narios in 2006 before the road tunnel opened, and in 2007 (n=27) an<br>
the tunnel opened. Exposure downwind of the stack was compared to<br>
to a We conducted a randomised cross-over cohort study in 36 volunteers who underwent three exposure scenarios in 2006 before the road tunnel opened, and in 2007 ( $n=27$ ) and 2008 (n=20) after the tunnel opened. Exposure downwind of the stack was compared to upwind of the stack and to a distant heavily trafficked location adjacent to a main road. Spirometry, exhaled nitric oxide (eNO), and symptom scores were measured repeatedly during each two hour exposure session.

# **Results**

Downwind locations were associated with increased reports of "dry nose" (score difference 0.36; 95% CI 0.09-0.63) compared with the control location (2006 vs 2007/08), but not with impaired lung function, increased airway inflammation or other symptoms. The heavily trafficked location was associated with significantly increased eNO (ratio=1.09; 95% CI: 1.04-1.14), eye (score difference 0.05; 95% CI: 0.01-0.10) and chest (score difference 0.21; 95% CI: 0.09-0.33) symptoms compared to the stack locations.

# **Conclusions**

There was no consistent evidence of adverse respiratory effects from short term exposures downwind of the tunnel ventilation stack, except for dry nose symptoms. However, the findings of increased airway inflammation and symptoms in subjects after only two hours exposure at the heavily trafficked location, is suggestive of detrimental effects of short-term exposures to traffic related air pollution.

# **ARTICLE SUMMARY**

# **Article focus**

- Road tunnels are being increasingly commissioned in major cities around the world to alleviate traffic on surface roads.
- There is little existing data on the health impacts of exposure to tunnel emissions on surrounding communities.
- Our objective was to estimate the short-term respiratory health effects of exposure to emissions from a road tunnel ventilation stack.

# **Key messages**

- There was no consistent evidence of adverse respiratory effects of being located downwind of the tunnel ventilation stack, except for an increase in dry nose symptoms.
- Short term (two hour) exposures to a heavily trafficked control site were associated with increased airway inflammation, eye and chest symptoms.
- ing communities.<br>
Etive was to estimate the short-term respiratory health effects of experts<br> **Form a road tunnel ventilation stack.**<br> **Form a road tunnel ventilation stack.**<br> **Form a road tunnel ventilation stack, except** • The findings suggest detrimental effects of short-term exposure to traffic related air pollution, but are not supportive of consistent adverse effects from short-term exposures downwind of the tunnel ventilation stacks.

# **Strengths and limitations of this study**

- Strengths of the study include: the randomised cross-over cohort design and blinding of participants; use of a heavily trafficked site as a positive control; the before/after nature of our study; the repeated measures within and across years; and the sensitivity analysis.
- A limitation of our study is the differential loss from the groups over time.

# **FUNDING STATEMENT**

The study was funded by the CRC for Asthma & Airways and the NSW Health Department.

## **INTRODUCTION**

Road tunnels are increasingly important components of urban infrastructure. However, few studies have investigated the adverse impacts of exposure to emissions from road tunnel portals or their ventilation stacks, or the beneficial effects of diverting traffic from surface roads into tunnels. This may be due to the difficulty in designing studies that distinguish between emissions from road tunnels and emissions from general surrounding traffic [1, 2]. Only one study has investigated the association between exposure to emissions from a road tunnel ventilation stack and health effects. That cross-sectional study found no evidence of increased respiratory or irritant symptoms in high exposure zones compared to low exposure zones [3].

ssions from road tunnels and emissions from general surrounding trady has investigated the association between exposure to emissions fr<br>dy has investigated the association between exposure to emissions fr<br>ation stack and h Several recent studies have reported designs that may be helpful for assessing this type of exposure. Each has investigated the respiratory $[4-8]$  or cardiovascular $[9-13]$  effects of very short term exposures (typically two hours or less) to traffic related air pollution (TRAP) and diesel emissions in real-world settings. All are characterised by their quasi-experimental study design in which participants were taken to locations (busy traffic, inner city with diesel only, quiet park like) or underwent realistic commutes (in buses, cars or bicycling) where the investigators expected elevated exposures or contrasts in exposures to TRAP to occur. However, these designs have not yet been applied to the assessment of new point sources, such as emissions from road tunnels and their ventilation stacks.

A proposal to build a new road tunnel in Sydney with emissions vented to the surface via ventilation stacks presented us with an opportunity to prospectively test the hypothesis that short-term exposure to emissions from the tunnel ventilation stacks has an adverse effect on respiratory health.

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## **METHODS**

 The study was conducted in Sydney, Australia between 2006 and 2008, and approved by the University of Sydney Human Research Ethics Committee (HREC) on 17 August, 2006.

## **Study design & subjects**

The study was a randomised, controlled cross-over trial. Thirty-six volunteers were recruited through a university, student and workplace networks, and the Woolcock Institute of Medical Research's (WIMR) Volunteer Database. Eligibility criteria included: non-smokers; ability to walk easily for 2 hours; able to participate in 2006 and 2007; sense of smell not compromised; and not pregnant.

as a randomised, controlled cross-over trial. Thirty-six volunteers we<br>iversity, student and workplace networks, and the Woolcock Institute<br>WIMR) Volunteer Database. Eligibility criteria included: non-smoke<br>or 2 hours; abl Study participants were randomised into three equal-size groups, which defined the order of exposure in the cross-over study. Three exposure conditions were tested: "downwind" of the LCT western ventilation stack (putative adverse exposure); "upwind" of the stack (negative control); and a "heavily trafficked site" (positive control). The three exposures took place over six days in October-November in 2006, before the tunnel opened, and during the same months in 2007 and 2008, after the tunnel opened. Each exposure comprised two morning sessions during peak traffic periods (7am-9am) on two consecutive weekdays (Tuesdays and Wednesdays) (Figure 1). There was a six day washout period between exposure conditions. The order of group allocation to exposure type was the same each year, except in 2008 where order of testing for two groups was changed twice due to changing wind conditions.

## **Exposure locations**

The tunnel ventilation stacks are located at its eastern and western ends, respectively. As the eastern stack is located close to tunnel feeder roads, attribution of an effect solely due to the

stack would be difficult. Hence, we decided to conduct the study at sites located around the western ventilation stack using predetermined walking routes (Figure 2). The heavily trafficked (positive control) site was located alongside a major urban commuter road close to Sydney's central business district and remote from the study area. The annual average daily traffic volume at this site was 64,418 vehicles per day during 2005.[14]

six sites in a 500-1000m circumference around the stack and two site<br>ajor commuter road. The stack sites were chosen based on the feasit<br>ing testing procedures and identification of 20 min walking routes, w<br>esidential stre We selected six sites in a 500-1000m circumference around the stack and two sites on either side of the major commuter road. The stack sites were chosen based on the feasibility of accommodating testing procedures and identification of 20 min walking routes, which were all on quiet residential streets with low traffic volumes, and were not undulating. Participants were asked to walk at a comfortable, even pace.

For each study day, two of the six stack sites were selected and defined as "upwind" or "downwind" sites based on the wind direction recorded at 6am that morning.

#### **Baseline assessment**

Prior to fieldwork we conducted spirometry, measured nitric oxide concentration in exhaled breath (eNO) and performed allergen skin prick tests (SPT) on participants in the WIMR laboratory. Participants completed a questionnaire recording asthma diagnosis, wheeze, medication use, respiratory symptoms, and odour perception,[15-17] as experienced in the preceding three months. Participants also completed a series of 14 visual analogue scales (VAS) to rate their perception of respiratory and chest, nose, throat and eye symptoms. They rated odour perception (strength and annoyance) on a VAS if they responded affirmatively to detecting odour. Symptoms were ranked on VAS scales from "no symptoms" (0) to "the

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worst I can imagine" (100). These data were used to determine pre-existing disease and to characterise the study population.

#### **Fieldwork**

Participants were either transported in a mini-van or made their own way to the field site. They were asked to wear N95 mask (Alpha Protech N95 particulate face mask certified to NIOSH standards) from when they left home each morning until measurements were made, to minimise exposure on route. They were asked to withhold long-acting bronchodilators for 12 hours and short-acting bronchodilators for 6 hours before fieldwork.

Examples 16 wear N95 mask (Alpha Protech N95 particulate face mask centrals) from when they left home each morning until measurements versions and the peer of the sector of hours before fieldwork.<br> **For the synchic short-a** Each exposure period commenced at 7am and lasted approximately two hours, incorporating two 20 minute periods of walking and two rest periods (Figure 1). During each exposure, subjects underwent three rounds of testing including completion of VAS forms, eNO measurement and spirometry, in that order, and at timepoints as indicated in Figure 1. Participants were allocated to pairs for testing and walking, with each pairing maintained (as much as possible) throughout the three years of testing.

#### *Clinical tests*

We used a hand-held spirometer (QRS Diagnostic: Z-5000-2668) linked to Office Medic software (Version 4.5i) to record forced expiratory volume in 1 second ( $FEV<sub>1</sub>$ ), and forced vital capacity (FVC). Spirometry was performed according to the American Thoracic Society/European Respiratory Society (ATS/ERS) criteria with technically unsatisfactory curves identified at the time of testing and again after fieldwork. These technically unsatisfactory curves were excluded from analysis. To improve reliability of the measures we chose the highest of two values reproducible within 150 ml for  $FEV<sub>1</sub>$  and FVC. Where not

reproducible, curves were rechecked for quality and, if acceptable, the highest measures for FEV<sub>1</sub> and FVC were selected.

We used an off-line technique to measure eNO, according to ATS/ERS criteria[18] at an expiratory flow rate of 200 ml/sec[19]. Exhaled breath was collected into a 3L chemically inert bag through a rotameter incorporating a filter and an NO scrubber on the inlet valve, to exclude the influence of ambient NO levels on personal NO concentrations. The NO concentration was measured within 24 hours of collection using a chemiluminescence analyser (ThermoEnvironmental  $42^{\circ}$ C).

The articular articular and an NO scrubber on the influence of ambient NO levels on personal NO concentrations. The nuas measured within 24 hours of collection using a chemiluminescermoEnvironmental 42°C).<br> **For a was mea** Skin prick tests were performed using airborne allergens for house dust mites (D. pteronyssinus, D. farinae), mould (Alternaria, Aspergillus), cat pelt, dog, cockroach, ryegrass, and grass mix (Hollister-Stier).[20] Wheal sizes were measured as the average of the largest diameter and its perpendicular at 15 minutes. Wheals larger than 3mm in diameter and larger than the negative control were considered positive. Subjects with any positive skin prick test were classified as atopic.

## *Air quality monitoring*

Air quality monitoring took place on each field testing day at each visited site, constituting a total of 18 measurement days. We collected samples for particulate matter ( $PM_{10}$ ,  $PM_{2.5}$  and PM<sub>1</sub>), volatile organic compounds (VOCS), total oxides of nitrogen (NOx), nitrogen dioxide  $(NO<sub>2</sub>)$  and nitrogen oxide  $(NO)$ .

We used Osiris instruments to collect PM data in 2006 and 2007 and Topas units for collection of PM data in 2008 due to the unavailability of Osiris instruments. Concentrations of PM data were recorded as 2 minute averages.

d sample over the two hours of fieldwork, providing an average read<br>
e analysed at the CSIRO Laboratory in 2006 and 2008 for 1,3-butadi<br>
nene, ethylbenzene, *m*- + *p*-xylene, *o*-xylene. Samples were analysed<br>
boratory in VOC samples were collected using a stainless steel canister and inlet designed to collect an evenly spaced sample over the two hours of fieldwork, providing an average reading. The samples were analysed at the CSIRO Laboratory in 2006 and 2008 for 1,3-butadiene, benzene, toluene, ethylbenzene, *m*- + *p*-xylene, *o*-xylene. Samples were analysed by an alternative laboratory in 2007 but the limit of detection (DL) was higher and given that many of the readings were below this DL we excluded the 2007 data.

Samples for ambient  $NOx/NO<sub>2</sub>$  were collected via grab samples every 15 minutes at each site using identical collection bags as those used for eNO collection. These were analysed (within 24 hours) as for eNO analysis.

We obtained continuously recorded data for 2006-2007 for  $NO<sub>2</sub> PM<sub>10</sub> PM<sub>2.5</sub>$  from two air quality monitors sited in elevated positions near the ventilation stacks and established by the tunnel proponents for statutory purposes.[21]

# **Wind direction**

Wind direction for each field day was determined using meteorological forecasts from the Bureau of Meteorology and by releasing slightly buoyant balloons filled with a mixture of air and helium from an elevated location near the western stack.

To validate our choice of testing sites, wind speed and direction data were obtained from three air quality monitoring stations located near the western ventilation stack, including the elevated site mentioned above. The other two were ground level stations, operated for all three study years. We constructed wind roses for the three monitoring sites using Lakes Environmental WRPlot View (version 6.5.1).

For dation of assignment of downwind and upwind study sites for each tusing wind rose data. Where the chosen locations were within  $\pm$  45 downwind or upwind direction, they were deemed to be accurate doins, respectively. Post-hoc validation of assignment of downwind and upwind study sites for each testing day was checked using wind rose data. Where the chosen locations were within  $\pm 45$  degrees of the measured downwind or upwind direction, they were deemed to be accurate downwind or upwind locations, respectively. This validation was used in the sensitivity analysis. Data for the heavily trafficked site (north or south side of the road) were included in the sensitivity analysis in the same way. Data were also excluded where wind speed was <1m/sec.

#### **Analysis**

#### *Health outcomes*

The distributions of eNO were skewed and so these values were log-transformed. All other dependent variables were analysed without transformation.

Principal components analysis (PCA) was conducted to reduce the 14 symptom scores to five factors which were found to explain 77% of the overall variance. The scree plot suggested that this was the optimum number of factors. Hence, a five factor structure was fitted with varimax rotation using data for all three years. The five factor scores were output for each observation and these were used as continuous measures of symptom intensity. The five symptom factors were 1) itchy/dry/sore eyes and blocked nose (EYE); 2) sore/dry throat, sore chest and runny nose (THROAT); 3) chest tightness, difficulty breathing (CHEST); 4)

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sneezing, cough, watery eyes; (SNEEZE) and, 5) dry nose symptoms (DRY NOSE). The distributions of VAS scores for "strong smell" and "annoying smell" were highly skewed and included a large proportion of zero values, and so were treated as binary outcomes: zero ("no") and greater than zero ("yes").

To estimate the air pollutant means we first calculated the daily means for each pollutant, and then calculated the mean pollutant level by location and year.

the air pollutant means we first calculated the daily means for each p<br>ed the mean pollutant level by location and year.<br>the data using an 'intention to treat' approach, based on the location<br>or each field testing session. We analysed the data using an 'intention to treat' approach, based on the locations chosen for each group for each field testing session. Data for timepoint 0 were excluded from the main analyses based on the assumption that exposure had not started when that measurement was made. We used a mixed effects linear regression model in which the fixed effects were location (upwind, downwind, control) and year (2006, 2007, 2008), and used contrast statements to test specific interactions between them. We also used contrast statements to estimate the overall difference (averaged over all years) in effect between the heavily trafficked location and the upwind and downwind locations combined. Since we expected that health outcomes would be worse at the heavily trafficked site, this served as a validation of the design. We included a random intercept term for subjects, with a variance components covariance structure, to account for repeated measures within subjects. Spirometric outcomes, eNO and the five symptom factors were separately included as outcome variables. Significance of the fixed effects was tested using F tests. Least squares means were estimated for each combination of year and location. Analyses were conducted using SAS Version 9.2 PROC MIXED procedures.

The binary variables "strong smell" and "annoying smell" were treated identically except that a generalised linear mixed model was fitted with Proc GLIMMIX using the adaptive quadrature estimation method, a logistic link and a binomial error distribution in SAS version 9.2.

## *Air pollutant concentrations*

We used a similar mixed effects linear regression model with contrast statements, to test the effects of location and year, and their interaction, on air pollutant measurements  $(NOx, NO<sub>2</sub>,$  $PM_{10}$ ,  $PM_{2.5}$ , and  $PM_1$ ), excluding VOCs where there were too few readings. A random intercept term for days was used to account for repeated measures. We conducted a sensitivity analysis for the air quality data by re-fitting the models using data that were validated using the wind-rose method, as described earlier.

*concentrations*<br>
milar mixed effects linear regression model with contrast statements<br>
ation and year, and their interaction, on air pollutant measurements<br>
(and PM<sub>1</sub>), excluding VOCs where there were too few readings. A For the elevated monitor data, we used an autoregressive error model with Yule-Walker estimates and stepwise regression to determine whether there was a difference in air quality between 2006 (pre tunnel) and 2007 (post-tunnel opening), by comparing daily averages for the two years. We adjusted this analysis for changes in regional air quality measured at three regulatory monitoring stations operated by the Department of Environment, Climate Change and Water (DECCW), including as a covariate the change in average daily readings from the DECCW sites combined. We tested up to 20 lags, residuals were plotted to check for normality, and the autocorrelation function (ACF) and partial autocorrelation function (PACF) were checked for autocorrelation. These models were implemented in SAS 9.2 using the AUTOREG procedure.

#### **Sample size and power calculation**

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The sample size for this study (36 subjects) was selected by reference to a study conducted in London that was able to detect significant within-subject differences in lung function in 60 subjects exposed to a high pollution area and a low pollution area [5], noting that our study had a greater number of repeated measures.

# **RESULTS**

6 of the 36 participants reported having had a diagnosis of asthma and<br>ing had wheeze, including 25% who had wheeze in the last three mo<br>rity of participants were atopic. Of the 36 subjects who started the st<br>ticipated in In 2006, 25% of the 36 participants reported having had a diagnosis of asthma and 42% reported having had wheeze, including 25% who had wheeze in the last three months (Table 1). The majority of participants were atopic. Of the 36 subjects who started the study in 2006, 26 (72%) participated in 2007 and 20 (56%) in 2008. Whilst the cohort characteristics remained relatively stable between 2006 and 2007, in 2008 the remaining cohort was less symptomatic in the three months prior to 2008 field testing (Table 1, Supplementary Material).

## **Health Outcomes**

 $FEV<sub>1</sub>$  and FVC were similar across locations, but were generally higher in 2006 compared to 2007 or 2008 (Table 2). Mean values for eNO were highest at the heavily trafficked (positive control) location for all years except for 2008 and did not differ between the downwind and upwind locations. EYE, THROAT and CHEST symptom scores were highest at the heavily trafficked location compared to the downwind or upwind locations during all years, with one exception (Table 2).

There was little evidence that the downwind locations, when compared with the heavily trafficked or upwind locations, were associated with adverse health outcomes in 2007 or

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2008, compared to 2006 (Table 3). The only exception was that changes in DRY NOSE

scores were higher for downwind vs heavily trafficked locations. Downwind locations were

associated with a decrease in reporting of chest symptoms for almost all year comparisons.







a Inhaled steroids=having taken either seretide or pulmicort b

 Question: Do you usually regard your eyes as being sensitive to cigarette smoke, smog, air conditioning or central heating?

 $\mathbf c$  Question: Have you had any of the following eye symptoms in the last 3 mths, when you did not have a cold or flu (eye redness, burning feeling, scratchiness, grittiness, dryness, watery eyes)?

d Any allergen  $SPT \geq 3$ mm

Location	Year	$FEV1$ (L)	FVC(L)	eNO (ppb) (median)	<b>EYE</b> symptoms score*	<b>THROAT</b> symptoms score*	<b>CHEST</b> symptoms score*	<b>SNEEZE</b> symptoms score*	<b>DRY NOSE</b> symptoms score*
<b>Heavily</b> trafficked	Average	2.77	3.71	13.6	0.086	0.030	0.149	0.393	$-0.047$
	2006	2.82	3.72	11.6	0.076	0.162	0.097	0.007	0.219
	2007	2.76	3.70	20.9	0.095	0.057	0.100	0.567	$-0.224$
	2008	2.71	3.70	11.1	0.095	$-0.215$	0.280	0.845	$-0.300$
Downwind	Average	2.78	3.73	12.1	0.046	$-0.055$	$-0.032$	0.387	$-0.158$
	2006	2.83	3.74	10.2	0.017	$-0.016$	0.001	$-0.076$	$-0.094$
	2007	2.78	3.75	17.8	0.035	$-0.037$	$-0.086$	0.813	$-0.089$
	2008	2.70	3.69	10.6	0.104	$-0.137$	$-0.027$	0.712	$-0.336$
<b>Upwind</b>	Average	2.79	3.75	12.8	0.007	$-0.144$	$-0.073$	0.118	$-0.030$
	2006	2.85	3.80	10.9	$-0.047$	$-0.059$	$-0.046$	$-0.109$	0.138
	2007	2.76	3.74	18.2	0.078	$-0.066$	$-0.169$	0.347	$-0.042$
	2008	2.73	3.69	11.3	0.024	$-0.358$	$-0.019$	0.258	$-0.294$
EYE symptoms: Itchy/dry/sore eyes and blocked nose THROAT symptoms: Sore/dry throat, sore chest & runny nose CHEST symptoms: Chest tightness, difficulty breathing SNEEZE symptoms: Sneezing, cough, watery eyes DRY NOSE symptoms: Dry nose									

**Table 2. Mean values for lung function and symptom scores by location and year** 

EYE symptoms: Itchy/dry/sore eyes and blocked nose THROAT symptoms: Sore/dry throat, sore chest & runny nose CHEST symptoms: Chest tightness, difficulty breathing SNEEZE symptoms: Sneezing, cough, watery eyes DRY NOSE symptoms: Dry nose

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# **Table 3. Contrasts\* in lung function, eNO and symptom scores between the downwind site and the other two sites (upwind and Heavily trafficked) adjusted for change from baseline (2006) to subsequent years**





<sup>a</sup> Contrasts are derived from mixed effects regression model in which location and year, and the interaction between them are fixed effects, and individual subject intercepts are a random effect. Main contrasts of interest are those included in the table and consist of downwind versus other exposures, as the downwind exposure is the primary exposure of interest in this study.

<sup>b</sup> Downwind vs (heavily trafficked and upwind) is the comparison of downwind versus heavily trafficked and upwind exposures combined.

 $\degree$  2007/08 vs 2006 represents the comparison of both post-tunnel years combined (2007/08) versus the pre-tunnel year (2006).

<sup>d</sup> Represents the overall comparison of the heavily trafficked site compared to the upwind and downwind sites combined (data from all years included), as we anticipated the heavily trafficked site to have higher air pollution levels and that health effects would be more prominent.

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Over the three year period as a whole, exposure at the heavily trafficked locations, compared with both the downwind and upwind stack locations, was associated with more airway inflammation (eNO) (ratio=1.09), and higher scores for EYE, CHEST, and THROAT symptoms (last row, Table 3). These findings support the validity of the study design and analysis.

The change in "smell strong" from 2006 to subsequent years was higher for downwind locations than the other two locations (Table 3). However, in general this pattern was not observed for "smell-annoying". Overall, both "smell-strong" and "smell-annoying" were higher at the heavily trafficked locations than the two stack locations (last row, Table 3).

# *Sensitivity analysis for health outcomes*

In the sensitivity analysis we excluded data which might have led to exposure misclassification due to unexpected wind direction, as described earlier. Data for: 5 "downwind" sites in 2006 and 2 in 2008; for 2 "upwind" sites in 2006 and 2 in 2008; and for one heavily trafficked site in 2006; were excluded.

"smell\_strong" from 2006 to subsequent years was higher for down<br>wo locations (Table 3). However, in general this pattern was not obs<br>g". Overall, both "smell-strong" and "smell-annoying" were higher<br>tions than the two sta In this analysis FVC was increased for downwind locations vs the other two locations, for most year comparisons (Table 2, Supplementary Material).  $FEV<sub>1</sub>$  was also increased for downwind vs heavily trafficked locations. Conversely, eNO was higher at downwind compared to upwind locations for 2007 vs 2006. The increased odds of "smell-strong" at downwind locations did not differ to the other locations in the sensitivity analysis. Overall, eNO levels, CHEST symptoms, "smell-strong" and "smell-annoying" were higher at the heavily trafficked locations compared to upwind or downwind locations, confirming the findings of the intention-to-treat analysis (Table 2, Supplementary Material).

### **Air pollutant levels**

Material). NOx and NO<sub>2</sub> ratios were lower for the downwind sites t<br>007 vs 2006. In contrast, the NO<sub>2</sub> ratio for 2008 vs 2006 was higher is<br>than at the heavily trafficked sites (Table 5). The change in PM (fo<br>2006 to subs As expected, the concentrations of all pollutants were higher at the heavily trafficked sites than the upwind or downwind sites. However, there was little difference in mean pollutant levels between the downwind and upwind locations around the ventilation stack (Table 4 and Table 3, Supplementary Material). NOx and  $NO<sub>2</sub>$  ratios were lower for the downwind sites than the other two sites for 2007 vs 2006. In contrast, the  $NO<sub>2</sub>$  ratio for 2008 vs 2006 was higher at the downwind sites than at the heavily trafficked sites (Table 5). The change in PM (for all size fractions) from 2006 to subsequent years did not vary by site. Overall, the change in pollutants was greatest at the heavily trafficked sites compared to the upwind and downwind sites combined.

In the sensitivity analysis the pollutant ratios for downwind locations compared to upwind locations increased for most pollutants supporting the validity of the exclusion criteria (Table 4, Supplementary Material).

NO2 and PM10 decreased at the western and eastern elevated monitoring sites between 2006 and 2007, after adjusting for changes in regional air quality (Table 6).

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# **Table 4. Mean pollutant levels by location (2006-2008 overall)**

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a Means and medians calculated by taking the mean of daily averages at each site. IQR, minimum and maximum values using all (not averaged) data.

b Grab sample for  $NO<sub>2</sub>$  and NOx collected at each site every 15 min and averaged per two hr exposure period.

c PM data collected every two min.

**For Formal Phillips** d One sample collected at site providing an average VOC reading per two hr exposure period. 2007 data not used in analysis due to different detection limits used by different laboratory. 95<sup>th</sup> percentile not reported for VOCs as sample size too small; 95<sup>th</sup> percentile equivalent to maximum reading.

**Table 5. Ratios of pollutant levels<sup>a</sup> contrasting the downwind site and the other two sites (upwind and heavily trafficked) adjusted for change from baseline (2006) to subsequent years** 



<sup>a</sup> Ratios are derived from a mixed effects regression model in which location and year, and the interaction between them are fixed effects, and individual subject intercepts are a random effect. Main contrasts of interest are those included in the table and consist of downwind versus other exposures, as the downwind exposure is the primary exposure of interest in this study.

<sup>b</sup> Downwind vs (heavily trafficked and upwind) is the comparison of downwind versus heavily trafficked and upwind exposures combined.

 $\degree$  2007/08 vs 2006 represents the comparison of both post-tunnel years combined (2007/08) versus the pre-tunnel year (2006).

<sup>d</sup> Represents the overall comparison of the heavily trafficked site compared to the two stack sites (downwind and upwind) combined, (data from all years included).

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**Table 6. Mean pollutant concentrations for elevated monitors by year and difference in pollutant concentration from post-tunnel year compared to pre-tunnel year** 

a. Represents the change in pollutant concentration between pre-tunnel (25 Mar 06-24 Mar 07) and post-tunnel (25 Mar 2007-24 Mar 2008)

b. Adjusted for change in air quality at the 3 regional sites

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## **DISCUSSION**

We did not detect any consistent evidence of detrimental effects on lung function, airway inflammation or respiratory symptoms attributable to short-term exposure to emissions from the tunnel ventilation stack with the exception of increased reporting of dry nose symptoms. This absence of any adverse health effect might be explained by the fact that we did not measure higher pollutant concentrations at the downwind stack sites after the tunnel opened. The finding of increased pollutant levels and increased symptoms and airway inflammation during periods of exposure at the heavily trafficked site supports previous evidence of the detrimental effects of short-term exposures to TRAP.

For pollutant concentrations at the downwind stack sites after the tuninal of increased pollutant levels and increased symptoms and airway inflused so fexposure at the heavily trafficked site supports previous evident offe This study relied on the proposition that the downwind sites we selected were, in fact, exposed to emissions from the tunnel ventilation stack during peak traffic periods. For a downwind site to be impacted by the emission plume from the stack the plume needed to be mixed at ground level prior to reaching the site and subjected to consistent wind direction during the sampling period. We used weather balloons early on each study morning in order to select the most appropriate downwind site. However, it is possible that we were not always able to accurately predict the ground level location of the plume. The sensitivity analysis we conducted was designed to address this potential weakness. The findings of that analysis support the original analysis.

Our finding that eNO, an indicator of airway inflammation, was increased at the heavily trafficked site is consistent with evidence of increased airway inflammation after exposure to TRAP (or diesel) in previous studies. These studies have varied in their measured end-points but include "real-world" exposure studies investigating neutrophilic airway inflammation[4, 5], associations between eNO and  $PM<sub>2.5</sub>$ , black carbon and fine particle counts[6] and

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chamber studies reporting increased neutrophilic airway inflammation,[22-25] increases in pro-inflammatory cytokines IL-6[26] and IL-8,[25] increased exhaled carbon monoxide (CO),[22] and increased airway resistance[26, 27].

The absence of effects on spirometric function, even at the heavily-trafficked site, is consistent with findings of chamber studies of diesel exhaust,[22, 24-27] high particle concentrations, [28] and NO<sub>2</sub> [29] an in-tunnel study[30] and a commuter study[8]. However, our findings are inconsistent with the London study of diesel exhaust exposures which demonstrated small reductions in  $FEV_1$  after 1-2 hr exposures.[5]

th findings of chamber studies of diesel exhaust, [22, 24-27] high parts, [28] and NO<sub>2</sub>, [29] an in-tunnel study[30] and a commuter study[8 are inconsistent with the London study of diesel exhaust exposures v and l reduct Strengths of the study include the randomised cross-over design and blinding of participants by using upwind and downwind stack sites. It is unlikely that the participants were aware of the upwind/downwind status of their location on any given day. The study was further strengthened by including the heavily trafficked sites as a positive control. The finding that pollutant levels were higher, and that subjects had greater airway inflammation, eye and chest symptom scores at this site, support the validity of the study methodology. The before/after nature of our study and repeated measures within and across years also strengthened the study design.

A limitation of our study is the differential loss from the groups over time. Common reasons for non-participation in follow-up years were travel overseas, residence relocation and changing work practices, all factors unrelated to the experience of study participation.

In conclusion, while we found that this ventilation stack, under current operating conditions, had little measurable short-term health impact, we did demonstrate adverse health effects

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attributable to short term exposures to TRAP at heavily trafficked locations. These findings support evidence from existing studies of the adverse health effects attributable to short-term exposures to TRAP.

## **Competing interests:**

Nectarios Rose was employed with the NSW Health Department on the Biostatistics Training Scheme at the time the study was conducted.

# **Authors' Contributions:**

See was employed with the NSW Health Department on the Biostatis<br>
e time the study was conducted.<br>
Intributions:<br>
I and participated in all aspects of the study including study design,<br>
seruiting subjects, field work, stat CC organised and participated in all aspects of the study including study design, ethics clearances, recruiting subjects, field work, statistical analysis and led manuscript preparation. WE was involved in all field work. WX contributed to study design and data analysis. WL conducted air quality exposure assessment measurements in the field and provided advice on air quality measurement. NR contributed biostatistical expertise. MR conducted air quality measurements in the field and wind rose analyses. GM participated in the study design, statistical analysis, and provided general oversight of the study. All authors contributed to the preparation of the manuscript.

## **Acknowledgments:**

We are grateful to the volunteers for their participation in this study. Professor Bert Brunekreef, Vicky Sheppeard and Michael Staff gave initial advice on the study design. We acknowledge the invaluable assistance of the following research staff: Brett Toelle, Elena Belousova, Adriana Cortes-Waterman, Kitty Ng, Angelica Ordonez, Paula Garay, Adrian Forero, Jennifer Manning, Renate Thielbeer. The study was overseen by a Scientific Advisory Group established by the NSW Health Department which provided advice on study

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design, methodology, analysis and interpretation. It consisted of Professor Bruce Armstrong, Professor Judy Simpson, Associate Professor Bin Jalaludin, Dr Peter Lewis, and Dr Gavin Fisher.

## **Data sharing:**

and risk of identification is low. Ethics clearance will need to be sought.

Participants did not provide consent for data sharing but the presented data are anonymised<br>and risk of identification is low. Ethics clearance will need to be sought.

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**Figure Legends** 

Figure 1 Schematic of fieldwork

Figure 2 Study area and field sites




Figure 1 Schematic of fieldwork 254x190mm (96 x 96 DPI)



Figure 2 Study area and field sites 254x190mm (96 x 96 DPI)



**Table 1 Characteristics of cohort members participating in each study year** 

<sup>a</sup><br>Measurements made at WIMR lung function laboratory<br> $\frac{b}{c}$ 

Field measurements made at field sites

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**Table 2 Sensitivity analysis**a **for contrasts**b **in lung function, eNO and symptom scores between the downwind site and the other two sites (upwind and heavily trafficked) adjusted for change from baseline (2006) to subsequent years.** 





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**Table 3 Mean pollutant levels by location and year**

Pollutant	Year	Location	<b>Total</b> samples	$Meana$	Median <sup>a</sup>	<b>IQR</b>	<b>Minimum</b>	<b>Maximum</b>	$95^{\text{th}}$ percentile
Nitrogen oxides <sup>b</sup>									
NOx	2006	Heavily trafficked	55	112.0	107.9	49.9-142.9	28.7	426.0	269.0
		Downwind	56	34.6	37.2	16.7-50.5	5.7	89.7	79.4
		Upwind	51	35.4	31.1	12.6-47.3	2.4	100.8	91.0
	2007	Heavily trafficked	51	239.8	128.6	67.1-234.0	14.8	1912.0	784.0
		Downwind	52	30.2	16.2	8.4-44.5	4.6	117.9	96.8
		Upwind	46	37.0	20.7	12.2-57.2	3.5	152.4	93.1
	2008	Heavily trafficked	54	26.7	16.8	9.9-28.6	5.5	115.7	98.0
		Downwind	51	6.8	7.1	$4.3 - 8.5$	2.1	18.9	13.0
		Upwind	53	6.2	5.1	$3.6 - 8.2$	$2.5\,$	16.7	12.5
NO2	2006	Heavily trafficked	55	4.3	4.1	$3.0 - 5.2$	1.3	12.0	$\ \, 8.0$
		Downwind	56	$3.2\,$	$3.2\,$	$2.1 - 4.3$	1.3	5.7	5.1
		Upwind	51	2.4	2.1	$1.7 - 3.0$	0.5	16.0	4.2
	2007	Heavily trafficked	51	13.5	13.6	5.0-17.6	1.4	69.0	31.0
		Downwind	52	4.9	5.1	$1.6 - 7.1$	0.1	18.2	9.5
		Upwind	47	5.9	4.8	$2.9 - 6.8$	0.1	15.7	13.8
	2008	Heavily trafficked	54	0.6	$1.5$	$0.1 - 1.7$	0.1	4.1	3.2
		Downwind	51	$1.1\,$	$1.1\,$	$0.8 - 1.4$	0.1	2.7	2.2
		Upwind	53	$1.1\,$	$1.0\,$	$0.7 - 1.4$	0.1	2.3	2.0
Particulate matter <sup>c</sup> <b>PM10</b>	2006	Heavily trafficked	372	74.2	81.2	46.3-99.1	17.9	168.1	118.7
		Downwind	374	58.1	57.0	33.4-84.0	15.1	195.1	106.7



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<sup>a</sup> Means and medians calculated by taking the mean of daily averages at each site. IQR, minimum and maximum values using all (not averaged) data.

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For beam-period collected revery two minutes.<br>
FM duata collected every two minutes.<br> **FOR USE CONTO GRAD SATE AND SATE OF A CONTO C** <sup>c</sup> PM data collected every two minutes.

d One VOC sample collected at each site to provide average reading for each two hour exposure period. 2007 data not used in analysis due to different detection limits used by different laboratory.

95<sup>th</sup> percentile not reported for VOCs as sample size too small;  $95<sup>th</sup>$  percentile equivalent to maximum reading.





<sup>a</sup> 5 downwind sites and 2 upwind sites in 2006, 2 downwind sites and 2 upwind sites in 2008, and 1 heavily trafficked site in 2006, were excluded for the sensitivity analysis

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# **A randomised cross-over cohort study of exposure to emissions from a road tunnel ventilation stack**



# **TITLE PAGE**

#### **Title**

A randomised cross-over cohort study of exposure to emissions from a road tunnel ventilation stack

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# **Keywords**

Road tunnel, respiratory, emissions, stack, acute exposures

**Word count:** 3698

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#### **ABSTRACT**

#### **Background & objective**

Road tunnels are increasingly important components of urban infrastructure. However knowledge of their health impact on surrounding communities is limited. Our objective was to estimate the short-term respiratory health effects of exposure to emissions from a road tunnel ventilation stack.

### **Methods**

ation stack.<br>
In a randomised cross-over cohort study in 36 volunteers who under<br>
narios in 2006 before the road tunnel opened, and in 2007 (n=27) an<br>
the tunnel opened. Exposure downwind of the stack was compared to<br>
to a We conducted a randomised cross-over cohort study in 36 volunteers who underwent three exposure scenarios in 2006 before the road tunnel opened, and in 2007 ( $n=27$ ) and 2008 (n=20) after the tunnel opened. Exposure downwind of the stack was compared to upwind of the stack and to a distant heavily trafficked location adjacent to a main road. Spirometry, exhaled nitric oxide (eNO), and symptom scores were measured repeatedly during each two hour exposure session.

### **Results**

Downwind locations were associated with increased reports of "dry nose" (score difference 0.36; 95% CI 0.09-0.63) compared with the control location (2006 vs 2007/08), but not with impaired lung function, increased airway inflammation or other symptoms. The heavily trafficked location was associated with significantly increased eNO (ratio=1.09; 95% CI: 1.04-1.14), eye (score difference 0.05; 95% CI: 0.01-0.10) and chest (score difference 0.21; 95% CI: 0.09-0.33) symptoms compared to the stack locations.

### **Conclusions**

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There was no consistent evidence of adverse respiratory effects from short term exposures downwind of the tunnel ventilation stack, except for dry nose symptoms. However, the findings of increased airway inflammation and symptoms in subjects after only two hours exposure at the heavily trafficked location, is suggestive of detrimental effects of short-term exposures to traffic related air pollution.

# **ARTICLE SUMMARY**

# **Article focus**

- Road tunnels are being increasingly commissioned in major cities around the world to alleviate traffic on surface roads.
- There is little existing data on the health impacts of exposure to tunnel emissions on surrounding communities.
- Our objective was to estimate the short-term respiratory health effects of exposure to emissions from a road tunnel ventilation stack.

## **Key messages**

- There was no consistent evidence of adverse respiratory effects of being located downwind of the tunnel ventilation stack, except for an increase in dry nose symptoms.
- Short term (two hour) exposures to a heavily trafficked control site were associated with increased airway inflammation, eye and chest symptoms.
- ing communities.<br>
Etive was to estimate the short-term respiratory health effects of experts<br> **Form a road tunnel ventilation stack.**<br> **Form a road tunnel ventilation stack.**<br> **Form a road tunnel ventilation stack, except** • The findings suggest detrimental effects of short-term exposure to traffic related air pollution, but are not supportive of consistent adverse effects from short-term exposures downwind of the tunnel ventilation stacks.

# **Strengths and limitations of this study**

- Strengths of the study include: the randomised cross-over cohort design and blinding of participants; use of a heavily trafficked site as a positive control; the before/after nature of our study; the repeated measures within and across years; and the sensitivity analysis.
- A limitation of our study is the differential loss from the groups over time.

# **FUNDING STATEMENT**

The study was funded by the CRC for Asthma & Airways and the NSW Health Department.

#### **INTRODUCTION**

Road tunnels are increasingly important components of urban infrastructure. However, few studies have investigated the adverse impacts of exposure to emissions from road tunnel portals or their ventilation stacks, or the beneficial effects of diverting traffic from surface roads into tunnels. This may be due to the difficulty in designing studies that distinguish between emissions from road tunnels and emissions from general surrounding traffic [1, 2]. Only one study has investigated the association between exposure to emissions from a road tunnel ventilation stack and health effects. That cross-sectional study found no evidence of increased respiratory or irritant symptoms in high exposure zones compared to low exposure zones [3].

ssions from road tunnels and emissions from general surrounding trady has investigated the association between exposure to emissions fr<br>dy has investigated the association between exposure to emissions fr<br>ation stack and h Several recent studies have reported designs that may be helpful for assessing this type of exposure. Each has investigated the respiratory $[4-8]$  or cardiovascular $[9-13]$  effects of very short term exposures (typically two hours or less) to traffic related air pollution (TRAP) and diesel emissions in real-world settings. All are characterised by their quasi-experimental study design in which participants were taken to locations (busy traffic, inner city with diesel only, quiet park like) or underwent realistic commutes (in buses, cars or bicycling) where the investigators expected elevated exposures or contrasts in exposures to TRAP to occur. However, these designs have not yet been applied to the assessment of new point sources, such as emissions from road tunnels and their ventilation stacks.

A proposal to build a new road tunnel in Sydney with emissions vented to the surface via ventilation stacks presented us with an opportunity to prospectively test the hypothesis that short-term exposure to emissions from the tunnel ventilation stacks has an adverse effect on respiratory health.

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### **METHODS**

 The study was conducted in Sydney, Australia between 2006 and 2008, and approved by the University of Sydney Human Research Ethics Committee (HREC) on 17 August, 2006.

#### **Study design & subjects**

The study was a randomised, controlled cross-over trial. Thirty-six volunteers were recruited through a university, student and workplace networks, and the Woolcock Institute of Medical Research's (WIMR) Volunteer Database. Eligibility criteria included: non-smokers; ability to walk easily for 2 hours; able to participate in 2006 and 2007; sense of smell not compromised; and not pregnant.

as a randomised, controlled cross-over trial. Thirty-six volunteers we<br>iversity, student and workplace networks, and the Woolcock Institute<br>WIMR) Volunteer Database. Eligibility criteria included: non-smoke<br>or 2 hours; abl Study participants were randomised into three equal-size groups, which defined the order of exposure in the cross-over study. Three exposure conditions were tested: "downwind" of the LCT western ventilation stack (putative adverse exposure); "upwind" of the stack (negative control); and a "heavily trafficked site" (positive control). The three exposures took place over six days in October-November in 2006, before the tunnel opened, and during the same months in 2007 and 2008, after the tunnel opened, to account for seasonal influences. Data collection extended to 2008, the second follow-up year, as it was expected that traffic volumes in the tunnel would be increased between 2007 and 2008. Traffic volumes in November 2007 and 2008 were 49,218 vehicles per day (vpd) and 58,218 vpd respectively. Each exposure comprised two morning sessions during peak traffic periods (7am-9am) on two consecutive weekdays (Tuesdays and Wednesdays) (Figure 1). There was a six day washout period between exposure conditions. The order of group allocation to exposure type was the same each year, except in 2008 where order of testing for two groups was changed twice due to changing wind conditions.

#### **Exposure locations**

The tunnel ventilation stacks are located at its eastern and western ends, respectively. As the eastern stack is located close to tunnel feeder roads, attribution of an effect solely due to the stack would be difficult. Hence, we decided to conduct the study at sites located around the western ventilation stack using predetermined walking routes (Figure 2). The heavily trafficked (positive control) site was located alongside a major urban commuter road close to Sydney's central business district and remote from the study area. The annual average daily traffic volume at this site was 64,418 vehicles per day during 2005.[14]

ilation stack using predetermined walking routes (Figure 2). The heasitive control) site was located alongside a major urban commuter r tratal business district and remote from the study area. The annual ave e at this site We selected six sites in a 500-1000m circumference around the stack and two sites on either side of the major commuter road. The stack sites were chosen based on the feasibility of accommodating testing procedures and identification of 20 min walking routes, which were all on quiet residential streets with low traffic volumes, and were not undulating. Participants were asked to walk at a comfortable, even pace.

For each study day, two of the six stack sites were selected and defined as "upwind" or "downwind" sites based on the wind direction recorded at 6am that morning.

#### **Baseline assessment**

Prior to fieldwork we conducted spirometry, measured nitric oxide concentration in exhaled breath (eNO) and performed allergen skin prick tests (SPT) on participants in the WIMR laboratory. Participants completed a questionnaire recording asthma diagnosis, wheeze, medication use, respiratory symptoms, and odour perception,[15-17] as experienced in the preceding three months. Participants also completed a series of 14 visual analogue scales

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(VAS) to rate their perception of respiratory and chest, nose, throat and eye symptoms. They rated odour perception (strength and annoyance) on a VAS if they responded affirmatively to detecting odour. Symptoms were ranked on VAS scales from "no symptoms" (0) to "the worst I can imagine" (100). These data were used to determine pre-existing disease and to characterise the study population.

### **Fieldwork**

Participants were either transported in a mini-van or made their own way to the field site. They were asked to wear N95 mask (Alpha Protech N95 particulate face mask certified to NIOSH standards) from when they left home each morning until measurements were made, to minimise exposure on route. They were asked to withhold long-acting bronchodilators for 12 hours and short-acting bronchodilators for 6 hours before fieldwork.

were either transported in a mini-van or made their own way to the fisked to wear N95 mask (Alpha Protech N95 particulate face mask ce lards) from when they left home each morning until measurements v exposure on route. Th Each exposure period commenced at 7am and lasted approximately two hours, incorporating two 20 minute periods of walking and two rest periods (Figure 1). During each exposure, subjects underwent three rounds of testing including completion of VAS forms, eNO measurement and spirometry, in that order, and at timepoints as indicated in Figure 1. Participants were allocated to pairs for testing and walking, with each pairing maintained (as much as possible) throughout the three years of testing.

### *Clinical tests*

We used a hand-held spirometer (QRS Diagnostic: Z-5000-2668) linked to Office Medic software (Version 4.5i) to record forced expiratory volume in 1 second ( $FEV<sub>1</sub>$ ), and forced vital capacity (FVC). Spirometry was performed according to the American Thoracic Society/European Respiratory Society (ATS/ERS) criteria with technically unsatisfactory

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curves identified at the time of testing and again after fieldwork. These technically unsatisfactory curves were excluded from analysis. To improve reliability of the measures we chose the highest of two values reproducible within 150 ml for  $FEV<sub>1</sub>$  and  $FVC$ . Where not reproducible, curves were rechecked for quality and, if acceptable, the highest measures for FEV<sub>1</sub> and FVC were selected.

off-line technique to measure eNO, according to ATS/ERS criteria[1]<br>bw rate of 200 ml/sec[19]. Exhaled breath was collected into a 3L ch<br>ough a rotameter incorporating a filter and an NO scrubber on the inl<br>filuence of amb We used an off-line technique to measure eNO, according to ATS/ERS criteria<sup>[18]</sup> at an expiratory flow rate of 200 ml/sec[19]. Exhaled breath was collected into a 3L chemically inert bag through a rotameter incorporating a filter and an NO scrubber on the inlet valve, to exclude the influence of ambient NO levels on personal NO concentrations. The NO concentration was measured within 24 hours of collection using a chemiluminescence analyser (ThermoEnvironmental  $42^{\circ}$ C).

Skin prick tests were performed using airborne allergens for house dust mites (D. pteronyssinus, D. farinae), mould (Alternaria, Aspergillus), cat pelt, dog, cockroach, ryegrass, and grass mix (Hollister-Stier).[20] Wheal sizes were measured as the average of the largest diameter and its perpendicular at 15 minutes. Wheals larger than 3mm in diameter and larger than the negative control were considered positive. Subjects with any positive skin prick test were classified as atopic.

### *Air quality monitoring*

Air quality monitoring took place on each field testing day at each visited site, constituting a total of 18 measurement days. We collected samples for particulate matter ( $PM<sub>10</sub>$ ,  $PM<sub>2.5</sub>$  and  $PM<sub>1</sub>$ ), volatile organic compounds (VOCS), total oxides of nitrogen (NOx), nitrogen dioxide  $(NO<sub>2</sub>)$  and nitrogen oxide  $(NO)$ .

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We used Osiris instruments to collect PM data in 2006 and 2007 and Topas units for collection of PM data in 2008 due to the unavailability of Osiris instruments. Concentrations of PM data were recorded as 2 minute averages.

s were collected using a stainless steel canister and inlet designed to d sample over the two hours of fieldwork, providing an average read analysed at the CSIRO Laboratory in 2006 and 2008 for 1,3-butadinene, ethylbenzene VOC samples were collected using a stainless steel canister and inlet designed to collect an evenly spaced sample over the two hours of fieldwork, providing an average reading. The samples were analysed at the CSIRO Laboratory in 2006 and 2008 for 1,3-butadiene, benzene, toluene, ethylbenzene, *m*- + *p*-xylene, *o*-xylene. Samples were analysed by an alternative laboratory in 2007 but the limit of detection (DL) was higher and given that many of the readings were below this DL we excluded the 2007 data.

Samples for ambient  $NOX/NO<sub>2</sub>$  were collected via grab samples every 15 minutes at each site using identical collection bags as those used for eNO collection. These were analysed (within 24 hours) as for eNO analysis.

We obtained continuously recorded data for 2006-2007 for  $\overline{NO_2}$  PM<sub>10</sub>, PM<sub>2.5</sub> from two air quality monitors sited in elevated positions near the ventilation stacks and established by the tunnel proponents for statutory purposes.[21]

# **Wind direction**

Wind direction for each field day was determined using meteorological forecasts from the Bureau of Meteorology and by releasing slightly buoyant balloons filled with a mixture of air and helium from an elevated location near the western stack.

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To validate our choice of testing sites, wind speed and direction data were obtained from three air quality monitoring stations located near the western ventilation stack, including the elevated site mentioned above. The other two were ground level stations, operated for all three study years. We constructed wind roses for the three monitoring sites using Lakes Environmental WRPlot View (version 6.5.1).

For dation of assignment of downwind and upwind study sites for each tusing wind rose data. Where the chosen locations were within  $\pm$  45 downwind or upwind direction, they were deemed to be accurate doins, respectively. Post-hoc validation of assignment of downwind and upwind study sites for each testing day was checked using wind rose data. Where the chosen locations were within  $\pm 45$  degrees of the measured downwind or upwind direction, they were deemed to be accurate downwind or upwind locations, respectively. This validation was used in the sensitivity analysis. Data for the heavily trafficked site (north or south side of the road) were included in the sensitivity analysis in the same way. Data were also excluded where wind speed was <1m/sec.

#### **Analysis**

#### *Health outcomes*

The distributions of eNO were skewed and so these values were log-transformed. All other dependent variables were analysed without transformation.

Principal components analysis (PCA) was conducted to reduce the 14 symptom scores to five factors which were found to explain 77% of the overall variance. The scree plot suggested that this was the optimum number of factors. Hence, a five factor structure was fitted with varimax rotation using data for all three years. The five factor scores were output for each observation and these were used as continuous measures of symptom intensity. The five symptom factors were 1) itchy/dry/sore eyes and blocked nose (EYE); 2) sore/dry throat, sore chest and runny nose (THROAT); 3) chest tightness, difficulty breathing (CHEST); 4)

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sneezing, cough, watery eyes; (SNEEZE) and, 5) dry nose symptoms (DRY NOSE). The distributions of VAS scores for "strong smell" and "annoying smell" were highly skewed and included a large proportion of zero values, and so were treated as binary outcomes: zero ("no") and greater than zero ("yes").

To estimate the air pollutant means we first calculated the daily means for each pollutant, and then calculated the mean pollutant level by location and year.

the air pollutant means we first calculated the daily means for each p<br>ed the mean pollutant level by location and year.<br>the data using an 'intention to treat' approach, based on the location<br>or each field testing session. We analysed the data using an 'intention to treat' approach, based on the locations chosen for each group for each field testing session. Data for timepoint 0 were excluded from the main analyses based on the assumption that exposure had not started when that measurement was made. We used a mixed effects linear regression model in which the fixed effects were location (upwind, downwind, control) and year (2006, 2007, 2008), and used contrast statements to test specific interactions between them. We also used contrast statements to estimate the overall difference (averaged over all years) in effect between the heavily trafficked location and the upwind and downwind locations combined. Since we expected that health outcomes would be worse at the heavily trafficked site, this served as a validation of the design. We included a random intercept term for subjects, with a variance components covariance structure, to account for repeated measures within subjects. Spirometric outcomes, eNO and the five symptom factors were separately included as outcome variables. Significance of the fixed effects was tested using F tests. Least squares means were estimated for each combination of year and location. Analyses were conducted using SAS Version 9.2 PROC MIXED procedures.

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The binary variables "strong smell" and "annoying smell" were treated identically except that a generalised linear mixed model was fitted with Proc GLIMMIX using the adaptive quadrature estimation method, a logistic link and a binomial error distribution in SAS version 9.2. Examination of the distribution of residuals indicated that there were no major departures from normality and transformation of the data was not required.

### *Air pollutant concentrations*

We used a similar mixed effects linear regression model with contrast statements, to test the effects of location and year, and their interaction, on air pollutant measurements  $(NOx, NO<sub>2</sub>,$  $PM_{10}$ ,  $PM_{2.5}$ , and  $PM_1$ ), excluding VOCs where there were too few readings. A random intercept term for days was used to account for repeated measures. We conducted a sensitivity analysis for the air quality data by re-fitting the models using data that were validated using the wind-rose method, as described earlier.

*Concentrations*<br> **EXECUTE:** This end their interaction, on air pollutant measurements (and PM<sub>1</sub>), excluding VOCs where there were too few readings. A ra for days was used to account for repeated measures. We conducted For the elevated monitor data, we used an autoregressive error model with Yule-Walker estimates and stepwise regression to determine whether there was a difference in air quality between 2006 (pre tunnel) and 2007 (post-tunnel opening), by comparing daily averages for the two years. We adjusted this analysis for changes in regional air quality measured at three regulatory monitoring stations operated by the Department of Environment, Climate Change and Water (DECCW), including as a covariate the change in average daily readings from the DECCW sites combined. We tested up to 20 lags, residuals were plotted to check for normality, and the autocorrelation function (ACF) and partial autocorrelation function (PACF) were checked for autocorrelation. These models were implemented in SAS 9.2 using the AUTOREG procedure.

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### **Sample size and power calculation**

The sample size for this study (36 subjects) was selected by reference to a study conducted in London that was able to detect significant within-subject differences in lung function in 60 subjects exposed to a high pollution area and a low pollution area [5], noting that our study had a greater number of repeated measures.

### **RESULTS**

of the 36 participants reported having had a diagnosis of asthma and ming had wheeze, including 25% who had wheeze in the last three mority of participants were atopic. Of the 36 subjects who started the st ticipated in 20 In 2006, 25% of the 36 participants reported having had a diagnosis of asthma and 42% reported having had wheeze, including 25% who had wheeze in the last three months (Table 1). The majority of participants were atopic. Of the 36 subjects who started the study in 2006, 26 (72%) participated in 2007 and 20 (56%) in 2008. Whilst the cohort characteristics remained relatively stable between 2006 and 2007, in 2008 the remaining cohort was less symptomatic in the three months prior to 2008 field testing (Table 1, Supplementary Material).

## **Health Outcomes**

 $FEV<sub>1</sub>$  and FVC were similar across locations, but were generally higher in 2006 compared to 2007 or 2008 (Table 2). Mean values for eNO were highest at the heavily trafficked (positive control) location for all years except for 2008 and did not differ between the downwind and upwind locations. EYE, THROAT and CHEST symptom scores were highest at the heavily trafficked location compared to the downwind or upwind locations during all years, with one exception (Table 2).

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There was little evidence that the downwind locations, when compared with the heavily trafficked or upwind locations, were associated with adverse health outcomes in 2007 or 2008, compared to 2006 (Table 3). The only exception was that changes in DRY NOSE scores were higher for downwind vs heavily trafficked locations. Downwind locations were associated with a decrease in reporting of chest symptoms for almost all year comparisons.



## **Table 1. Baseline characteristics as measured at recruitment in 2006**

a Inhaled steroids=having taken either seretide or pulmicort b

 Question: Do you usually regard your eyes as being sensitive to cigarette smoke, smog, air conditioning or central heating?



Location	Year	No. of obs <sup>a</sup>	$FEV1$ (L)	FVC(L)	$eNO$ (ppb) (median)	<b>EYE</b> symptoms score*	<b>THROAT</b> symptoms score*	<b>CHEST</b> symptoms score*	<b>SNEEZE</b> symptoms score*	<b>DRY NOSE</b> symptoms score*
<b>Heavily</b> trafficked	Average	463	2.72(0.80)	3.64(1.00)	17.6(13.0)	$-0.048(0.714)$	0.052(1.187)	0.120(1.098)	0.351(1.590)	$-0.158(1.076)$
	2006	208	2.84(0.81)	3.70(1.00)	15.2(12.1)	$-0.022(0.881)$	0.120(1.146)	0.095(1.120)	0.059(1.153)	0.086(1.146)
	2007	156	2.67(0.83)	3.62(1.01)	24.34 (14.4)	$-0.054(0.587)$	0.108(1.305)	0.137(1.084)	0.547(1.929)	$-0.409(1.003)$
	2008	99	2.58(0.69)	3.53(0.89)	12.2(6.6)	$-0.092(0.467)$	$-0.182(1.049)$	0.143(1.082)	0.651(1.692)	$-0.268(0.923)$
<b>Downwind</b>	Average	461	2.68(0.79)	3.62(0.99)	15.2(10.9)	$-0.033(0.840)$	$-0.012(0.962)$	$-0.071(0.874)$	0.365(1.867)	$-0.248(1.025)$
	2006	207	2.81(0.84)	3.73(1.05)	12.8(9.5)	0.033(1.073)	$-0.031(0.873)$	$-0.054(0.907)$	$-0.005(0.867)$	$-0.139(0.933)$
	2007	143	2.57(0.73)	3.53(0.95)	21.3(12.9)	$-0.139(0.414)$	0.010(1.043)	$-0.023(0.720)$	0.663(2.422)	$-0.264(1.155)$
	2008	111	2.59(0.74)	3.55(0.90)	12.5(7.5)	$-0.024(0.728)$	$-0.002(1.024)$	$-0.163(0.984)$	0.688(2.289)	$-0.432(0.997)$
<b>Upwind</b>	Average	447	2.68(0.81)	3.61(0.95)	15.7(11.4)	$-0.042(0.801)$	$-0.065(0.980)$	$-0.015(1.045)$	0.159(1.264)	$-0.113(1.046)$
	2006	204	2.79(0.83)	3.69(0.97)	13.5(10.3)	$-0.013(1.039)$	$-0.090(0.958)$	$-0.040(0.961)$	$-0.056(0.964)$	0.060(0.891)
	2007	138	2.63(0.82)	3.57(0.96)	20.9(13.2)	$-0.048(0.545)$	0.022(1.141)	0.020(1.317)	0.409(1.529)	$-0.161(1.313)$
	2008	105	2.54(0.73)	3.50(0.88)	12.7(7.6)	$-0.087(0.544)$	$-0.135(0.757)$	$-0.015(0.745)$	0.220(1.305)	$-0.374(0.834)$
$\ast$ EYE symptoms: Itchy/dry/sore eyes and blocked nose THROAT symptoms: Sore/dry throat, sore chest & runny nose CHEST symptoms: Chest tightness, difficulty breathing SNEEZE symptoms: Sneezing, cough, watery eyes DRY NOSE symptoms: Dry nose										

**Table 2. Mean (SD) values for lung function and symptom scores by location and year** 

<sup>a</sup> Number of observations for FEV1 provided, as observations varied slightly with health outcomes measured.

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# **Table 3. Contrasts\* in lung function, eNO and symptom scores between the downwind site and the other two sites (upwind and Heavily trafficked) adjusted for change from baseline (2006) to subsequent years**





<sup>a</sup> Contrasts are derived from mixed effects regression model in which location and year, and the interaction between them are fixed effects, and individual subject intercepts are a random effect. Main contrasts of interest are those included in the table and consist of downwind versus other exposures, as the downwind exposure is the primary exposure of interest in this study.

<sup>b</sup> Downwind vs (heavily trafficked and upwind) is the comparison of downwind versus heavily trafficked and upwind exposures combined.

 $\degree$  2007/08 vs 2006 represents the comparison of both post-tunnel years combined (2007/08) versus the pre-tunnel year (2006).

<sup>d</sup> Represents the overall comparison of the heavily trafficked site compared to the upwind and downwind sites combined (data from all years included), as we anticipated the heavily trafficked site to have higher air pollution levels and that health effects would be more prominent.

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Over the three year period as a whole, exposure at the heavily trafficked locations, compared with both the downwind and upwind stack locations, was associated with more airway inflammation (eNO) (ratio=1.09), and higher scores for EYE, CHEST, and THROAT symptoms (last row, Table 3). These findings support the validity of the study design and analysis.

The change in "smell strong" from 2006 to subsequent years was higher for downwind locations than the other two locations (Table 3). However, in general this pattern was not observed for "smell-annoying". Overall, both "smell-strong" and "smell-annoying" were higher at the heavily trafficked locations than the two stack locations (last row, Table 3).

# *Sensitivity analysis for health outcomes*

In the sensitivity analysis we excluded data which might have led to exposure misclassification due to unexpected wind direction, as described earlier. Data for: 5 "downwind" sites in 2006 and 2 in 2008; for 2 "upwind" sites in 2006 and 2 in 2008; and for one heavily trafficked site in 2006; were excluded.

"smell\_strong" from 2006 to subsequent years was higher for down<br>wo locations (Table 3). However, in general this pattern was not obs<br>g". Overall, both "smell-strong" and "smell-annoying" were higher<br>tions than the two sta In this analysis FVC was increased for downwind locations vs the other two locations, for most year comparisons (Table 2, Supplementary Material).  $FEV<sub>1</sub>$  was also increased for downwind vs heavily trafficked locations. Conversely, eNO was higher at downwind compared to upwind locations for 2007 vs 2006. The increased odds of "smell-strong" at downwind locations did not differ to the other locations in the sensitivity analysis. Overall, eNO levels, CHEST symptoms, "smell-strong" and "smell-annoying" were higher at the heavily trafficked locations compared to upwind or downwind locations, confirming the findings of the intention-to-treat analysis (Table 2, Supplementary Material).

#### **Air pollutant levels**

Material). NOx and NO<sub>2</sub> ratios were lower for the downwind sites t<br>007 vs 2006. In contrast, the NO<sub>2</sub> ratio for 2008 vs 2006 was higher is<br>than at the heavily trafficked sites (Table 5). The change in PM (fo<br>2006 to subs As expected, the concentrations of all pollutants were higher at the heavily trafficked sites than the upwind or downwind sites. However, there was little difference in mean pollutant levels between the downwind and upwind locations around the ventilation stack (Table 4 and Table 3, Supplementary Material). NOx and  $NO<sub>2</sub>$  ratios were lower for the downwind sites than the other two sites for 2007 vs 2006. In contrast, the  $NO<sub>2</sub>$  ratio for 2008 vs 2006 was higher at the downwind sites than at the heavily trafficked sites (Table 5). The change in PM (for all size fractions) from 2006 to subsequent years did not vary by site. Overall, the change in pollutants was greatest at the heavily trafficked sites compared to the upwind and downwind sites combined.

In the sensitivity analysis the pollutant ratios for downwind locations compared to upwind locations increased for most pollutants supporting the validity of the exclusion criteria (Table 4, Supplementary Material).

NO2 and PM10 decreased at the western and eastern elevated monitoring sites between 2006 and 2007, after adjusting for changes in regional air quality (Table 6).

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# **Table 4. Mean pollutant levels by location (2006-2008 overall)**



a Means and medians calculated by taking the mean of daily averages at each site. IQR, minimum and maximum values using all (not averaged) data.

b Grab sample for  $NO<sub>2</sub>$  and NOx collected at each site every 15 min and averaged per two hr exposure period.

c PM data collected every two min.

**For Formal Phillips** d One sample collected at site providing an average VOC reading per two hr exposure period. 2007 data not used in analysis due to different detection limits used by different laboratory. 95<sup>th</sup> percentile not reported for VOCs as sample size too small; 95<sup>th</sup> percentile equivalent to maximum reading.
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**Table 5. Ratios of pollutant levels<sup>a</sup> contrasting the downwind site and the other two sites (upwind and heavily trafficked) adjusted for change from baseline (2006) to subsequent years** 



<sup>a</sup> Ratios are derived from a mixed effects regression model in which location and year, and the interaction between them are fixed effects, and individual intercepts are a random effect. Main contrasts of interest are those included in the table and consist of downwind versus other exposures, as the downwind exposure is the primary exposure of interest in this study.

<sup>b</sup> Downwind vs (heavily trafficked and upwind) is the comparison of downwind versus heavily trafficked and upwind exposures combined.

 $\degree$  2007/08 vs 2006 represents the comparison of both post-tunnel years combined (2007/08) versus the pre-tunnel year (2006).

<sup>d</sup> Represents the overall comparison of the heavily trafficked site compared to the two stack sites (downwind and upwind) combined, (data from all years included).



**Table 6. Mean pollutant concentrations for elevated monitors by year and difference in pollutant concentration from post-tunnel year compared to pre-tunnel year** 

a. Represents the change in pollutant concentration between pre-tunnel (25 Mar 06-24 Mar 07) and post-tunnel (25 Mar 2007-24 Mar 2008)

b. Adjusted for change in air quality at the 3 regional sites

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### **DISCUSSION**

We did not detect any consistent evidence of detrimental effects on lung function, airway inflammation or respiratory symptoms attributable to short-term exposure to emissions from the tunnel ventilation stack with the exception of increased reporting of dry nose symptoms. This absence of any adverse health effect might be explained by the fact that we did not measure higher pollutant concentrations at the downwind stack sites after the tunnel opened. The finding of increased pollutant levels and increased symptoms and airway inflammation during periods of exposure at the heavily trafficked site supports previous evidence of the detrimental effects of short-term exposures to TRAP.

For pollutant concentrations at the downwind stack sites after the tuninary of increased pollutant levels and increased symptoms and airway inflused so fexposure at the heavily trafficked site supports previous evident eff This study relied on the proposition that the downwind sites we selected were, in fact, exposed to emissions from the tunnel ventilation stack during peak traffic periods. For a downwind site to be impacted by the emission plume from the stack the plume needed to be mixed at ground level prior to reaching the site and subjected to consistent wind direction during the sampling period. We used weather balloons early on each study morning in order to select the most appropriate downwind site. However, it is possible that we were not always able to accurately predict the ground level location of the plume. Compared to the upwind locations, pollutant concentrations at the downwind locations were similar, or sometimes slightly lower. This may have reflected actual conditions in the field suggesting that there was little measurable impact of the plume on the pollutants measured. However, this finding may have also arisen due to inaccurate prediction of the ground level location of the plume. Nevertheless, tThe sensitivity analysis we conducted was designed to address this potential weakness. The findings of that analysis support the original analysis.

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Our finding that eNO, an indicator of airway inflammation, was increased at the heavily trafficked site is consistent with evidence of increased airway inflammation after exposure to TRAP (or diesel) in previous studies. These studies have varied in their measured end-points but include "real-world" exposure studies investigating neutrophilic airway inflammation[4, 5], associations between eNO and  $PM_{2.5}$ , black carbon and fine particle counts[6] and chamber studies reporting increased neutrophilic airway inflammation,[22-25] increases in pro-inflammatory cytokines IL-6[26] and IL-8,[25] increased exhaled carbon monoxide (CO),[22] and increased airway resistance[26, 27].

The absence of effects on spirometric function, even at the heavily-trafficked site, is consistent with findings of chamber studies of diesel exhaust,[22, 24-27] high particle concentrations, [28] and  $NO<sub>2</sub>[29]$  an in-tunnel study[30] and a commuter study[8]. However, our findings are inconsistent with the London study of diesel exhaust exposures which demonstrated small reductions in  $FEV_1$  after 1-2 hr exposures.[5]

lies reporting increased neutrophilic airway inflammation, [22-25] in<br>atory cytokines IL-6[26] and IL-8, [25] increased exhaled carbon mo<br>d increased airway resistance[26, 27].<br>of effects on spirometric function, even at t Strengths of the study include the randomised cross-over design and blinding of participants by using upwind and downwind stack sites. It is unlikely that the participants were aware of the upwind/downwind status of their location on any given day. The study was further strengthened by including the heavily trafficked sites as a positive control. The finding that pollutant levels were higher, and that subjects had greater airway inflammation, eye and chest symptom scores at this site, support the validity of the study methodology. The before/after nature of our study and repeated measures within and across years also strengthened the study design.

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A limitation of our study is the differential loss from the groups over time. Common reasons for non-participation in follow-up years were travel overseas, residence relocation and changing work practices, all factors unrelated to the experience of study participation.

In conclusion, while we found that this ventilation stack, under current operating conditions, had little measurable short-term health impact, we did demonstrate adverse health effects attributable to short term exposures to TRAP at heavily trafficked locations. These findings support evidence from existing studies of the adverse health effects attributable to short-term exposures to TRAP.

### **Competing interests:**

Nectarios Rose was employed with the NSW Health Department on the Biostatistics Training Scheme at the time the study was conducted.

### **Authors' Contributions:**

Example short-term health impact, we did demonstrate adverse health<br>
o short term exposures to TRAP at heavily trafficked locations. Thes<br>
ence from existing studies of the adverse health effects attributable to<br> **FRAP.**<br> CC organised and participated in all aspects of the study including study design, ethics clearances, recruiting subjects, field work, statistical analysis and led manuscript preparation. WE was involved in all field work. WX contributed to study design and data analysis. WL conducted air quality exposure assessment measurements in the field and provided advice on air quality measurement. NR contributed biostatistical expertise. MR conducted air quality measurements in the field and wind rose analyses. GM participated in the study design, statistical analysis, and provided general oversight of the study. All authors contributed to the preparation of the manuscript.

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#### **Data sharing:**

For data sharing but the presented data are anony<br>on is low. Ethics clearance will need to be sought. Participants did not provide consent for data sharing but the presented data are anonymised and risk of identification is low. Ethics clearance will need to be sought.

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**Figure Legends** 

Figure 1 Schematic of fieldwork

Figure 2 Study area and field sites





Figure 1 Schematic of fieldwork 254x190mm (96 x 96 DPI)



Figure 2 Study area and field sites 254x190mm (96 x 96 DPI)



**Table 1 Characteristics of cohort members participating in each study year** 

<sup>a</sup><br>Measurements made at WIMR lung function laboratory<br> $\frac{b}{c}$ 

Field measurements made at field sites

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**Table 2 Sensitivity analysis**a **for contrasts**b **in lung function, eNO and symptom scores between the downwind site and the other two sites (upwind and heavily trafficked) adjusted for change from baseline (2006) to subsequent years.** 





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**Table 3 Mean pollutant levels by location and year**

<b>Pollutant</b>	Year	Location	<b>Total</b> samples	$Meana$	Median <sup>a</sup>	<b>IQR</b>	<b>Minimum</b>	<b>Maximum</b>	$95^{\text{th}}$ percentile
Nitrogen oxides <sup>b</sup>									
NOx	2006	Heavily trafficked	55	112.0	107.9	49.9-142.9	28.7	426.0	269.0
		Downwind	56	34.6	37.2	16.7-50.5	5.7	89.7	79.4
		Upwind	51	35.4	31.1	12.6-47.3	2.4	100.8	91.0
	2007	Heavily trafficked	51	239.8	128.6	67.1-234.0	14.8	1912.0	784.0
		Downwind	52	30.2	16.2	8.4-44.5	4.6	117.9	96.8
		Upwind	46	37.0	20.7	12.2-57.2	3.5	152.4	93.1
	2008	Heavily trafficked	54	26.7	16.8	9.9-28.6	5.5	115.7	98.0
		Downwind	51	6.8	7.1	$4.3 - 8.5$	2.1	18.9	13.0
		Upwind	53	6.2	5.1	$3.6 - 8.2$	$2.5\,$	16.7	12.5
NO2	2006	Heavily trafficked	55	4.3	4.1	$3.0 - 5.2$	1.3	12.0	$\ \, 8.0$
		Downwind	56	$3.2\,$	$3.2\,$	$2.1 - 4.3$	1.3	5.7	5.1
		Upwind	51	2.4	2.1	$1.7 - 3.0$	0.5	16.0	4.2
	2007	Heavily trafficked	51	13.5	13.6	5.0-17.6	1.4	69.0	31.0
		Downwind	52	4.9	5.1	$1.6 - 7.1$	0.1	18.2	9.5
		Upwind	47	5.9	4.8	$2.9 - 6.8$	0.1	15.7	13.8
	2008	Heavily trafficked	54	0.6	$1.5$	$0.1 - 1.7$	0.1	4.1	3.2
		Downwind	51	$1.1\,$	$1.1\,$	$0.8 - 1.4$	0.1	2.7	2.2
		Upwind	53	$1.1\,$	$1.0\,$	$0.7 - 1.4$	0.1	2.3	2.0
Particulate matter <sup>c</sup> <b>PM10</b>	2006	Heavily trafficked	372	74.2	81.2	46.3-99.1	17.9	168.1	118.7
		Downwind	374	58.1	57.0	33.4-84.0	15.1	195.1	106.7



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<sup>a</sup> Means and medians calculated by taking the mean of daily averages at each site. IQR, minimum and maximum values using all (not averaged) data.

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For beam-period collected revery two minutes.<br>
FM duata collected every two minutes.<br> **FOR USE CONTO GRAD SATE AND SATE OF A CONTO C** <sup>c</sup> PM data collected every two minutes.

d One VOC sample collected at each site to provide average reading for each two hour exposure period. 2007 data not used in analysis due to different detection limits used by different laboratory.

95<sup>th</sup> percentile not reported for VOCs as sample size too small;  $95<sup>th</sup>$  percentile equivalent to maximum reading.





<sup>a</sup> 5 downwind sites and 2 upwind sites in 2006, 2 downwind sites and 2 upwind sites in 2008, and 1 heavily trafficked site in 2006, were excluded for the sensitivity analysis

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