

1 **Microbial regulation of biogeochemical cycles: evidence from a study on methane flux**
2 **and land-use change**

3

4 Loïc Nazaries ^{a,b}, Yao Pan ^{c,d}, Levente Bodrossy ^e, Elizabeth M. Baggs ^f, Peter Millard ^g, J.
5 Colin Murrell ^h, Brajesh K. Singh ^{a,#}

6

7 **Supplementary section**

8

9 **Materials and methods**

10 **Field site description**

11 Bad à Cheo is situated near Thurso, in Northern Scotland, along the A9 (national grid
12 reference ND169503; 58°25'47.35"N, 3°25'48.27"W) and is part of the Rumster forest. It had
13 been the subject of detailed hydrochemical studies (1). The study site is composed of an open
14 bog of deep blanket peat dominated by a mixture of peat moss (*Sphagnum* spp.), deergrass
15 (*Trichophorum cespitosum*) and cotton-grass (*Eriophorum* spp.). Adjacent to the bog,
16 experimental forestry plots were drained, ploughed and planted in 1968 and 1988 with a
17 mixture of Sitka spruce (*Picea sitchensis*) and lodgepole pine (*Pinus contorta*) (2). Therefore,
18 at the time of this study, the young and old pine forests were about 20- and 40-years old,
19 respectively. Because of the age difference, the younger pine forest lay on much wetter
20 ground compared to the older pine forest. Also, at the time of sampling, natural colonisation
21 of the bog by conifers was observed.

22 The Glensaugh Research Station is part of the James Hutton Institute (formerly known as the
23 Macaulay Land Use Research Institute) and is located in Laurencekirk, Aberdeenshire,
24 Scotland, UK (national grid reference NO671782; 56°53'55.91"N, 2°33'0.28"W). The study
25 site is an agroforestry plot that was used to study experimental planting and grazing
26 management between 1988 and 2001 (Glensaugh Agroforestry Demonstration,

27 <http://www.macaulay.ac.uk/aboutus/researchstations/agroforestry.html>). The site is a pasture
28 occupied by ewes and lambs with regular fertiliser applications during each grazing season.
29 In 1988, a subplot was planted with Scots pine (*Pinus sylvestris*). Therefore, the pine forest
30 was about 20-years old at the time of this study.

31 The Craggan forest is also described elsewhere (3, 4). It is located in Moray, near the Spey
32 River, Scotland, UK (national grid reference NJ190322; 57°22'20.50"N, 3°20'45.42"W) and
33 was originally used in 1978 in a study testing the durability of changes caused by *Betula* spp.
34 on moorland (3). Open *Calluna*-dominated moorland is adjacent to a natural chronosequence
35 of birch trees (*Betula* spp.) aged about 53, 62 and 88 years. During the colonisation phase,
36 heather (*Calluna vulgaris*) was replaced by wavy hairgrass (*Deschampsia flexuosa*) and
37 bilberry (*Vaccinium myrtillus*) before long-term establishment of birch woodland (5, 6). For
38 our study, soil samples were taken from the 62-year-old stand (young birch forest) and the
39 88-year-old stands (old birch forest) only. It is worth noting that the moorland was cleared of
40 trees in 1974 but has since been progressively naturally colonised by birch. Also, due to the
41 old age of the 88-year-old stands, few trees were left standing and alive, with mainly colonial
42 bentgrass (*Agrostis capillaris*) present as understorey vegetation. The Craggan site was
43 situated on the slope of a hill.

44 The Tulchan Estate is described in more detail elsewhere (3, 5). Briefly, the study site is
45 located on the Tulchan Estate, Speyside, Scotland, UK (national grid reference NJ154373;
46 57°24'42.78"N, 3°26'28.65"W). The site contains a natural heather moorland-birch woodland
47 chronosequence. The open *Calluna*-dominated moorland is adjacent to two stands of birch
48 trees (*Betula pubescens*) following natural invasion of the heathland in *ca.* 1953 (young birch
49 forests, 55-year-old) and *ca.* 1943 (old birch forests, 65-year-old) (7). Like in Tulchan,
50 similar changes in vegetation occurred (5).

51

52

53 **Table S1. Sites and land uses for this study.**

54 For each land use, there was n=4 replicates per seasonal sampling. N/A means non-
55 applicable.

| Site | Land use | Age of the forest | National grid reference (GPS coordinates) |
|-------------|-----------------|--------------------------|--|
| Bad à Cheo | Bog | N/A | ND169503 |
| | Young Pine | 20 years | (58°25'47.35"N, 3°25'48.27"W) |
| | Old Pine | 40 years | |
| Glensaugh | Grassland | N/A | NO671782 |
| | Young Pine | 20 years | (56°53'55.91"N, 2°33'0.28"W) |
| Craggan | Moorland | N/A | NJ190322 |
| | Young Birch | 62 years | (57°22'20.50"N, 3°20'45.42"W) |
| | Old Birch | 88 years | |
| Tulchan | Moorland | N/A | NJ154373 |
| | Young Birch | 55 years | (57°24'42.78"N, 3°26'28.65"W) |
| | Old Birch | 65 years | |

56

57

58 **Soil sampling**

59 The sampling procedure was the same for each site and similar to a method used in a
60 previous study (8). In brief, stainless steel rings (10 cm diameter, 0-10 cm depth) were used
61 to extract soil cores after removal of the L and FH layers. For each site, twelve replicates per
62 habitat were sampled at random, and were randomly grouped in four sets of three cores for
63 measurement of net CH₄ fluxes (see below). Therefore, n=4 for each habitat, for each site, for
64 each season (total n=176). Within a few hours of sampling, the soil cores were taken to the
65 laboratory and left overnight in an environment-controlled chamber (minimum 70%
66 humidity). For each seasonal experiment, the temperature of the chamber was set using a
67 value close to the air temperature of the site at the time of sampling: 5°C in winter, 10°C in
68 spring, 15°C in summer and 20°C in autumn. The following day, measurements of net CH₄
69 fluxes were performed and the soil cores were then stored at 4°C.

70 Also, during summer, smaller intact cores (5 cm diameter, 0-5 cm depth) were taken in
71 triplicate from each habitat from each site for bulk density, porosity and water retention
72 analysis (total n=33).

73

74 **Soil analyses**

75 Field-moist 5.6-mm sieved soils were extracted with 1 M KCl for one hour and extracts were
76 analysed colorimetrically for mineral N (NH₄⁺-N and NO₃⁻-N). Moisture content was
77 measured after drying the fresh soil samples in an oven at 105°C overnight. Fresh soils were
78 dried at 30°C and then dry-sieved through a 2 mm-mesh sieve for pH and particle size
79 analysis. pH was measured in water after mixing thoroughly the soil water slurry (1:2.5
80 suspension) for 30 minutes. Particle size distribution analysis was performed using laser
81 diffraction on a Malvern Mastersizer 2000 particle size analyser fitted with a Malvern Hydro
82 2000G sample dispersion tank (Malvern, UK). A sub-sample of the dry-sieved soils was

83 milled (Retsch mill, 5 minutes at 60 strokes per second) for subsequent use for total C and N
84 analysis by combustion in a Thermo-Finnigan Elemental Analyser (FlashEA 1112 Series).
85 Small soil cores (5 cm diameter, 0-5 cm depth) were dried using hanging water columns for
86 measurement of soil bulk density, porosity and WFPS. The field-moist cores were saturated
87 with water and gravimetric soil water content was estimated after equilibrating the cores at
88 10, 50, 100 and 150 kPa suction pressures. WFPS was estimated as the ratio of the volumetric
89 soil moisture content to the total pore space, or porosity. Porosity was estimated to be
90 equivalent to the volumetric water content at water saturation. Volumetric water content was
91 calculated as the product of the gravimetric water content and the bulk density. Bulk density
92 corresponded to the oven-dry soil weight (105°C overnight) divided by the core volume.
93 Table S2 provides some of the physico-chemical properties of the soils under the different
94 habitats.

95

96

97 **Table S2. Chemical and physical soil properties.**

98 The data are means \pm s.e.m. (n=8 replicates – autumn and summer combined, except for bulk density, porosity and WFPS (n=4 – summer only))
 99 of each habitat. For each soil characteristic, Greek letters (α , β , γ) indicate statistical differences between habitats within each site, according to
 100 multiple pairwise comparison ($P < 0.05$).

| Site | Habitat | pH | Total C (g.kg ⁻¹) | Total N (g.kg ⁻¹) | C:N ratio | NH ₄ ⁺ -N (mg.kg ⁻¹) | NO ₃ ⁻ -N (mg.kg ⁻¹) | Moisture (%) | Particle size (% of total) | | | Bulk density (g.cm ⁻³) | Porosity (%) | WFPS at field capacity (at 50 kPa) (%) |
|------------|----------------|--|---|---|---|---|---|---|--|---|---|--|---|--|
| | | | | | | | | | 0.02-2.00 μ m | 2-20 μ m | 20- 2000 μ m | | | |
| Bad à Cheo | Bog | 3.6 \pm 0.04 ^{α} | 94 \pm 3 | 3.2 \pm 0.1 ^{α} | 30 \pm 1 | 154 \pm 15 ^{α} | 167 \pm 16 | 87 \pm 1 ^{α} | 1.5 \pm 0.32 ^{α} | 21 \pm 3 ^{α} | 77 \pm 3 ^{α} | 0.39 \pm 0.07 | 85 \pm 3 | 76 \pm 6 |
| | Young Pine | 3.5 \pm 0.04 ^{β} | 90 \pm 2 | 2.9 \pm 0.2 ^{$\alpha\beta$} | 31 \pm 1 | 89 \pm 7 ^{β} | 142 \pm 12 | 88 \pm 0 ^{α} | 0.38 \pm 0.14 ^{β} | 10 \pm 2 ^{β} | 90 \pm 2 ^{β} | 0.25 \pm 0.04 | 91 \pm 2 | 71 \pm 8 |
| | Old Pine | 3.2 \pm 0.07 ^{γ} | 92 \pm 1 | 2.7 \pm 0.07 ^{β} | 34 \pm 1 | 81 \pm 7 ^{β} | 132 \pm 16 | 79 \pm 1 ^{β} | 0.40 \pm 0.17 ^{β} | 10 \pm 2 ^{β} | 90 \pm 2 ^{β} | 0.34 \pm 0.03 | 87 \pm 1 | 66 \pm 6 |
| Glensaugh | Grassland | 4.2 \pm 0.05 ^{α} | 5.9 \pm 0.93 ^{α} | 0.58 \pm 0.08 ^{α} | 10 \pm 0 ^{α} | 30 \pm 4 ^{α} | 131 \pm 20 | 32 \pm 1 | 3.8 \pm 0.38 | 34 \pm 1 ^{α} | 63 \pm 2 ^{α} | 1.39 \pm 0.06 | 48 \pm 2 | 72 \pm 6 |
| | Young Pine | 3.9 \pm 0.04 ^{β} | 3.9 \pm 0.26 ^{β} | 0.36 \pm 0.03 ^{β} | 11 \pm 0 ^{β} | 17 \pm 1 ^{β} | 148 \pm 19 | 30 \pm 1 | 4.6 \pm 0.31 | 39 \pm 1 ^{β} | 56 \pm 2 ^{β} | 1.57 \pm 0.12 | 41 \pm 5 | 59 \pm 3 |
| Craggan | Moorland | 3.4 \pm 0.03 | 76 \pm 5 ^{α} | 2.9 \pm 0.2 ^{α} | 27 \pm 1 ^{α} | 83 \pm 10 ^{α} | 71 \pm 10 ^{α} | 79 \pm 1 ^{α} | 1.8 \pm 0.27 | 20 \pm 3 | 78 \pm 4 | 0.36 \pm 0.07 | 87 \pm 3 | 60 \pm 3 ^{α} |
| | Young Birch | 3.4 \pm 0.04 | 45 \pm 7 ^{β} | 2.0 \pm 0.3 ^{β} | 22 \pm 1 ^{β} | 105 \pm 12 ^{$\alpha\beta$} | 35 \pm 4 ^{β} | 66 \pm 2 ^{β} | 3.5 \pm 0.42 | 36 \pm 3 | 60 \pm 3 | 0.38 \pm 0.01 | 86 \pm 1 | 69 \pm 2 ^{$\alpha\beta$} |
| | Old Birch | 3.4 \pm 0.01 | 63 \pm 5 ^{α} | 3.0 \pm 0.3 ^{α} | 22 \pm 1 ^{β} | 116 \pm 13 ^{β} | 37 \pm 6 ^{β} | 70 \pm 1 ^{β} | 2.9 \pm 0.69 | 33 \pm 8 | 64 \pm 8 | 0.49 \pm 0.05 | 82 \pm 2 | 72 \pm 1 ^{β} |
| Tulchan | Moorland | 3.5 \pm 0.03 ^{α} | 79 \pm 4 ^{α} | 3.2 \pm 0.2 ^{α} | 25 \pm 1 ^{α} | 84 \pm 13 ^{α} | 131 \pm 19 ^{α} | 85 \pm 1 ^{α} | 0.55 \pm 0.17 ^{α} | 11 \pm 2 ^{α} | 88 \pm 2 ^{α} | 0.26 \pm 0.02 ^{α} | 90 \pm 1 ^{α} | 64 \pm 3 |
| | Young Birch | 3.7 \pm 0.04 ^{β} | 6.6 \pm 0.82 ^{β} | 0.36 \pm 0.05 ^{β} | 19 \pm 1 ^{β} | 30 \pm 4 ^{β} | 23 \pm 3 ^{β} | 37 \pm 2 ^{β} | 3.9 \pm 0.05 ^{β} | 28 \pm 3 ^{β} | 68 \pm 3 ^{β} | 0.99 \pm 0.08 ^{β} | 63 \pm 3 ^{β} | 62 \pm 5 |
| | Old Birch | 3.6 \pm 0.03 ^{$\alpha\beta$} | 11 \pm 4 ^{β} | 0.51 \pm 0.15 ^{β} | 19 \pm 2 ^{β} | 38 \pm 5 ^{β} | 33 \pm 7 ^{β} | 42 \pm 3 ^{β} | 3.1 \pm 0.45 ^{β} | 24 \pm 2 ^{β} | 73 \pm 2 ^{β} | 1.48 \pm 0.22 ^{β} | 44 \pm 8 ^{β} | 62 \pm 9 |

102 **Gas fluxes measurements**

103 Headspace gas samples were taken under dim light using closed PVC chambers (~9 L) fitted
104 with a gas sampling tube and a 3-way tap. Out of the twelve replicates from each habitat,
105 three soil cores per chamber were used, so for each habitat n=4. Before starting any
106 measurements, the soil cores were unwrapped and left in the open chamber for 2-3 hours.
107 Immediately after locking the lid of the chambers (T₀), air (12 mL) was sampled from the
108 chamber's headspace using a plastic syringe fitted with Luer lock (Fisher Scientific, UK) and
109 3-way tap, and quickly injected into a pre-evacuated 12-mL glass Exetainer (Labco Ltd, UK)
110 using a Luer syringe needle 24 mm, 25G (Fisher Scientific, UK). Headspace sampling was
111 repeated after 30, 60 and 90 minutes (T₃₀, T₆₀ and T₉₀, respectively).
112 Using a gas-tight precision injection glass syringe, an air sample (1 mL) was taken from the
113 Exetainer and injected into the column of the gas chromatograph. A 20-ppm CH₄ standard
114 (CryoService Limited, UK) was run every 20 samples to check for accuracy. Precision was
115 3.31% with a method-detection limit of 0.19 ppm. The atmospheric CH₄ concentrations (at
116 T₀, T₃₀, T₆₀ and T₉₀) of the unknown samples were calculated by comparing the peak area
117 from the chromatogram to the peak area of the CH₄ standard. The results were then used to
118 estimate the CH₄ flux inside the chamber headspace as follows (9, 10):

$$F = \rho \frac{V}{A} \times \frac{\Delta c}{\Delta t}$$

119 where F is the CH₄-C flux (μg.m⁻².h⁻¹); V the volume of the chamber (m³); A the base area of
120 the chamber (m²); Δc/Δt the average rate of change of CH₄ concentration (ppmv) with time
121 (h); ρ the density of CH₄-C (kg.m⁻³) at the corresponding experimental temperature. The gas
122 density ρ was calculated as follows:

$$\rho = \frac{P \times M}{R \times T}$$

123 where P is the air pressure (1 atm); M the molecular weight of CH₄-C (g.mol⁻¹); R the
124 universal gas constant (0.082057 atm.L.mol⁻¹.K⁻¹); T the experimental air temperature (K).

125

126 **Analysis of methanotrophic community by molecular ecology approaches**

127 PCR conditions

128 The amplification of the *pmoA* genes used the following optimised master mix (final
129 concentrations given): 1x NH_4^+ reaction buffer, 6 mM MgCl_2 , 50 μM of each
130 deoxynucleotide, 0.02 $\text{U}\cdot\mu\text{L}^{-1}$ BioTaq™ DNA polymerase (all reagents from Bioline, UK),
131 0.3 $\mu\text{g}\cdot\mu\text{L}^{-1}$ bovine serum albumin (Roche diagnostic, UK), 0.3 μM of each primer and 3
132 $\text{ng}\cdot\mu\text{L}^{-1}$ of DNA template.

133 An optimised touchdown PCR program was used: initial denaturation at 95°C for 7 min,
134 denaturation at 94°C for 1 min, annealing at 65°C for 1.5 min, extension at 72°C for 1 min for
135 15 cycles with a decrement of 0.8°C/cycle of the annealing temperature, and then
136 denaturation at 94°C for 1 min, annealing at 53°C for 1 min, extension at 72°C for 1 min for
137 20 cycles, and a final extension at 72°C for 10 min. PCRs were performed on a DYAD™
138 DNA Engine® Peltier thermal cycler (MJ Research, USA). Purity and size of the PCR
139 amplicons were checked by loading 5 μL of each reaction mix on a 1% (w/v) agarose gel
140 stained with ethidium bromide, and observed under UV light. PCR products were purified
141 using the UltraClean-htp™ 96-well PCR Clean-up™ kit (MoBio, USA) according to the
142 manufacturer's instructions, except that DNA was eluted in 35 μL instead of the
143 recommended 100 μL , in order to increase the final concentration. Concentrations of the
144 purified PCR products were then measured on the Nano-Drop® ND-1000.

145 Terminal-restriction fragment length polymorphism (T-RFLP) analysis

146 In a 10- μL reaction mix, the final concentrations of the different components were the
147 following: 10 $\text{ng}\cdot\mu\text{L}^{-1}$ of DNA template, 1x of enzyme solution, 1x of enzyme buffer and 0.1
148 $\mu\text{g}\cdot\mu\text{L}^{-1}$ of bovine serum albumin (all reagents from Promega, UK). Samples were then
149 digested for 3 hours at 37°C on a DYAD™ thermal cycler, and the enzymatic reaction was
150 stopped by an incubation at 95°C for 15 min. Aliquots of digested PCR products (1 μL) were

151 transferred onto a MicroAmp® optical 96-well plate (Applied Biosystems, UK) and mixed
152 with 12 µL of Hi-Di™ formamide. 0.3 µl of LIZ-labelled GeneScan™-500 internal size
153 standard (all reagents from Applied Biosystems, UK) was added and the reaction was
154 denatured at 95°C for 5 min.

155 Terminal-restriction fragments (T-RFs) generated by the sequencer were analysed using the
156 size-calling software GeneMapper™ 4.0 (Applied Biosystems, UK) and quantified by
157 advanced mode using second order algorithm. T-RFs in a T-RFLP profile were selected by
158 the software if their minimum peak height was above the noise observed with the negative
159 control (usually above 25 relative fluorescence units).

160 Diagnostic *pmoA* microarray analysis

161 *In vitro* transcription

162 *In vitro* transcription was carried out under RNAase-free conditions. The procedure was as
163 follows (20 µl final volume): 8 µL purified PCR product (50 ng.µL⁻¹), 4 µL 5x T7 RNA
164 polymerase buffer, 2 µL DTT (100 mM), 0.5 µL RNAsin (40 U.µL⁻¹) (Promega), 1 µL of
165 each ATP, CTP, GTP (10 mM), 0.5 µL UTP (10 mM), 1 µL T7 RNA polymerase (40 U.µL⁻¹)
166 (Invitrogen) and 1 µL Cy3-UTP (5 mM) were added into a 1.5 mL Eppendorf tube and
167 incubated at 37°C for 4 hours. RNA was purified immediately based on the RNeasy Mini Kit
168 (Qiagen): 80 µL of DEPC-treated water were added to IVT mixture, followed by adding 350
169 µL of RLT and 250 µL of ethanol, and then mixed thoroughly. Samples were transferred to
170 an Rneasy mini tube and 500 µL of RPE were added. Tubes were centrifuged at 10,000 rpm
171 for 15 sec. Another 500 µL of RPE were added, and then centrifugation at 10,000 rpm for 2
172 min. Purified RNA was eluted into 50 µL of dH₂O. RNA yields and dye incorporation rates
173 were measured by spectrophotometry. Purified RNA was fragmented by incubating with 9.5
174 mM ZnCl₂ and 24 mM TrisCl (pH7.4) at 60°C for 30 min. Fragmentation was stopped by the

175 addition of 12 mM EDTA (pH 8.0) to the reaction and putting it on ice. 1 μL of RNAsin (40
176 $\text{U}\cdot\mu\text{L}^{-1}$) was added to the fragmented target.

177 Hybridisation

178 Hybridisation was carried out (in triplicate) in an aluminium block on a Belly Dancer (Stovall
179 Life Sciences, USA), which was preheated to 55°C for at least 1 hour. For each hybridisation,
180 the following was added to a 1.5 mL Eppendorf tube (100 μl final volume) and incubated at
181 65°C for 1 min: 62 μL of DEPC-treated water, 1 μL of 10% SDS, 30 μl of 20x SSC (3 M
182 sodium chloride, 0.3 M sodium citrate, pH 7.0), 2 μl of 50x Denhardt's reagent (Sigma) and 5
183 μl of target RNA (corresponding to about 200 ng of RNA). Preheated hybridisation mixtures
184 were applied onto the preheated slides containing the arrays. The assembled microarray slides
185 were incubated overnight in the HybriWell hybridisation chambers (Grace BioLabs) at 55°C
186 at maximum bending and lowest rotation. Following hybridisation, the slides were washed by
187 shaking at room temperature for 5 min in 2x SSC, 0.1% (w/v) SDS; twice for 5 min in 0.2x
188 SSC and finally for 5 min in 0.1x SSC. Slides were dried using an airgun.

189 Scanning and data analysis

190 Hybridised slides were scanned at 10 μm resolution with a GenePix 4000 laser scanner
191 (Axon, USA) at a wavelength of 532 nm. Fluorescent images were analyzed with the
192 GenePix software (Axon, USA). Microsoft Excel was used for statistical analysis and
193 presentation of results.

194 Results were normalised to a positive control. The hybridisation signal for each probe was
195 expressed as a percentage of the signal (median of signal minus background) of the positive
196 control probe mtrof173 on the same array (11). As each slide contained triplicate arrays,
197 normalised signal intensities of the triplicate spots on a slide were used to determine average
198 results and standard deviations. Hybridisation between a probe and a target was considered
199 positive if the signal was at least 5% of the strongest signal obtained for that probe with the

200 validation set of reference strains/clones. For probes where no perfect match reference target
201 was available or the strongest signal was less than 60 (% of the signal obtained for mtrof173),
202 this reference value was arbitrarily set to 60. This was found to minimize false positive calls
203 while not creating any false negative calls (12).

204

205 **Identifying active methanotrophs by stable isotope probing of phospholipids fatty acids** 206 **(PLFA-SIP)**

207 Microcosm experiments and PLFA-SIP

208 Field-moist 5.6-mm sieved soils (10 g) were transferred into 125-mL Wheaton glass serum
209 bottles (Sigma-Aldrich, UK), and left overnight in the dark at 20°C. The following day,
210 bottles were sealed and injected through the rubber septum with 2.5 mL of $^{13}\text{C-CH}_4$ (>99
211 atom%, CK Gas, UK) from a ~5,000 ppm master mix in order to have a starting headspace
212 concentration of ~100 ppm. Soils were incubated in the dark at 20°C.

213 PLFA-SIP was performed on the autumn and summer soils only, and on all chamber replicate
214 soils (n=4) from each habitat from each site (total n=88). The autumn samples were all
215 incubated for 14 days whereas the summer samples were incubated until >90% of $^{13}\text{C-CH}_4$
216 had been incorporated (between 4 and 32 days depending on the activity of the soils).

217 After incubation was complete, ^{13}C -enriched soils were kept frozen at -20°C.

218 Compound-specific isotope analysis

219 The isotopic composition of individual PLFAs was determined using a GC Trace Ultra with
220 combustion column attached via a GC Combustion III to a Delta V Advantage isotope ratio
221 mass spectrometer (all Thermo Finnigan, Germany). Samples (2 μL) were injected in
222 splitless mode onto a J&W Scientific HP-5 column, 50 m length, id 0.2 mm with a film
223 thickness of 0.33 μm (Agilent Technologies Inc, USA). All other running conditions were as
224 described elsewhere (13). The carbon isotope ratios were calculated with respect to Vienna-

225 PDB ($\delta^{13}\text{C}_{\text{V-PDB}}$) through the use of a CO_2 reference gas injected with every sample and
226 traceable to International Atomic Energy Agency reference material NBS 19 TS-Limestone.
227 Repeated analysis, over a two-month period, of the $\delta^{13}\text{C}$ value of a C19:0 FAME internal
228 standard gave a standard deviation of 1.11‰ (n=18).

229

230

231 **Results**

232

233 **Table S3. Linear regression of the net CH₄ fluxes from each site with some abiotic**
 234 **properties of the soil.**

235 The data are P values from simple linear regression analyses. Values in bold represent
 236 significance ($\alpha=0.05$).

| | | Site | Bad à Cheo | Glensaugh | Craggan | Tulchan |
|------------------|---------------------------------|------|--------------|--------------|------------------|------------------|
| Abiotic property | | | | | | |
| | pH | | 0.443 | 0.041 | 0.709 | 0.054 |
| | Total C | | 0.625 | 0.293 | 0.014 | <0.001 |
| | Total N | | 0.211 | 0.159 | 0.302 | <0.001 |
| | C:N ratio | | 0.244 | 0.012 | 0.010 | 0.012 |
| | NH ₄ ⁺ -N | | 0.013 | 0.047 | 0.663 | 0.005 |
| | NO ₃ ⁻ -N | | 0.537 | 0.959 | 0.095 | <0.001 |
| | Moisture | | 0.202 | 0.211 | <0.001 | <0.001 |
| Particle size | 0.02-2.00 µm | | 0.043 | 0.655 | 0.343 | <0.001 |
| | 2-20 µm | | 0.018 | 0.169 | 0.254 | 0.001 |
| | 20-2000 µm | | 0.019 | 0.216 | 0.258 | <0.001 |
| | Bulk density | | 0.966 | 0.467 | 0.282 | 0.062 |
| | Porosity | | 0.966 | 0.218 | 0.282 | 0.062 |
| | WFPS | | 0.720 | 0.006 | 0.004 | 0.370 |

237

238

239 It should be noted that the overall fluorescence detected in the T-RFLP profiles for the
240 samples from the old pine forest at Bad à Cheo was very low due to the quality of the soil and
241 the difficulty to extract DNA. Thus, T-RFLP results from the old pine forest at Bad à Cheo
242 are not included (see Table S4 and Table S5).

243

244

245

246 **Table S4. Relative (seasonal and annual) abundance (\pm s.e.m.) and phylogenetic affiliation of the most abundant T-RFs (digestion of**
 247 ***pmoA* with the restriction enzyme *HhaI*) found in soils (n=16 replicates for each habitat).**

248 For each T-RF, statistical differences between seasons within each habitat are indicated by different Roman letters (a, b), while Greek letters (α ,
 249 β) indicate statistical differences between habitats within each site, according to multiple pairwise comparison ($\alpha=0.05$).

| Site | Habitat | Season | T-RF ID (<i>enzyme-bp</i>) | | | | | | T-RF relative total (%) |
|------------|------------|--------|------------------------------|------------|----------------|---------------|---------------|----------------|-------------------------------|
| | | | <i>Hha-32</i> | | <i>Hha-129</i> | | <i>Hha-81</i> | | |
| | | | Seasonal | Annual | Seasonal | Annual | Seasonal | Annual | |
| Bad à Cheo | Bog | Autumn | 17 \pm 6 | | 30 \pm 3 | | | 30 \pm 6 | 77 |
| | | Spring | 17 \pm 3 | 17 \pm 2 | 25 \pm 2 | 29 \pm 2 | 33 \pm 5 | 31 \pm 3 | |
| | | Summer | 19 \pm 2 | | 28 \pm 7 | | 26 \pm 6 | | |
| | | Winter | 17 \pm 3 | | 33 \pm 6 | | 34 \pm 7 | | |
| | Young Pine | Autumn | 12 \pm 4 | | 34 \pm 6 | | | 31 \pm 8 | 88 |
| | | Spring | 15 \pm 7 | 21 \pm 4 | 29 \pm 7 | 29 \pm 3 | 48 \pm 12 | 38 \pm 4 | |
| | | Summer | 23 \pm 8 | | 36 \pm 3 | | 38 \pm 5 | | |
| | | Winter | 34 \pm 7 | | 19 \pm 5 | | 35 \pm 5 | | |
| Glensaugh | Grassland | Autumn | 58 \pm 21 | | 15 \pm 8 | | | 12 \pm 12 | 74 |
| | | Spring | 51 \pm 6 | 61 \pm 7 | 8.7 \pm 3.3 | 9.0 \pm 2.3 | 1.8 \pm 1.8 | 3.8 \pm 2.7 | |
| | | Summer | 74 \pm 21 | | 3.4 \pm 3.4 | | 0 | | |
| | | Winter | 62 \pm 10 | | 8.7 \pm 3.8 | | 2.2 \pm 2.2 | | |
| | Young Pine | Autumn | 75 \pm 8 | | 9.4 \pm 1.3 | | | 2.7 \pm 2.7 | 83 |
| | | Spring | 65 \pm 25 | 71 \pm 6 | 8.2 \pm 4.0 | 11 \pm 1 | 0 | 0.8 \pm 0.78 | |
| | | Summer | 69 \pm 4 | | 10 \pm 2 | | 0 | | |
| | | Winter | 74 \pm 8 | | 16 \pm 4 | | 0 | | |

| | | | | | | | | | |
|----------------------------|-------------|---------------------|---|--|--|-----------------------|----------------------|--------------------|----|
| Craggan | Moorland | Autumn | 14±3 | | 19±4 | | 25±3 | 36±6 ^{αβ} | 64 |
| | | Spring | 29±0 | 18±2 ^α | 0 | 10±3 ^α | 35±35 | | |
| | | Summer | 19±2 | | 3.7±2.2 | | 59±7 | | |
| | | Winter | 15±2 | | 14±4 | | 25±8 | | |
| | Young Birch | Autumn | 19±7 | | 4.6±2.7 | | 67±2 | 46±7 ^α | 94 |
| | | Spring | 42±8 | 45±8 ^β | 3.0±2.9 | 2.5±1.2 ^β | 43±8 | | |
| | | Summer | 46±22 | | 0 | | 49±19 | | |
| | | Winter | 81±15 | | 2.7±2.7 | | 16±16 | | |
| | Old Birch | Autumn | 64±19 | | 4.2±4.2 | | 13±8 | 17±6 ^β | 81 |
| | | Spring | 37±18 | 58±9 ^β | 8.4±8.4 | 7.8±2.7 ^{αβ} | 25±16 | | |
| | | Summer | 65±5 | | 13±6 | | 12±10 | | |
| | | Winter | 68±22 | | 6.4±3.7 | | 17±11 | | |
| Tulchan | Moorland | Autumn | 10±5 | | 13±2 | | 58±8 | 55±6 ^α | 82 |
| | | Spring | 21±12 | 13±3 ^α | 9.8±3.4 | 14±2 ^α | 53±18 | | |
| | | Summer | 6.5±1.8 | | 14±6 | | 73±6 | | |
| | | Winter | 14±2 | | 18±3 | | 38±6 | | |
| | Young Birch | Autumn | 70±21 | | 2.8±1.6 | | 27±19 | 13±7 ^β | 98 |
| | | Spring | 70±20 | 82±8 ^β | 4.5±3.0 | 2.7±1.0 ^β | 22±15 | | |
| | | Summer | 98±1 | | 1.9±1.0 | | 0 | | |
| | | Winter | 95±3 | | 1.4±1.4 | | 0.94±0.94 | | |
| | Old Birch | Autumn | 58±13 | | 12±6 | | 7.7±7.6 ^a | 13±8 ^β | 84 |
| | | Spring [†] | 0 | 61±9 ^β | 0 | 10±3 ^α | 90 ^b | | |
| | | Summer | 68±19 | | 12±6 | | 3.1±3.0 ^a | | |
| | | Winter | 73±11 | | 10±5 | | 5.3±4.3 ^a | | |
| Associated organism | | | Distant relative of <i>Methylocapsa</i> sp./ USC _α | Distant relative of <i>Methylocapsa</i> sp./ Cluster 5 | Distant relative of <i>Methylocystaceae</i> | | | | |
| Reference | | | | Nazaries <i>et al.</i> (8) Singh <i>et al.</i> (14) | | | | | |

250

251 **Table S5. Effects of land-use change (or tree growth) and seasonal changes on the**
 252 **methanotrophic community (digestion of *pmoA* with the restriction enzyme *HhaI*).**

253 The data are *P* values corresponding to the first four IPC scores of the AMMI analyses, and
 254 were obtained by nested ANOVA and MANOVA. Within each column, statistical differences
 255 between seasons within each habitat are indicated by different Roman letters (a, b), while
 256 Greek letters (α , β) indicate statistical differences between habitats, according to multiple
 257 pairwise comparison ($\alpha=0.05$).

| | | IPC 1 | IPC 2 | IPC 3 | IPC 4 | MANOVA |
|-------------------|----------------|----------------------------|----------------------------|----------------------------|----------------------------|----------------------------|
| Bad à Cheo | % variation | 42.0 | 26.5 | 14.2 | 9.3 | |
| | Habitat | 0.008 | 0.374 | 0.088 | 0.007 | <0.001 |
| | Habitat/Season | 0.755 | 0.295 | 0.239 | 0.386 | 0.390 |
| Bog | Autumn | | | | | |
| | Spring | a | a | a | a | |
| | Summer | α | α | α | α | |
| | Winter | | | | | |
| Young Pine | Autumn | | | | | |
| | Spring | a | a | a | a | |
| | Summer | β | α | α | β | |
| | Winter | | | | | |
| | | IPC 1 | IPC 2 | IPC 3 | IPC 4 | MANOVA |
| Glensaugh | % variation | 55.8 | 23.2 | 8.0 | 5.1 | |
| | Habitat | 0.840 | 0.044 | 0.003 | 0.027 | <0.001 |
| | Habitat/Season | 0.476 | 0.011 | 0.208 | 0.507 | 0.014 |
| Grassland | Autumn | | a | | | |
| | Spring | a | a | a | a | |
| | Summer | α | b | α | α | |
| | Winter | | a | | | |
| Young Pine | Autumn | | | | | |
| | Spring | a | a | a | a | |
| | Summer | α | β | β | β | |
| | Winter | | | | | |
| | | IPC 1 | IPC 2 | IPC 3 | IPC 4 | MANOVA |
| Craggan | % variation | 47.9 | 24.5 | 10.8 | 7.2 | |
| | Habitat | <0.001 | 0.001 | 0.063 | 0.089 | <0.001 |
| | Habitat/Season | 0.116 | 0.334 | 0.852 | 0.007 | 0.019 |
| Moorland | Autumn | | | | a | |
| | Spring | a | a | a | b | |
| | Summer | α | α | α | ab | α |
| | Winter | | | | a | |

| | | | | | | |
|----------------|--------------------------------------|------------------|--------------|--------------|--------------|------------------|
| Young Birch | Autumn Spring Summer Winter | a α | a β | a α | a α | |
| Old Birch | Autumn Spring Summer Winter | a β | a β | a α | a α | |
| | | IPC 1 | IPC 2 | IPC 3 | IPC 4 | MANOVA |
| | % variation | 75.7 | 11.4 | 6.8 | 2.2 | |
| Tulchan | Habitat | <0.001 | 0.060 | 0.037 | 0.026 | <0.001 |
| | Habitat/Season | 0.053 | 0.341 | 0.411 | 0.389 | 0.233 |
| Moorland | Autumn Spring Summer Winter | a α | a α | a αβ | a αβ | |
| Young Birch | Autumn Spring Summer Winter | a β | a α | a α | a α | |
| Old Birch | Autumn Spring Summer Winter | a β | a α | a β | a β | |

258

259

260 **Table S6. Effects of tree growth on the methanotrophic community (*pmoA* microarray).**

261 The data presented are some of the *pmoA* probes that showed higher levels of hybridisation, and their statistical difference (Greek letters [α , β ,
 262 γ]) between habitat within each site, according to multiple pairwise comparison ($P < 0.05$). ND means that the probe showed no detectable
 263 hybridisation.

| Site | Habitat | Probe | | | | | | | | | | |
|------------|-------------|----------|----------|----------|----------|----------|---------------|---------------|---------------|---------------|----------|----------|
| | | McyM309 | Mcy522 | Mcy459 | Mcy413 | Msi233 | Peat264 | RA14-594 | RA14-591 | RA14-299 | Wsh1-566 | Wsh2-491 |
| Bad à Cheo | Bog | α | α | α | α | α | α | α | α | α | α | α |
| | Young Pine | β | α | α | β | α | α | α | α | α | β | α |
| | Old Pine | β | α | α | β | α | α | α | β | α | γ | β |
| Glensaugh | Grassland | ND | ND | α | α | ND | α | α | α | α | α | α |
| | Young Pine | ND | ND | α | α | ND | α | α | α | α | α | α |
| Craggan | Moorland | ND | α | α | α | α | α | α | α | α | α | ND |
| | Young Birch | ND | α | α | α | α | α | α | α | $\alpha\beta$ | α | ND |
| | Old Birch | ND | α | α | α | α | α | β | β | β | α | ND |
| Tulchan | Moorland | ND | α | α | α | α | α | α | α | α | α | ND |
| | Young Birch | ND | β | β | β | β | β | β | β | β | β | ND |
| | Old Birch | ND | β | β | β | β | $\alpha\beta$ | $\alpha\beta$ | $\alpha\beta$ | $\alpha\beta$ | β | ND |

264

265

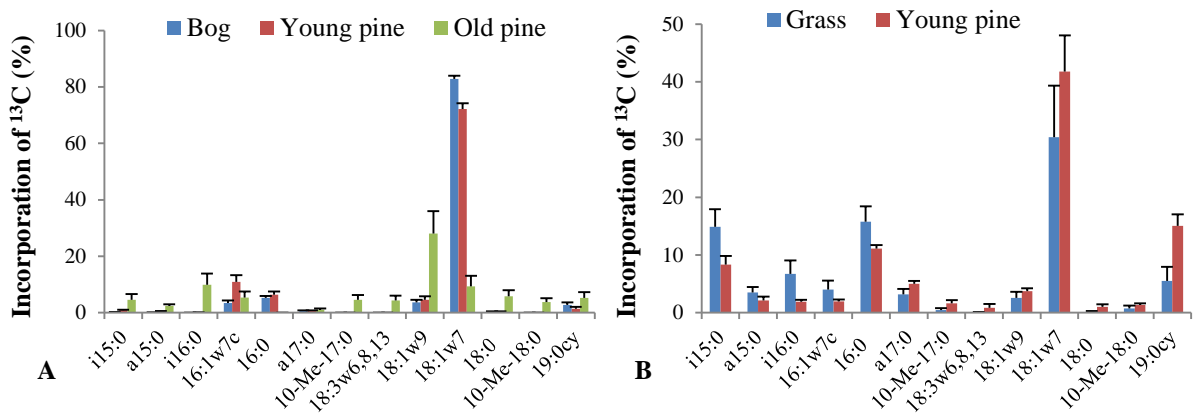
266 **Table S7. Effect of tree growth on the methanotrophic community (PCA from the *pmoA***
 267 **microarray).**

268 The data are *P* values corresponding to the first five PC scores of the probe hybridisation
 269 intensities, and were obtained by MANOVA. Within each column and for each site, results
 270 followed by different Greek letters (α , β , γ) are statistically different for each habitat,
 271 according to multiple pairwise comparison ($\alpha=0.05$).

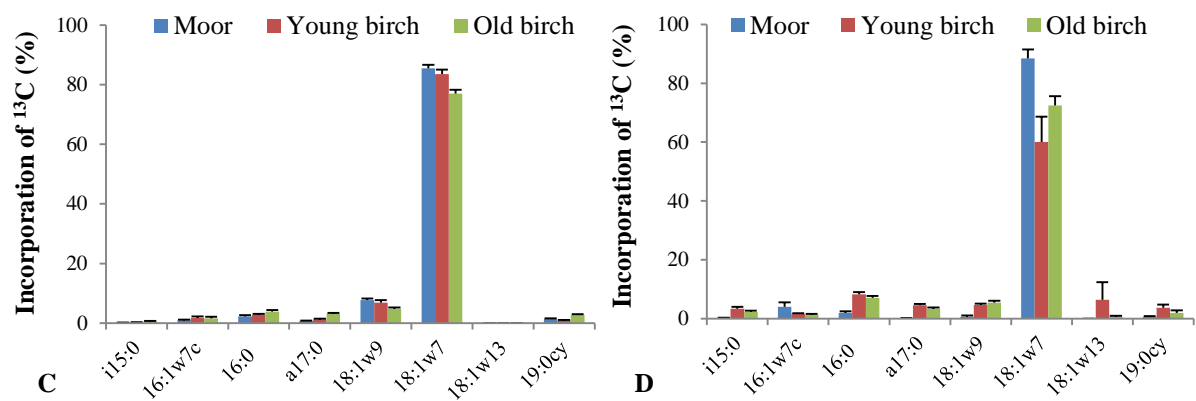
| Site/Habitat | | PC 1 | PC 2 | PC 3 | PC 4 | PC 5 | MANOVA |
|--------------|--------------------|------------------|--------------|------------------|----------|----------|------------------|
| | % variation | 74.23 | 15.93 | 6.07 | 2.22 | 0.66 | |
| | <i>P</i> | <0.001 | 0.215 | 0.469 | 0.640 | 0.695 | 0.003 |
| Bad à Cheo | Bog | α | α | α | α | α | |
| | Young Pine | β | α | α | α | α | |
| | Old Pine | γ | α | α | α | α | |
| | % variation | 63.88 | 34.65 | 1.170 | 0.230 | 0.050 | |
| | <i>P</i> | 0.934 | 0.003 | 0.821 | 0.811 | 0.598 | 0.299 |
| Glensaugh | Grassland | α | α | α | α | α | |
| | Young Pine | α | β | α | α | α | |
| | | | | | | | |
| | % variation | 90.4 | 6.09 | 2.62 | 0.66 | 0.15 | |
| | <i>P</i> | 0.601 | 0.559 | <0.001 | 0.125 | 0.758 | 0.002 |
| Craggan | Moorland | α | α | α | α | α | |
| | Young Birch | α | α | α | α | α | |
| | Old Birch | α | α | β | α | α | |
| | % variation | 89.71 | 5.27 | 3.65 | 1.04 | 0.21 | |
| | <i>P</i> | 0.011 | 0.501 | 0.482 | 0.064 | 0.762 | <0.001 |
| Tulchan | Moorland | α | α | α | α | α | |
| | Young Birch | β | α | α | α | α | |
| | Old Birch | β | α | α | α | α | |

272

273



274



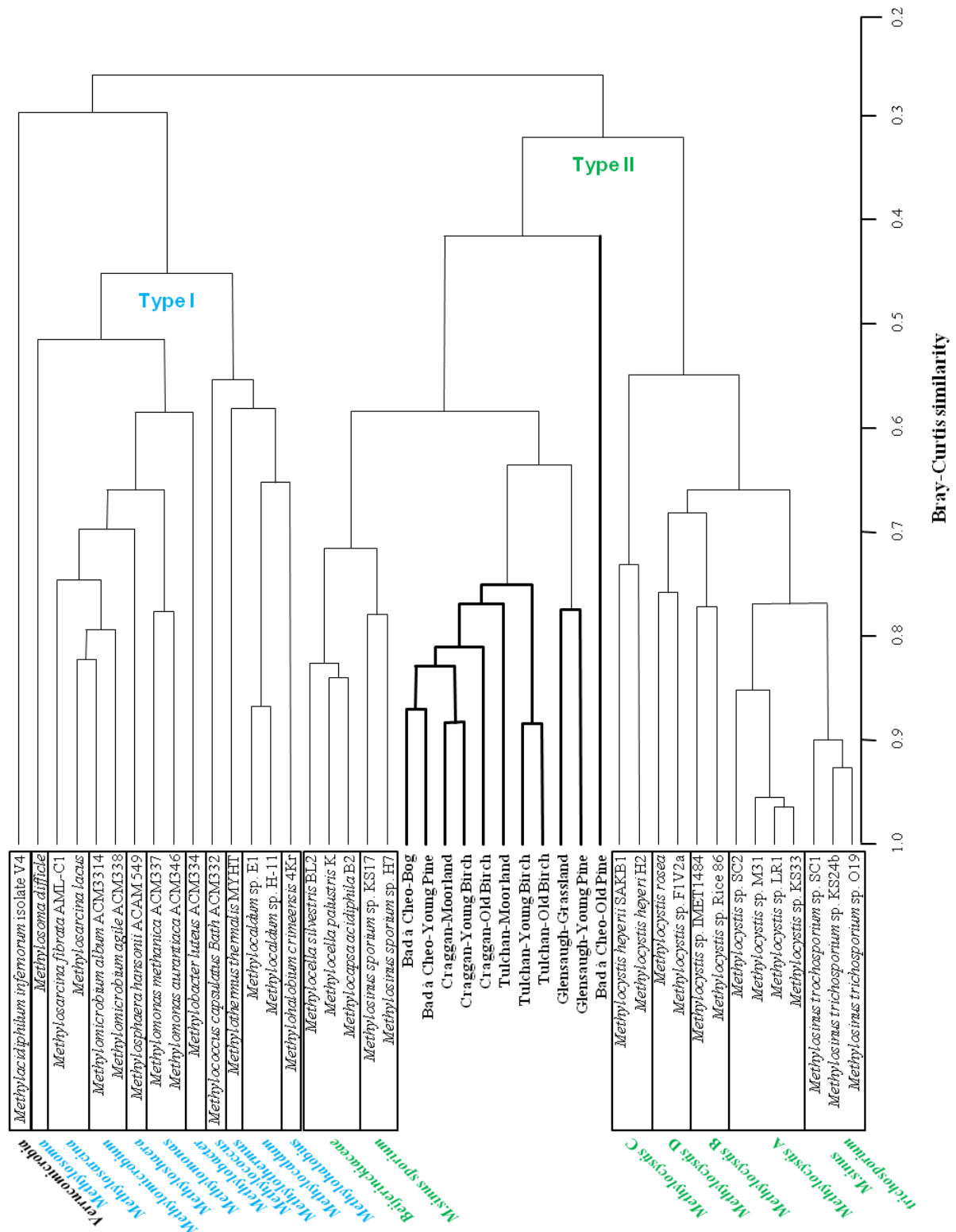
275

276 **Fig. S1. Percentage of incorporation of ^{13}C within the PLFAs after incubation with ~100**
277 **ppm of $^{13}\text{C}\text{-CH}_4$ at (A) Bad à Cheo, (B) Glensaugh, (C) Craggan and (D) Tulchan.**

278 The data are seasonal average \pm s.e.m. (n=8 replicates – autumn and summer combined) of
279 the enriched PLFA content.

280

281



282

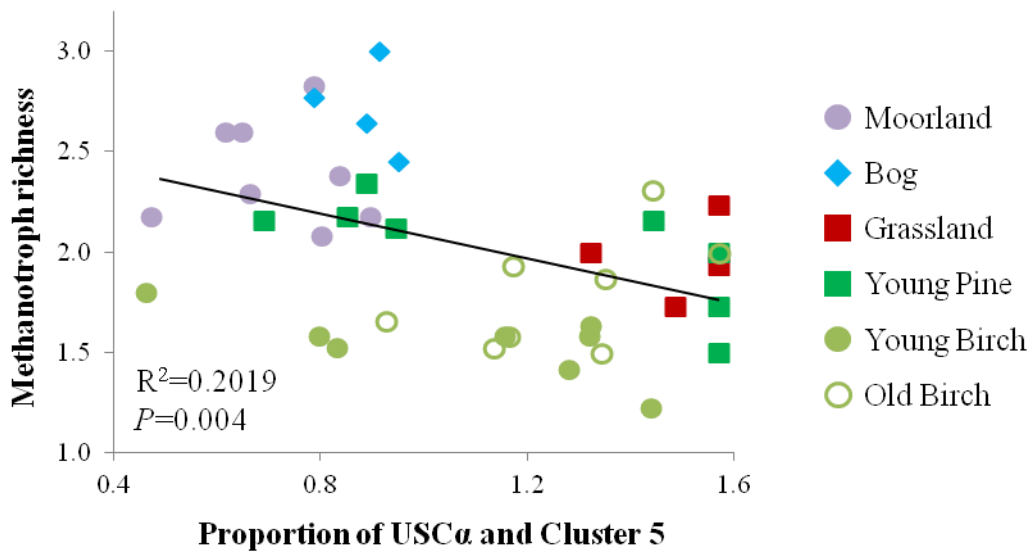
283 **Fig. S2. Cluster analysis of the PLFA-SIP profiles (based on % of ¹³C-incorporation) of**
 284 **methanotrophs in the enriched (~100 ppm ¹³C-CH₄) soils (n=8 replicates - autumn and**
 285 **summer combined).**

286 The above dendrogram was built using data from this study, combined with data from the
287 literature (15). Thus, the active methanotroph population present in the soils investigated
288 could be affiliated with published methanotrophs. This is because PLFA-SIP allows for the
289 identification of ^{13}C -labelled fatty acids produced by active methanotrophs feeding on ^{13}C -
290 CH_4 . The PLFA pattern of an environmental sample from a PLFA-SIP incubation with ^{13}C -
291 CH_4 can then be compared to the PLFA content of a pure culture of methanotrophs. A Bray-
292 Curtis similarity matrix was used, from the square-root transformation of the PLFA data (see
293 Fig. S1), to perform a group average linking cluster analysis with GenStat® software.

294

295

296



297

298 **Fig. S3. Relationship between methanotroph richness and proportion of the USCα**
299 **members associated with land-use change (n=40).**

300 The methanotroph richness was calculated as the square-root transformation of the number of
301 T-RFs present in each sample (among the 15 most abundant T-RFs of the T-RFLP profiles,
302 which constituted >94% coverage). The proportion of USCα microorganisms was calculated
303 as the angular transformation (arcsine of the square root) of the ratio of the relative
304 abundance of the T-RFs specific to *Methylocapsa* sp. (USCα/Cluster 5 – T-RFs *Hha*-32 and
305 *Hha*-129) to the sum of the T-RFs specific to USCα and the *Methylocystaceae* family (T-RF
306 *Hha*-81). Refer to Table 1 in main text for T-RF reference values.

307

308

309 **References**

- 310 1. **Miller JD, Anderson HA, Ray D, Anderson AR.** 1996. Impact of some initial
311 forestry practices on the drainage waters from blanket peatlands. *Forestry* **69**:193-203.
- 312 2. **Anderson A, Pyatt D, Sayers JM, Blackhall SR, Robinson HD.** 1992. Volume and
313 mass budgets of blanket peat in the north of Scotland. *Suo (Helsinki)* **43**:195-198.
- 314 3. **Miles J, Young WF.** 1980. The effects on heathland and moorland soils in scotland
315 and northern england following colonization by birch (*Betula* spp.). *B.Ecol.* **11**:233-
316 242.
- 317 4. **Mitchell RJ, Campbell CD, Chapman SJ, Osler GHR, Vanbergen AJ, Ross LC,**
318 **Cameron CM, Cole L.** 2007. The cascading effects of birch on heather moorland: a
319 test for the top-down control of an ecosystem engineer. *J.Ecol.* **95**:540-554.
- 320 5. **Hester AJ, Miles J, Gimingham CH.** 1991. Succession from heather moorland to
321 birch woodland. I. Experimental alteration of specific environmental conditions in the
322 field. *J.Ecol.* **79**:303-315.
- 323 6. **Miles J.** 1981. Effects of birch on moorlands. Institute of Terrestrial Ecology,
324 Banchory, UK.
- 325 7. **Keith AM, van der Wal R, Brooker RW, Osler GHR, Chapman SJ, Burslem**
326 **DFRP.** 2006. Birch invasion of heather moorland increases nematode diversity and
327 trophic complexity. *Soil Biol.Biochem.* **38**:3421-3430.
- 328 8. **Nazaries L, Tate KR, Ross DJ, Singh J, Dando J, Saggarr S, Baggs EM, Millard**
329 **P, Murrell JC, Singh BK.** 2011. Response of methanotrophic communities to
330 afforestation and reforestation in New Zealand. *ISME J.* **5**:1832-1836.

- 331 9. **Levy PE, Gray A, Leeson SR, Gaiawyn J, Kelly MPC, Cooper MDA, Dinsmore**
332 **KJ, Jones SK, Sheppard LJ.** 2011. Quantification of uncertainty in trace gas fluxes
333 measured by the static chamber method. *Eur.J.Soil Sc.* **62**:811-821.
- 334 10. **Matthias AD, Blackmer AM, Bremner JM.** 1980. A simple chamber technique for
335 field measurement of emissions of nitrous oxide from soils. *J.Environ.Qual.* **9**:251-
336 256.
- 337 11. **Bodrossy L, Stralis-Pavese N, Murrell JC, Radajewski S, Weilharter A, Sessitsch**
338 **A.** 2003. Development and validation of a diagnostic microbial microarray for
339 methanotrophs. *Environ.Microbiol.* **5**:566-582.
- 340 12. **Stralis-Pavese N, Sessitsch A, Weilharter A, Reichenauer T, Riesing J, Csontos J,**
341 **Murrell JC, Bodrossy L.** 2004. Optimization of diagnostic microarray for
342 application in analysing landfill methanotroph communities under different plant
343 covers. *Environ.Microbiol.* **6**:347-363.
- 344 13. **Paterson E, Gebbing T, Abel C, Sim A, Telfer G.** 2007. Rhizodeposition shapes
345 rhizosphere microbial community structure in organic soil. *New Phytol.* **173**:600-610.
- 346 14. **Singh BK, Tate KR, Ross DJ, Singh J, Dando J, Thomas N, Millard P, Murrell**
347 **JC.** 2009. Soil methane oxidation and methanotroph responses to afforestation of
348 pastures with *Pinus radiata* stands. *Soil Biol.Biochem.* **41**:2196-2205.
- 349 15. **Bodelier PL, Gillisen MJ, Hordijk K, Damste JS, Rijpstra WI, Geenevasen JA,**
350 **Dunfield PF.** 2009. A reanalysis of phospholipid fatty acids as ecological biomarkers
351 for methanotrophic bacteria. *ISME J* **3**:606-617.

352

353