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Separation of NH₄⁺ and NO₃⁻

For isotopic analysis, NH_4^+ and NO_3^- were separated by distillation with magnesium oxide and Devarda's alloy^{1,2}. In detail, a portion of the extract was steam-distilled with MgO in a steam distillation system to isolate the NH_4^+ . The sample in the flask was then distilled again after the addition of Devarda's alloy to isolate the $NO₃$. The liberated NH_3 was trapped using a boric acid solution. The trapped N was acidified and converted to $(NH_4)_2SO_4$ using a 0.02 mol $L^{-1}H_2SO_4$ solution. The H_2SO_4 solution containing NH_4^+ was then evaporated to dryness at 65 $^{\circ}$ C in an oven and analysed for ¹⁵N abundance. Before separating NH_4^+ and NO_3^- in the extract using a steam distillation system, the recovery of NH_4^+ and NO_3^- from a standard solution (1 g NH_4^+ -N L⁻¹ and 1 g NO₃⁻-N L⁻¹) was determined. The results showed that almost all of the NH₄⁺-N in the solution could be recovered (>99%), and the recovery of NO₃⁻ was >95%.

Comparison of dilution and 15N tracing model methods

To verify the utility of the $15N$ tracing model for simulating the gross rates of N transformation that occur simultaneously, we compared the results obtained using the ¹⁵N dilution method³ with those obtained using the ¹⁵N tracing method employed in this study.

The gross rates of mineralization, nitrate production, and $NO₃$ consumption per time interval and the average of all time intervals were calculated using the method of Kirkam and Bathalomew $(1954)^3$. The results showed the calculated (dilution technique) and modelled $(^{15}N$ tracing technique) mineralization and $NO₃$ production rates were similar ($p < 0.01$; Fig. 2A, B). Based on the generally strong agreement between the values, we were confident enough to use the modelled rates, which provided details on the N transformation mechanisms and therefore a much more detailed view of actual N dynamics. Therefore, the modelled rates were used to compare soil N transformations between temperate and subtropical-tropical forest soils in the present investigation.

Samples 14 and 15 were from the same core, but the $15N$ tracing experiment was carried out in the laboratory for soil 14 (added 50 μ g N g⁻¹ soil) and in the field for soil 15 (added 2 μg N g^{-1} soil). The results showed that the N transformation rates determined in the laboratory and in the field were comparable (Table 2 in article), and thus the N transformations were not stimulated in the laboratory due to higher N application in the studied acid forest soils. The laboratory studies can provide essential information to mechanistically understand the observed N cycling process in the field^{4,5}, despite some reported problems⁶. Nitrogen enrichment in humid subtropical soils could therefore be explained using data obtained by comparing the gross rates of N transformation in laboratory-incubated soils with those of temperate soils (Fig. 3 in article).

Previous studies, where the $15N$ tracing approach (section 2.4 in article) was used, showed that N consumption/production rates can be separated into process-specific gross rates to provide more details about soil N transformations, despite the fact that there is still some uncertainty about the general nature of the application of the $\rm^{15}N$ tracing model^{5,7,8,9,10,11}. It should be stressed that, in general, individual gross N

transformation rates cannot be directly measured but can only be quantified with the help of analytic models, the most basic of which was presented by Kirkham and Bartholomew $(1954)^7$. Subsequently, more complete and realistic analytic models have been developed. The latest are numerical ^{15}N tracing models in which the N transformation rate parameters are identified by non-linear optimization techniques^{7,10}. We decided to calculate the total NH_4^+ and NO_3^- production and consumption rates via the dilution method (Kirkam and Bathalomew 1954)⁷ and the current ¹⁵N tracing model. Comparison of the model results showed that the two methods provided comparable results for mineralization and $NO₃$ production rates ($p < 0.01$; Fig. 1A, B). The advantage of the $15N$ tracing approach is that not only pool-specific gross N rates, but also individual N rates, can be determined. Furthermore, ^{15}N tracing models overcome the restrictions of zero-order kinetic rates that are assumed in the dilution approach. First-order and Michaelis-Menten kinetics are more realistic because they take into account a non-linear behaviour with respect to changing N concentrations. Discrepancies between the two methods may be associated with the use of different kinetics¹¹ (see also the result for NO_3 ⁻ immobilization for the two neutral temperate forest soils [soil 7, 8], Fig. 1C).

The results of analysis using average of soil properties and gross N transformation rates at each site

* M, mineralization rate of organic N pool; TNi, total nitrification rate (autotrophic nitrification + heterotrophic nitrification); O_{NH4} , autotrophic nitrification; I_{NO3} , immobilization of NO₃; NC, nitrification capacity; NR, NO₃ retention capacity; DNRA, dissimilatory NO_3^- reduction to NH_4^+ .

Table 2 Pearson correlation coefficient between soil pH and autotrophic nitrification

 (O_{NH4}) nitrification capacity (NC) by using site-averages

**. Correlation is significant at the 0.01 level (2-tailed).

Table 3 Pearson correlation coefficient between gross rates of mineralization and

**. Correlation is significant at the 0.01 level (2-tailed).

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Fig. 1 Comparison of modelled and calculated total mineralization (A), nitrification

(B) and NO₃ immobilization rates (C; mg N kg⁻¹ d⁻¹)

The modelled rates were simulated using the $15N$ tracing model (Müller et al., 2007)¹⁰, and the calculated rates were obtained using the method of Kirkam and Bathalomew $(1954)^7$.