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# <span id="page-2-0"></span>60 **S1. Inland Ramsar wetlands**

 According to the Ramsar convention, wetlands are defined as "areas of marsh, fen, peatland or water, whether natural or artificial, permanent or temporary, with water that is static or flowing, fresh, brackish or salt, including areas of marine water the depth of which at low tide does not exceed six 64 metres".<sup>[1](#page-28-3)</sup> The biological importance of the Ramsar sites, as indicated in the Ramsar Sites 65 Information Service  $(RIS)^2$ [,](#page-28-4) is shown in [Table S1](#page-2-2) and [Table S2.](#page-2-3) Note that each Ramsar site can be named several times within the biological importance category.

<span id="page-2-2"></span>67 **Table S1: Biological importance of the Ramsar sites. The number of sites is given for total sites, surface water-fed sites**  68 **and groundwater-fed sites. The percentages are related to the respective total (i.e. 1184 for total, 1033 for surface**  water-fed and 151 for groundwater-fed wetlands).



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#### 71 **Table S2: Biological importance of the Ramsar sites per geographical region.**

<span id="page-2-3"></span>

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# <span id="page-2-1"></span>73 **S2. Overview of species and data sources**

74 We have included different taxa for calculating effect factors of water consumption on biodiversity in

75 wetlands. All species combined can act as a proxy for biodiversity. [Table S3](#page-2-4) lists all considered taxa.

# 76 **Table S3: Overview of taxa, data sources and total number of species. The SI section indicates in which section more**

<span id="page-2-4"></span>information and the calculated maps for the respective taxon can be found.



# <span id="page-3-0"></span>**S3. Bird maps**

- 79 For each bird species a shape file is available from BirdLife and NatureServe  $2011<sup>3</sup>$  $2011<sup>3</sup>$ , indicating the
- range of distribution. Additionally, it gives information on Presence, Origin and Season (see [Table S4\)](#page-3-1)
- 81 that is equally valid for amphibians and reptiles.
- <span id="page-3-1"></span>**Table S4: Codes for presence, origin and season of the dataset of BirdLife and NatureServe, [3](#page-28-5) that are also valid for amphibians and reptiles.**



 Resulting bird maps for the number of non-residential birds and waterbirds and the respective vulnerability scores are shown in [Figure S1](#page-3-2) to [Figure S8.](#page-6-1) For the definition of waterbirds, non-residential birds and the calculation of the vulnerability scores, see the main document.



<span id="page-3-2"></span>**Figure S1: Bird richness map for the waterbird sample based on data from BirdLife and NatureServe[.](#page-28-5) <sup>3</sup> Presence values are chosen from categories 1 to 3, values for season are at 1 to 5 (se[e Table S4\)](#page-3-1).** 





 **Figure S2: Bird vulnerability score (VS) map for waterbirds. Presence values are chosen from categories 1 to 3, values for season are at 1 to 5 (see [Table S4\)](#page-3-1).**



- **Figure S3: Bird richness map for the waterbird sample based on data from BirdLife and NatureServe. [3](#page-28-5) Presence**
- **values are chosen from categories 1 to 4 (instead of 1 to 3), values for season remain at 1 to 5 (se[e Table S4\)](#page-3-1). Base map**
- with country boundaries adapted from ref[.](#page-29-2)<sup>8</sup>



 **Figure S4: Bird vulnerability score (VS) map for waterbirds. Presence values are chosen from categories 1 to 4**  (instead of 1 to 3), values for season remain at 1 to 5 (se[e Table S4\)](#page-3-1).



**Figure S5: Bird richness map for the non-residential birds based on data from BirdLife and NatureServ[e.](#page-28-5) <sup>3</sup> Presence values are chosen from categories 1 to 3, values for season are at 2 to 5 (see [Table S4\)](#page-3-1).**



- **Figure S6: Bird vulnerability score (VS) map for non-residential birds. Presence values are chosen from categories 1**
- **to 3, values for season are at 2 to 5 (see [Table S4\)](#page-3-1).**



**Figure S7: Bird richness map for the non-residential birds based on data from BirdLife and NatureServ[e.](#page-28-5) <sup>3</sup> Presence values are chosen from categories 1 to 4 (instead of 1 to 3), values for season remain at 2 to 5 (see [Table S4\)](#page-3-1).**



<span id="page-6-1"></span>112 **Figure S8: Bird vulnerability score (VS) map for non-residential birds. Presence values are chosen from categories 1**  113 **to 4 (instead of 1 to 3), values for season remain at 2 to 5 (see [Table S4\)](#page-3-1).**

114 The largest difference between species richness calculated with presence values 1 to 3, or with presence values 1 to 4, was 7 for waterbirds (The Bahamas), and 6 for non-residential birds presence values 1 to 4, was 7 for waterbirds (The Bahamas), and 6 for non-residential birds 116 (St.Helena). None of them were in areas where one of our 1184 Ramsar wetlands was located.

## <span id="page-6-0"></span>117 **S4. Reptile maps**

118 Reptile maps were derived based on data from IUCN.<sup>[4,](#page-28-6) [6](#page-29-0)</sup> We only used those species which were 119 classified as having "Wetland (inland)" as habitat. The categories for presence and seasonality from 120 [Table S4](#page-3-1) are valid for reptiles as well. All seasonality values were used, and for presence categories we changed between 1 to 3 and 1 to 4. The maps and corresponding vulnerability scores (VS) are we changed between 1 to 3 and 1 to 4. The maps and corresponding vulnerability scores  $(VS)$  are 122 shown in Figure S9 to [Figure S12.](#page-7-0)



123

**Figure S9: Species richness map of wetland reptiles based on data from IUCN.<sup>[4,](#page-28-6) [6](#page-29-0)</sup> Presence values are chosen from**  $125$  **categories 1 to 3. Base map with country boundaries adapted from ref.<sup>8</sup>** categories 1 to 3. Base map with country boundaries adapted from ref.<sup>[8](#page-29-2)</sup>



126

**Figure S10: Map of the vulnerability score (VS) of wetland reptiles based on data from IUCN.<sup>[4,](#page-28-6) [6](#page-29-0)</sup> Presence values are chosen from categories 1 to 3. Base map with country boundaries adapted from ref.<sup>8</sup>** chosen from categories 1 to 3. Base map with country boundaries adapted from re[f.](#page-29-2)<sup>8</sup>





**Figure S11: Species richness map of wetland reptiles based on data from IUCN.<sup>[4,](#page-28-6)6</sup> Presence values are chosen from categories 1 to 4 (instead of 1 to 3). Base map with country boundaries adapted from ref.<sup>8</sup>** categories 1 to 4 (instead of 1 to 3). Base map with country boundaries adapted from ref.<sup>[8](#page-29-2)</sup>



<span id="page-7-0"></span>**Figure S12:** Map of the vulnerability score (VS) of wetland reptiles based on data from IUCN.<sup>[4,](#page-28-6)6</sup> Presence values are chosen from categories 1 to 4 (instead of 1 to 3). Base map with country boundaries adapted from ref.<sup></sup> chosen from categories 1 to 4 (instead of 1 to 3). Base map with country boundaries adapted from ref.<sup>[8](#page-29-2)</sup>

## <span id="page-8-0"></span>**S5. Amphibian maps**

136 Amphibian maps were derived based on data from IUCN.<sup>[4,](#page-28-6) [5](#page-28-7)</sup> All amphibian species were used. The categories for presence and seasonality from [Table S4](#page-3-1) are valid for reptiles as well. All seasonality values were used, and for presence categories we altered between 1 to 3 and 1 to 4. The maps and corresponding vulnerability scores (VS) are shown in [Figure S13](#page-8-1) to [Figure S16.](#page-9-1)



- **Figure S13: Species richness map of amphibians based on data from IUCN.**<sup>[4,](#page-28-6)5</sup> Presence values are chosen from categories 1 to 3. Base map with country boundaries adapted from ref.<sup>8</sup> categories 1 to 3[.](#page-29-2) Base map with country boundaries adapted from ref.<sup>8</sup>
- 

<span id="page-8-1"></span>

- **Figure S14: Map of the vulnerability score (VS) of amphibians based on data from IUCN.<sup>[4,](#page-28-6)5</sup> Presence values are chosen from categories 1 to 3. Base map with country boundaries adapted from ref.<sup>8</sup>**
- **chosen from categories 1 to 3. Base map with country boundaries adapted from ref[.](#page-29-2)<sup>8</sup>**



**Figure S15: Species richness map of amphibians based on data from IUCN.<sup>[4,](#page-28-6)5</sup> Presence values are chosen from 148 categories 1 to 4 (instead of 1 to 3). Base map with country boundaries adapted from ref.<sup>8</sup>** categories 1 to 4 (instead of 1 to 3). Base map with country boundaries adapted from ref.<sup>[8](#page-29-2)</sup>



<span id="page-9-1"></span>**Figure S16: Map of the rarity score of amphibians based on data from IUCN.<sup>[4,](#page-28-6)5</sup> Presence values are chosen from**  $151$  **categories 1 to 4 (instead of 1 to 3). Base map with country boundaries adapted from ref.<sup>8</sup>** categories 1 to 4 (instead of 1 to 3). Base map with country boundaries adapted from ref.<sup>[8](#page-29-2)</sup>

## <span id="page-9-0"></span>**S6. Mammal maps**

- Maps for the number of water-dependent mammals and the respective vulnerability scores are shown
- in [Figure S17](#page-9-2) to [Figure S20.](#page-10-1) The mammal suitability model was developed by the Global Mammal
- 155 Assessment (see e.g.  $ref^7$ [\)](#page-29-1)



#### 

 **Figure S17: Species richness of water-dependent mammals based on the extent of occurrence of the mammals. Base**  map with country boundaries adapted from ref[.](#page-29-2)<sup>8</sup>

<span id="page-9-2"></span>

- **Figure S18: Vulnerability score (VS) of the water-dependent mammals, based on the extent of occurrence of the**
- mammals. Base map with country boundaries adapted from ref.<sup>[8](#page-29-2)</sup>



 **Figure S19: Species richness of water-dependent mammals based on the suitable habitat of the mammals (AOO). Base**  map with country boundaries adapted from ref[.](#page-29-2)<sup>8</sup>



<span id="page-10-1"></span> **Figure S20: Vulnerability score (VS) of the water-dependent mammals, based on the suitable habitat of the mammals (AOO).** Base map with country boundaries adapted from ref.<sup>[8](#page-29-2)</sup>

 Difference between species richness map with EOO and AOO is small. The reason is that, as water- dependent mammals are limited to waterbodies, the difference between AOO and EOO is not large, and that AOO represents a nested subset of the EOO data. The geographical outermost boundary can

only follow waterbody borders, as in the AOO.

### <span id="page-10-0"></span>**S7. CpA – Waterbody count per area**

 The waterbody count per area (CpA) data set is derived, as described in the main manuscript, based on [t](#page-29-3)he rivers of the world dataset<sup>9</sup> and the global lakes and wetland database<sup>[10](#page-29-4)</sup> by counting how many points (i.e. waterbodies or river sections) fall into each sub-watersheds area (N in [Equation S1\)](#page-11-1)[.](#page-29-3) Dividing the number of points by the area of the sub-watershed (A in [Equation S1\)](#page-11-1) and multiplying 177 with an aridity index (precipitation  $P<sup>11</sup>$  $P<sup>11</sup>$  $P<sup>11</sup>$  divided by potential evapotranspiration  $PET<sup>12</sup>$  $PET<sup>12</sup>$  $PET<sup>12</sup>$ ) leads to a value set that tells us for each pixel i how large the habitat loss risk in the network of waterbodies in each pixel is (taking into account a potential larger density of temporary pools by multiplying with the aridity index). By dividing all values with the maximum value, CpA is scaled between 0 and 1 and becomes unitless. If there is little water, the pixel had higher chances of becoming unsuitable as habitat, thus if the CpA is small, the habitat loss risk is large.

$$
CpA_i = \frac{N_{per\,subwatershed,i}}{A_{subwatershed,i}} \cdot \frac{P_i}{PET_i}
$$

$$
CpA_i = \frac{N_{subwatershed,i}}{max\,CpA_i}
$$

#### <span id="page-11-1"></span>**Equation S1**

 Iceland, Norway, Finland, as well as parts of Sweden and Eastern Russia are not covered in the dataset 186 for the watersheds[.](#page-29-3)<sup>9</sup> The closest available CpA values were thus assumed to be valid in the administrative regions of these countries which were missing for calculating the CpA. As they have a high CpA, they are not relevant and this simplification is acceptable. For remote islands for which no P, PET, rivers and lakes data were available in global databases (e.g. Azores), a CpA of 1 was assumed. Since there was no indication about counts of waterbodies, we decided to set CpA to 1 in these cases, although this was not a conservative assumption and the damage is likely to be underestimated. However, this concerns only very few wetlands on individual islands and these small, data deficient wetlands would need a closer look in future. For islands close to the mainland, the closest mainland value was assigned to the island (e.g. Malta received its value from Sicily).

The resolution of the CpA data set [\(Figure S21\)](#page-11-2) is 0.167°x0.167° since this was the resolution of the

precipitation dataset (coarsest dataset).



<span id="page-11-2"></span>**Figure S21: Habitat loss risk index CpA.**

 CpA=1 are areas with a large density of waterbodies, and as the value approaches zero, wetland habitats get more rare. The smaller the CpA, the rarer the waterbodies and the larger the threat of losing the habitat type "wetland/waterbody" in the region, when water is consumed.

## <span id="page-11-0"></span>**S8. Species-area relationship and z-values**

 The formula of the species-area relationship is shown in [Equation S2.](#page-11-3) The species richness S can be predicted from a habitat area A, an exponent z indicating the slope of the species richness curve and a 205 constant c. z is often determined for specific curves, but a common value of  $0.25$  is often applied.<sup>[13](#page-29-7)</sup>

$$
S = c \cdot A^Z
$$

<span id="page-11-3"></span>**Equation S2**

208 For a known area change a new species richness  $S<sub>new</sub>$  can thus be predicted based on the original area and species richness, as shown in [Equation S3.](#page-12-1)

$$
S_{new} = \left(\frac{A_{new}}{A_{original}}\right)^{z} \cdot S_{original}
$$

212 The number of lost species  $S<sub>lost</sub>$  from the original area is thus [\(Equation S4\)](#page-12-2).

$$
S_{lost} = S_{original} - S_{new} = S_{original} - \left(\frac{A_{new}}{A_{original}}\right)^{z} \cdot S_{original} = \left(1 - \left(\frac{A_{new}}{A_{original}}\right)^{z}\right) \cdot S_{original}
$$

$$
\frac{213}{244}
$$

<span id="page-12-2"></span>**Equation S4**

 The range of the z values applied for the different taxa are shown in [Table S5.](#page-12-3) All values are taken  $\frac{14}{14}$  $\frac{14}{14}$  $\frac{14}{14}$  For birds and mammals, we only used the z-values from nested studies because they best represent pure diversity change over different sampling area sizes, are best suited to the power model employed, and are more suited for extrapolation beyond the range of area sizes used to 219 derive the z-value.<sup>[15](#page-29-9)</sup> For reptiles and amphibians, we used z-values from independent (non-nested) studies from the same data source, since no values for nested studies were available. Values across 221 multiple studies for a single taxon were averaged to derive taxon-specific<sup>[14](#page-29-8)</sup> values. All z-values are 222 close to the commonly used z-value of  $0.25^{13, 16, 17}$  $0.25^{13, 16, 17}$  $0.25^{13, 16, 17}$  $0.25^{13, 16, 17}$  $0.25^{13, 16, 17}$ 

**Taxon zmin zmax zaverage No. of studies used** birds 0.15 0.63 0.37 8 (nested) mammals  $-0.24$  0.93 0.34 4 (nested) reptiles 0.08 0.81 0.33 10 (independent) amphibians  $0.04$   $0.36$   $0.20$   $18$  (independent)

<span id="page-12-3"></span>**Table S5: Minimum, maximum and average slopes of the species-area relationship for the different taxa.[14](#page-29-8)**

## <span id="page-12-0"></span>**S9. Characterization factors – determining individual catchments**

 All characterization factors (CFs) are applicable on a larger scale than just the wetland itself. The reasoning is explained below for surface water-fed and groundwater-fed wetlands separately.

#### **Surface water-fed wetlands**

 A surface water-fed wetland is fed by inflowing water from the catchment upstream of the wetland. If water is consumed anywhere in the area upstream of the wetland, inflow will be reduced and the wetland will be damaged. Therefore, the CF for this wetland is applicable in the whole watershed of the wetland (e.g. blue watershed in [Figure S22\)](#page-13-1). A second wetland, which is for instance situated upstream of the first one, may receive water from partly the same sources. But other rivers, for example, may be completely irrelevant for the second wetland, because they do not flow into this specific wetland. Thus, the CF for the second wetland is applicable in another area, which is the individual catchment of the second wetland (e.g. orange area in [Figure S22\)](#page-13-1). However, water which is consumed in the range area does not reach both wetlands and therefore both of them are damaged and the CFs of both are applicable. That means that the CFs are summed in these areas. This procedure is repeated for all 1033 surface water-fed wetlands and leads to the global maps.

<span id="page-12-1"></span>**Equation S3**



<span id="page-13-1"></span> **Figure S22: Schematic representation of two surface water-fed wetlands (red dots) in their respective watersheds (blue and orange). The river network is shown in black.**

#### **Groundwater-fed wetlands**

 Here the relevant area is calculated according to the hydrogeological condition surrounding the wetland (not upstream-downstream as in the surface water-fed wetland cases). The Area of Relevance 245 (AoR) has been used for the calculation of the FF before (for details see ref<sup>[18](#page-29-12)</sup>). In principle, we determined circles around the wetlands from which water is being drawn to the wetland (imagining the wetland to act like a pump). The decrease in groundwater level due to pumping anywhere in this Area of Relevance influences the infiltrating amount into the wetland. Thus, any pumping within this area leads to damage and thus the CF is applicable in the whole AoR. If there is a second groundwater-fed wetland and their AoRs overlap, the CFs are summed, because it was assumed that pumping in that region will damage both wetlands.



 

 **Figure S23: Schematic representation of two Areas of Relevance (AoR, hatched circles) around two groundwater-fed wetlands (blue and violet circle). The orange part is the area where the AoRs overlap and where CFs are thus**  summed.

## <span id="page-13-0"></span>**S10. EFs and CFs**

 In [Figure S24,](#page-14-0) an overview of all the necessary parameters and modeling steps for the calculation of the EF is shown. As an example, all the values of the parameters for lake Naivasha and lake Elmenteita (both in Kenya) are given in [Table S6.](#page-14-1) These two wetlands are used in the application example and have individual catchments. For the location Bleiswijk, no example is provided since the used factor at the location consists of a several overlaying catchments of wetlands within the Rhine watershed.



<span id="page-14-0"></span>265 **Figure S24: Overview of the parameters and their origin, as well as the modeling steps that are required for**  266 **calculating the EFs of a wetland. Modeling steps are indicated in bold.**

<span id="page-14-1"></span>

267	Table S6: Examples for all required parameters for calculating EFs for lake Naivasha (SW-fed, Kenya) and lake
268	Elmenteita (GW-fed, Kenya).



- 269 Bird species are present in all wetlands. This does not apply to the other taxa. In [Table S7,](#page-15-0) the number
- 270 of wetlands is shown that do not harbour mammals, reptiles or amphibians.

#### <span id="page-15-0"></span>271 **Table S7: Number of wetlands that do not contain a certain taxa in absolute numbers and as percentage of wetland**  type. SW stands for surface water-fed wetlands and GW for groundwater-fed wetlands.



273

274 Effect and characterization factors calculated on the basis of the waterbody area are presented in [Table](#page-15-1) 

275 [S8.](#page-15-1)

<span id="page-15-1"></span>**Table S8: Effect factors [species-eq/m<sup>2</sup> ] and characterization factors [species-eq·yr/m<sup>3</sup>** 276 **] for waterbirds, non-residential birds, water-dependent mammals, wetland reptiles, amphibians and all combined based on the area of the waterbodies within the Ramsar sites. Factors are presented summarized for surface-fed wetlands with surface water consumption (SW) and groundwater-fed wetlands with groundwater consumption (GW). Presence categories are 1 to** 

280 **3 (birds, reptiles, amphibians). CV is the coefficient of variation.**



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- 286 Characterization factors (CFs) calculated with the waterbody areas within the Ramsar area are shown
- 287 in [Figure S25](#page-16-0) to [Figure S36.](#page-19-1)



289 Figure S25: CFs for surface water-fed wetlands with surface water consumption for waterbirds (presence 1 to 3) based on waterbody areas within the Ramsar areas. Base map with country boundaries adapted from ref.<sup>8</sup> **based on waterbody areas within the Ramsar areas. Base map with country boundaries adapted from re[f.](#page-29-2)<sup>8</sup>** 

<span id="page-16-0"></span>

291

292 **Figure S26: CFs for groundwater-fed wetlands with groundwater consumption for waterbirds (presence 1 to 3) based on waterbody areas within the Ramsar areas. Base map with country boundaries adapted from ref[.](#page-29-2)<sup>8</sup>** 293







297

298 **Figure S28: CFs for groundwater-fed wetlands with groundwater consumption for non-residential birds (presence 1**  to 3) based on waterbody areas within the Ramsar areas[.](#page-29-2) Base map with country boundaries adapted from ref<sup>8</sup>





302 **Figure S29: CFs for surface water-fed wetlands with surface water consumption for water-dependent mammals based on waterbody areas within the Ramsar areas. Base map with country boundaries adapted from ref[.](#page-29-2)<sup>8</sup>** 303



305 **Figure S30: CFs for groundwater-fed wetlands with groundwater consumption for water-dependent mammals based on waterbody areas within the Ramsar areas. Base map with country boundaries adapted from ref[.](#page-29-2)<sup>8</sup>** 306



308 **Figure S31: CFs for surface water-fed wetlands with surface water consumption for wetland reptiles based on**  waterbody areas within the Ramsar areas. Base map with country boundaries adapted from re[f.](#page-29-2)<sup>8</sup>





311

312 **Figure S32: CFs for groundwater-fed wetlands with groundwater consumption for wetland reptiles based on**  waterbody areas within the Ramsar areas. Base map with country boundaries adapted from re[f.](#page-29-2)<sup>8</sup>



315 **Figure S33: CFs for surface water-fed wetlands with surface water consumption for amphibians based on waterbody** 



318 **Figure S34: CFs for groundwater-fed wetlands with groundwater consumption for amphibians based on waterbody areas within the Ramsar areas. Base map with country boundaries adapted from ref[.](#page-29-2)<sup>8</sup>** 319



320

321 **Figure S35: CFs for surface water-fed wetlands with surface water consumption for all taxa combined, based on**  waterbody areas within the Ramsar areas. Base map with country boundaries adapted from re[f.](#page-29-2)<sup>8</sup>



323

<span id="page-19-1"></span>324 **Figure S36: CFs for groundwater-fed wetlands with groundwater consumption for all taxa combined, based on**  waterbody areas within the Ramsar areas. Base map with country boundaries adapted from re[f.](#page-29-2)<sup>8</sup>

# <span id="page-19-0"></span>326 **S11. Sensitivities and correlations of EF and CF**

- 327 The correlation between CFs of different taxa and correlations between different parts of the EF and
- 328 the EF itself are shown in Table S9 and Table S10.

329 **Table S9: Spearman's rank correlation coefficient ρ between CFs of different taxa and for both surface water- fed** 

330 **(SW) and groundwater-fed (GW) wetlands. The color code indicates whether there is medium (ρ > 0.3, green with** 

331 **green letters) or high correlation (ρ > 0.5, blue with yellow letters).**



Table S10: Correlations between components of the effect factor and effect factor itself (EF), as well as the fate factor (FF). Correlations are calculated for each taxa separately, except for the correlation between CpA a

353 **(FF). Correlations are calculated for each taxa separately, except for the correlation between CpA and FF, which is** 

354 **the same for all taxa and is thus only calculated for SW-fed and GW-fed wetlands. VS is the vulnerability score, S the**  species number and CpA is the habitat loss risk index.



 We calculated the histogram of species richness, a dominant factor for the EF, and EFs themselves for all taxa. They are shown for SW-fed and GW-fed wetlands in [Figure S37.](#page-22-0) In [Figure S37A](#page-22-0), mammals are with 912 wetlands highest in the bin category 10-20 species. Also, reptiles and amphibians are mostly represented by 10-20 species (840 wetlands and 539 wetlands, respectively). The largest number of wetlands for non-residential bird species is 235 in the species richness category 40-50 species. Only waterbirds show their maximum in an even higher category (80-90 species in 128 wetlands). The distribution of species richness is widest for waterbirds. In GW-fed wetlands [\(Figure](#page-22-0)  [S37B](#page-22-0)), waterbird species are present in 29 wetlands with between 70 and 80 species. Non-residential birds have their maximum with 50-60 species in 35 wetlands, and this is again a bit lower than for SW-fed wetlands. Mammals, reptiles and amphibians are all mostly present with 10-20 species (123, 123 and 83 wetlands, respectively). For the EFs, the most frequent bins for the EFs are those between  $10^{-11}$  species-eq/m<sup>2</sup> and  $10^{-8}$  species-eq/m<sup>2</sup> [\(Figure S37C](#page-22-0) and D).





<span id="page-22-0"></span>369 **Figure S37: Distribution of the species richness for A) SW-fed wetlands and B) GW-fed wetlands. The distribution for**  370 **the EF factors is shown in C) for SW-fed wetlands and D) for GW-fed wetlands.** 

371 A large part of the sensitivity of the characterization factor comes from the fate factor (FF). The 372 sensitivity of the FF is discussed in detail in ref<sup>[18](#page-29-12)</sup>. In [Figure S38,](#page-23-0) the differences between the FFs for 373 different amounts of water consumption is shown as factor of the FF with 10  $\text{m}^3/\text{yr}$  consumption 374 divided by the FF with  $1'000'000 \text{ m}^3/\text{yr}$  consumption. For the SW-fed wetlands, the differences are 375 small, since the factor varies over the whole globe only between 1 and 1.167. For the GW-fed 376 wetlands, the non-linearity of the well formula shows in the much larger differences between the FFs, 377 distributed over the world. As stated in ref<sup>[18](#page-29-12)</sup>, caution should thus be applied when using the factors for 378 GW-fed wetlands.



<span id="page-23-0"></span>379 **Figure S38: Factor between FFs with different amounts of water consumption. A) for SW-fed wetlands. B) for GW**fed wetlands. Base map with country boundaries adapted from re[f.](#page-29-2)<sup>8</sup>

 Whether the CF is calculated with the Ramsar area or the waterbody area can make a substantial difference, as is shown in [Table S11](#page-24-0) exemplarily for waterbirds (being the most dominant taxon). For the other taxa the trend is similar. The table shows the relation between values that are once calculated (CF, EF, FF) or extracted from maps (S, VS, CpA) based on Ramsar areas and once based on waterbody areas. The median, as well as the 2.5% and 97.5% percentile (between them are 95% of the data), are shown for these factors between Ramsar and waterbody areas based results. The difference between species richness, CpA and the vulnerability score is very low. However, the difference in the EF is larger, which is due to the non-linearity of the species-area relationship. In 95% of the data, the difference of the FF is smaller than for the EF. For the GW-fed wetlands, the difference between the FFs regarding waterbody or Ramsar area is negligible, because differences between the areas are often

- 391 smaller. This is probably a coincidence. Bear in mind that the sample for the SW-fed wetlands is 392 almost 10 times larger than for the GW-fed wetlands. The differences between CFs and EFs calculated 393 with either Ramsar or wetland areas for the other taxa, are shown in [Table S12.](#page-24-1) Since the differences 394 for the FFs and the underlying area are the same, they are not repeated. The minor differences from S,
- 395 VS and CpA are not indicated.

<span id="page-24-0"></span>396 **Table S11: Factors between different parameters calculated based on the waterbody and the Ramsar area of each**  397 **wetland, respectively. Between the 2.5% percentile and 97.5 % percentile 95% of the wetlands are found. These**  values are for waterbirds, as this is the most dominant taxon and covers all wetlands.

Factor between waterbody	SW-fed wetlands			GW-fed wetlands		
and Ramsar value	median	2.5% percentile	97.5% percentile	median	2.5% percentile	97.5% percentile
CF	1.25	0.92	70.14	1.60	0.64	62.44
EF	1.67	0.97	55.24	1.67	1.00	63.23
FF	0.93	0.12	10.53	1.00	0.19	1.00
Underlying wetland area (A)	0.60	0.02	1.00	0.60	0.02	1.00
Species richness $(S)$	1.00.	0.96	1.04	1.00	0.98	1.02
Habitat loss risk (CpA)	1.00	0.92	1.09	1.00	0.91	1.09
Vulnerability score (VS)	1.00	0.94	1.06	1.00	0.98	1.03

<span id="page-24-1"></span>400 **Table S12: Factors for the taxa other than waterbirds between CF and EF calculated based on the waterbody and the**  401 **Ramsar area of each wetland, respectively. Between the 2.5% percentile and 97.5 % percentile 95% of the wetlands**  are found. Factors between the FFs or the underlying areas are the same as for waterbirds.

Taxa	Factor between	SW-fed wetlands			GW-fed wetlands		
	waterbody and Ramsar value			median 2.5% percentile 97.5% percentile		median 2.5% percentile	97.5% percentile
non-residential	CF	1.24	0.88	73.31	1.60	0.64	61.74
birds	EF	1.67	0.97	58.29	1.67	1.00	63.43
water-dependent	CF	1.22	0.90	53.65	1.50	0.67	32.49
mammals	EF	1.67	0.99	34.12	1.67	1.00	41.31
	CF	1.20	0.93	55.82	1.58	0.57	23.13
reptiles	EF	1.67	1.00	42.12	1.67	1.00	28.27
	CF	1.24	0.85	57.64	1.57	0.63	44.07
amphibians	EF	1.67	0.96	49.39	l.67	1.00	51.32

403

 Overviews of the sensitivity analysis for the characterization factors (for each taxon and water source separately) are shown in [Figure S39](#page-26-0) and [Figure S40.](#page-27-0) The parameters that were important in the 406 sensitivity analyses of the fate factors  $(FF)^{18}$  $(FF)^{18}$  $(FF)^{18}$  were used here again. In addition, the influence of including possibly extinct species (presence category 4, see [Table S4\)](#page-3-1) was checked (for mammals: area of occupancy instead of extent of occurrence). As for the FFs, the amount of water consumed, surface water flow volumes, and hydraulic conductivity were most relevant. However, due to the non- linear species-area relationship, the underlying area is now relevant for both SW-fed and GW-fed wetlands, while for the FF it was only relevant for some GW-fed wetlands. The influence of including possibly extinct species is in the majority of cases negligible. For the mammals, the change from the extent of occurrence (EOO) to the actual area of occupancy (AOO) of the species had a slightly larger influence, showing that for future developments the derivation of AOOs is relevant.





surface flow volumes [-]

Change from EOO to AOO [-]

 $\sim$ 

**AOO** 

own model

**EOO** 

WaterGap



 **Figure S39: Sensitivity analysis overview for SW-fed wetlands and all taxa separately. The change in each parameter is given for the global median values, parameters are listed below the graphic. Too large changes are indicated next to the marker.**

<span id="page-26-0"></span>





<span id="page-27-0"></span>Figure S40: Sensitivity analysis overview for GW-fed wetlands and all taxa separately. The change in each parameter is given for the global median values, parameters are listed below the graphic. Too large changes are indi 418 **the graphic. Too large changes are indicated next to the marker.**

# <span id="page-28-0"></span>419 **S12. Example of comparison of CFs calculated with PDFs and species-eq.**

420 The CF in PDF $\cdot$ yr/m<sup>3</sup> of the two SW-fed wetlands "Cheyenne Bottoms" (USA) and "Lake Ånnsjön" 421 (Sweden) is the same  $(3.38 \cdot 10^{-9} \text{ PDF-yr/m}^3)$ . However, the number of waterbird species present is 422 very different, being 112 in "Cheyenne Bottoms" and 49 in "Lake Ånnsjön". Absolute species loss 423 was one order of magnitude smaller in the Swedish wetland. VS is the same order of magnitude in 424 both wetlands  $(2.5 \cdot 10^{-6}$  for Cheyenne Bottoms and  $1.1 \cdot 10^{-6}$  for Lake Ånnsjön) and CpA one order of 425 magnitude lower in "Cheyenne Bottoms", showing that the wetland habitat loss risk is larger than in 426 "Lake Ånnsjön". CFs are consequently different for those two wetlands in the species-eq approach 427 (CF for "Cheyenne Bottoms"  $1.83 \cdot 10^{-10}$  species-eq·yr/m<sup>3</sup>, for "Lake Ånnsjön"  $5.8 \cdot 10^{-12}$  species-428 eq·yr/m<sup>3</sup>).

# <span id="page-28-1"></span>429 **S13. Agricultural water requirement ratio**

430 The consumptive share of the water use can be used for estimating water consumption amounts from

431 withdrawn water amounts (for agriculture). It is based on data from AQUASTAT<sup>[19](#page-29-13)</sup> for 93 developing

432 countries and on data from Döll and Siebert<sup>[20](#page-29-14)</sup> for other world regions and is shown in [Figure S41.](#page-28-8)



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<span id="page-28-8"></span>**Figure S41: Water requirement ratios based on data from AQUASTAT<sup>[19](#page-29-13)</sup> and Döll and Siebert ([20](#page-29-14)02).<sup>20</sup> Base map with country boundaries adapted from ref.<sup>8</sup>** with country boundaries adapted from ref[.](#page-29-2)<sup>8</sup>

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