

Identifying Achilles' heel of multi-host pathogens: The concept of
keystone “host” species illustrated by *Mycobacterium ulcerans*
transmission

Supplementary materials

1 Model for environmental transmission and model fit

In addition of the model detailed in the main text, we have developed an alternative model where pathogen transmission is not transmitted between the different kind of hosts through a network of transmission, but only via free-living particles in the environmental aquatic reservoir (compartment B). This kind of model has been successfully applied for various systems, especially for cholera transmission (Codeço 2001; de Magny et al. 2005). We assume that contact rates between each (host) taxon and the contaminated environment are equal across the different aquatic taxa. Put in mathematical terms, that led to the following set of equations:

$$\frac{dS_i}{dt} = b_i N_i - \omega \frac{B}{B + \theta} S_i - d_i S_i \quad (1)$$

$$\frac{dI_i}{dt} = \beta \frac{B}{B + \theta} S_i - d_i I_i \quad (2)$$

$$\frac{dR_i}{dt} = \gamma I_i - d_i R_i \quad (3)$$

$$\frac{dB}{dt} = \sigma \sum_i^n I_i - \varepsilon B \quad (4)$$

where additional parameters ω , θ , σ and ε represent efficient contact rate to water, mycobacterial load needed to yields infection, bacterial production by infectious individuals into environment and mycobacterium lifespan stage into the aquatic environment, respectively.

The environmental transmission model shows also a significant correlation (figure S1), but only when taxa are considered without immunity ($r=0.90$, $p\text{-value}<0.05$ without immunity, and $r=0.3$, $p\text{-value}=0.6851$ with immunity). Moreover, this correlation coefficient is lower than in the case of the pathogen transmission through networks. Consequently, we have considered that a transmission of pathogen through ecological networks as the most likely scenario according to the field observations.

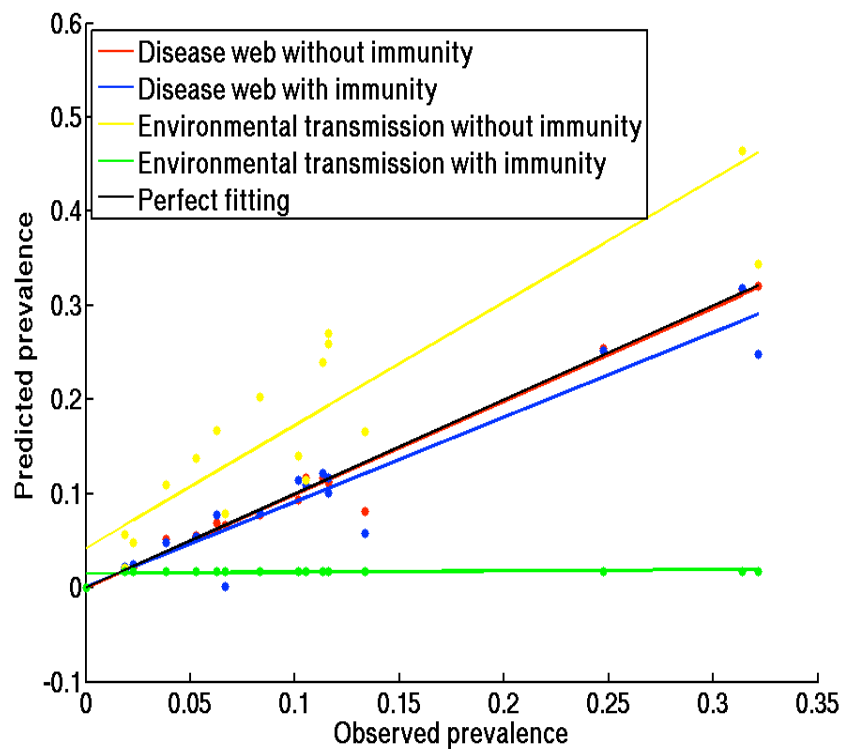


Figure 1: Fitting quality of the two models. The colored lines represent the linear regression between observed and predicted transmission prevalence means for the different hypothesis tested in the present work. The black line represents the perfect fit (*i.e.*, the linear regression without intercept and a slope coefficient equals to 1).

2 Quantifying intrinsic ecological parameters

Except taxa abundances that have been estimated from field data and parameters of networks construction that are estimated by our mathematical model, we have to quantify the demographic parameters of each taxon, *i.e.*, birth and death rate. To keep a realistic community characterization, we used a theoretical framework that has been already assessed in other studies on aquatic ecosystems (Cohen et al. 2003) by using allometric relationships. We first quantify the body mass of each taxon by using the following relationship:

$$\log(M_i) = e - f \log(i)$$

where M_i is the body mass of each taxon rank i , and e and f two constants. Since more abundant taxa tend to live longer (West et al., 1999), we sorted each (host) taxon into minimal to maximal values according to their abundance and we assumed that each taxon represented a rank in the previous relationship (i). To quantify a taxon's growth rate, we applied the following allometric relationship (De Leo & Dobson 1996; West et al. 1999):

$$r_i = 0.6M_i^{-0.27}$$

where r_i is the growth rate of taxa i . Since we kept population size constant ($b_i = d_i = r_i$), we have assumed throughout the manuscript that $e = 0.1$ and $f = 0.05$ (Jonsson et al. 2005).

3 Estimation of parameters

To estimate the parameters for network construction, we estimate the maximal likelihood by assuming that errors are normally distributed. Hence, the quantity to maximize is:

$$-LL(m(x)|s) = \log\left(\prod \frac{1}{\sqrt{2\pi\sigma^2}} \frac{(s(i) - m(i))^2}{2\sigma^2}\right) \quad (6)$$

where s and m represent respectively collected data and simulated values, index i the different (host) taxa, x is the set of parameters and σ^2 is the variance of observed prevalence. To minimize this value, we use the classical Nelder-Menson algorithm where seeds are randomly determined 25 times and we take the best fit solution. We repeat this procedure for the 17 sites in which *M. ulcerans* has been observed.

For network model, we constraint exploration of parameters values to $1 \leq \rho \leq n$, $1 \leq \delta \leq \rho$, $10e - 8 \leq \beta[ii] \leq 10e - 1$, $10e - 8 \leq \beta[ij] \leq 10e - 1$, and this $\forall [i, j]$. For model with environmental direct aquatic transmission, we assume that $0.1 < \omega < 10$, $1 < \theta < 10e4$, $0.1 < \sigma < 10$, $0.1 < \epsilon < 10$. These values have been visually chosen to identify extreme values, from a point of view of disease transmission, to cope with the larger range of parameters' values.

4 Data

| Taxon | Collected | Sampled for MU | PCR Positivity |
|-----------------|------------------|-----------------------|-----------------------|
| Acarina | 708 | 28 | 0.139 |
| Ancylidae | 59 | 4 | 0.000 |
| Anura | 1426 | 31 | 0.114 |
| Atyidae | 524 | 16 | 0.000 |
| Baetidae | 2863 | 44 | 0.051 |
| Belostomatidae | 401 | 37 | 0.091 |
| Caenidae | 617 | 28 | 0.083 |
| Calopterygidae | 2 | 1 | 0.000 |
| Ceratopogonidae | 261 | 28 | 0.000 |
| Chaoboridae | 36 | 7 | 0.000 |
| Chironomidae | 2793 | 57 | 0.058 |
| Cladocera | 769 | 13 | 0.000 |
| Coenagrionidae | 668 | 19 | 0.000 |
| Conchostraca | 6 | 1 | 0.000 |
| Copepoda | 1962 | 13 | 0.000 |
| Corduliidae | 167 | 5 | 0.000 |
| Corixidae | 264 | 15 | 0.000 |
| Culicidae | 1053 | 30 | 0.042 |
| Curculionidae | 11 | 1 | 0.000 |
| Dytiscidae | 358 | 44 | 0.069 |
| Elmidae | 58 | 5 | 0.500 |
| Ephydriidae | 12 | 1 | 0.000 |
| Gerridae | 40 | 17 | 0.000 |
| Gomphidae | 2 | 1 | 0.000 |

| | | | |
|-----------------|------|----|-------|
| Gyrinidae | 25 | 2 | 0.000 |
| Hebridae | 10 | 2 | 0.000 |
| Helicopsychidae | 2 | 1 | 0.000 |
| Heptageniidae | 41 | 4 | 0.000 |
| Hirudinea | 89 | 25 | 0.125 |
| Hydraenidae | 432 | 14 | 0.150 |
| Hydrometridae | 8 | 11 | 0.000 |
| Hydrophilidae | 664 | 58 | 0.093 |
| Hydropsychidae | 12 | 2 | 0.000 |
| Hygrobiiidae | 13 | 1 | 0.000 |
| Lampyridae | 14 | 5 | 0.000 |
| Leptoceridae | 24 | 6 | 0.000 |
| Leptophlebiidae | 15 | 3 | 0.000 |
| Libellulidae | 238 | 26 | 0.118 |
| Lymnaeidae | 12 | 2 | 0.000 |
| Mesoveliidae | 160 | 34 | 0.000 |
| Mysidae | 2 | 1 | 0.000 |
| Naucoridae | 55 | 18 | 0.143 |
| Nematoda | 17 | 3 | 0.000 |
| Nepidae | 34 | 21 | 0.139 |
| Noteridae | 334 | 38 | 0.029 |
| Notonectidae | 387 | 37 | 0.095 |
| Oligochaeta | 1264 | 25 | 0.075 |
| Ostracoda | 1081 | 14 | 0.133 |
| Physidae | 226 | 9 | 0.048 |
| Planorbidae | 525 | 34 | 0.069 |

| | | | |
|-------------------|-----|----|-------|
| Pleidae | 683 | 24 | 0.000 |
| Polycentropedidae | 3 | 1 | 0.000 |
| Polymitarcyidae | 11 | 4 | 0.000 |
| Protoneuridae | 372 | 24 | 0.177 |
| Psephenidae | 1 | 1 | 0.000 |
| Psychodidae | 27 | 1 | 1.000 |
| Pyralidae | 32 | 3 | 0.667 |
| Saldidae | 3 | 2 | 0.000 |
| Sciomycidae | 3 | 3 | 0.333 |
| Scirtidae | 96 | 10 | 0.071 |
| Sphaeriidae | 8 | 1 | 1.000 |
| Stratiomyidae | 24 | 7 | 0.000 |
| Syrphidae | 4 | 2 | 0.500 |
| Tabanidae | 14 | 3 | 0.000 |
| Thiaridae | 60 | 8 | 0.000 |
| Tipulidae | 21 | 6 | 0.000 |
| Veliidae | 43 | 14 | 0.000 |
| Viviparidae | 3 | 3 | 0.000 |

Table 1: Data summary for *Mycobacterium ulcerans* prevalence among aquatic invertebrate taxa across 27 localities of Ghana, West Africa.

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