

1 **Supplement for:**

2 **Climatic warming and the future of bison as grazers**

3 **Joseph M. Craine, E. Gene Towne, Mary Miller, Noah Fierer**

4

5

## 6 Supplemental methods

7 The barcodes used were 12-bp error-correcting barcodes unique to each sample<sup>1</sup>. The  
8 thermocycling program used an initial step at 94 °C for 2 minutes, a final extension at  
9 72 °C for 2 minutes and the following steps cycled 35 times: 2 minutes at 94 °C, 1 minute  
10 at 55 °C, and 30 seconds at 72 °C. Sequences were demultiplexed using a python script  
11 available from: [https://github.com/leffj/helper-code-for-](https://github.com/leffj/helper-code-for-uparse/blob/master/parse_fastq_for_uparse_paired.py)  
12 [uparse/blob/master/parse\\_fastq\\_for\\_uparse\\_paired.py](https://github.com/leffj/helper-code-for-uparse/blob/master/parse_fastq_for_uparse_paired.py) Paired end reads were then merged  
13 using `fastq_merge_pairs`<sup>2</sup>. Since merged reads often extended beyond the amplicon region  
14 of the sequencing construct, we used `fastx_clipper` to trim primer and adaptor regions  
15 from both ends ([https://github.com/agordon/fastx\\_toolkit](https://github.com/agordon/fastx_toolkit)). Sequences lacking a primer  
16 region on both ends of the merged reads were discarded.

17 Sequences were quality trimmed to have a maximum expected number of errors per read  
18 of less than 0.1 and only sequences with more than 3 identical replicates were included in  
19 downstream analyses. BLASTN 2.2.30+ was run locally, with a representative sequence  
20 for each OTU as the query and the current NCBI nt nucleotide and taxonomy database as  
21 the reference. The tabular BLAST hit tables for each OTU representative were then  
22 parsed so only hits with > 97% query coverage and identity were kept.

23

24 In the isotopic mixing model, the diet-feces carbon isotopic offset was set to  $-0.9\text{‰}$ <sup>3,4</sup>.  
25 Reference C<sub>3</sub> and C<sub>4</sub>  $\delta^{13}\text{C}$  values were derived from a survey of herbaceous Konza plant  
26 species collected at the time of first flowering<sup>5,6</sup>. These included 309 herbaceous C<sub>3</sub>  
27 species ( $-29.7 \pm 0.08 \text{‰}$ ) and 56 herbaceous C<sub>4</sub> species ( $-13.2 \pm 0.2 \text{‰}$ ). Data on foliar N

28 concentrations (Fig. S1 and *Ceanothus herbaceus* in text) were taken from ref 6. A  
29 conversion of 6.25:1 was used between [CP] and [N]<sup>7</sup>.  
30  
31 For NIRS analysis, after collection, fecal samples were frozen and later dried at 60 °C in a  
32 forced air oven, ground to 1 mm particle size and re-dried at 60 °C prior to scanning with  
33 near infrared spectroscopy (NIRS) by Texas A&M's Grazingland Animal Nutrition  
34 Lab<sup>8,9</sup>. Spectra (400–2500 nm) were collected on a Foss NIRS 6500 scanning  
35 monochromator with spinning cup attachment. Reference chemistry and chemometrics  
36 for NIRS calibration development that link forage chemistry and fecal spectra were  
37 generated for cattle<sup>10</sup>.

38

39

40

41

## 42 **Supplemental discussion**

### 43 **OTU Richness**

44 The richness of OTUs in diet for bison in the Kansas and South Dakota grasslands are  
45 similar. In the Kansas grassland, an average of  $14.1 \pm 0.5\%$  distinct OTUs per bison fecal  
46 sample were recovered with a total of 65 OTUs detected across all samples per year  
47 (minimum 1% relative read abundance). At the South Dakota grassland, bison consumed  
48 a similar richness of species in their diet as in Kansas—an average of 13.8 OTUs per  
49 sample from DOY 94-281 vs. 14.0 OTUs for the same period ( $P > 0.85$ ). During this  
50 time period, as the season progressed, dietary richness increased ( $DR = 9.49 +$   
51  $0.025 * DOY$ ,  $r^2 = 0.09$ ,  $P = 0.0026$ ) with no significant difference between sexes or sites  
52 in the relationship ( $P > 0.05$ ).

### 53 **Biomass vs. protein intake**

54 The *trnL* sequence examined here is a chloroplast intron with no known variation in copy  
55 number<sup>11</sup>. Because chloroplast density scales with foliar nitrogen concentrations and  
56 therefore protein concentrations<sup>7,12</sup>, the percentage of *trnL* sequences in fecal material  
57 should scale with the percentage of protein ingested from that plant species. That  
58 metabarcoding provides information on protein rather than C or biomass intake is not  
59 necessarily a drawback given that many herbivores are more limited by protein than  
60 energy<sup>13</sup>, allowing for resolution of relative importance of different plant species for a  
61 limiting nutrient to herbivores.

62 A feeding trial has indicated a strong 1:1 relationship between biomass intake and RRA<sup>14</sup>,  
63 but this could be explained by the two plant species used in the feeding trial having  
64 similar protein concentrations. The relatively greater proportion of C<sub>4</sub> species in bison  
65 diet here suggested by isotopic analysis as opposed to metabarcoding likely reflects the  
66 lower N concentrations of C<sub>4</sub> grasses relative to other species (Fig. S1). The slope of the  
67 relationship between the proportion of C<sub>4</sub> plants in bison diet estimated with isotopes and  
68 C<sub>4</sub> grass RRA indicates that the underrepresentation was greatest when C<sub>4</sub> grass  
69 component of diet was lowest (Fig. 4), which is probably the time of the greatest  
70 difference in protein concentrations between C<sub>4</sub> grasses and eudicots.

71

72

73

74 **Forage quality differences between sites**

75 Forage quality is primarily defined by the concentrations of crude protein (CP) and  
76 digestible organic matter (DOM) in plant matter consumed by grazers<sup>7</sup>. When the ratio  
77 of concentrations of DOM to CP is relatively high (e.g. > 4), weight gain is determined  
78 more by the concentration of CP in forage than DOM<sup>15</sup>. Typically, CP is more limiting to  
79 weight gain in North American grazers. In a synthesis of continental-scale patterns of  
80 forage quality, dietary DOM:CP for cattle on pastures and rangelands was generally high  
81 indicating greater protein-limitation than energy-limitation<sup>13</sup>. Geographic patterns in  
82 forage quality characteristics also demonstrated that protein limitation became greater as  
83 MAT increased as peak spring CP concentrations decreased with increasing  
84 temperatures<sup>13</sup>.

85 In 2003, we had measured seasonal patterns of [CP] at the South Dakota grassland and  
86 found that [CP] had peaked at 182 mg g<sup>-1</sup> in May. This is 48 mg g<sup>-1</sup> greater than peak  
87 [CP] for the Kansas grassland (ref 16 and this study). Concomitant with the greater  
88 measured forage quality, bison weights were also higher in the South Dakota than the  
89 Kansas grassland. For example, bison calf weight is a good index of the forage quality  
90 available to bison in a particular year. From 2004-2008—the first 5 years weights were  
91 measured at the South Dakota grassland—calf weights averaged 199 kg (Fig. S2). By  
92 comparison, the long-term average calf weight at the Kansas grassland was 134 kg<sup>16</sup>.  
93 Reproduction rates for adult females were also higher at the South Dakota than the  
94 Kansas grassland (87% vs. 65%, respectively).

95 Recently, it appears that forage quality has declined at the South Dakota site. When  
96 forage quality was measured on the 2014 South Dakota fecals, the highest [CP] measured  
97 was only 100 mg g<sup>-1</sup> (Fig. S3). Part of the relatively low peak [CP] might be influenced  
98 by the relatively low frequency of sampling which could have missed peak forage quality.  
99 Yet, 2014 South Dakota [CP] was lower throughout the year than in 2003 and were now  
100 similar to the Kansas grassland. [DOM] concentrations were also lower than in 2003, but  
101 still higher than the for the Kansas grassland. Examining the weights of bison from 2010-  
102 2013 in the SD grassland (bison weights were not measured in 2009 or 2014), bison calf  
103 weights were lower than 2004-2008, too, declining by an average of 26 kg (e.g. calf  
104 weights were 199.8 vs. 173.2 kg, 2010-2013 vs. 2004-2008 respectively;  $P = 0.005$ ),  
105 though still higher than the long-term average calf weights for the KS grassland.

106 Comparing the two periods, there was little difference in monthly precipitation patterns  
107 or mean temperatures that could explain the decline in forage quality and bison weights  
108 in the SD grassland over the ten-year period. However, budget reductions had forced a  
109 cessation of prescribed fires in 2007. No prescribed fires have been set since then. Late-  
110 spring prescribed fires had been implemented in 2004, 2005, and 2006 and there was a  
111 unprescribed fire in the spring of 2008. In these 4 years, an average of approximately 140  
112 ha was burned each year, which is approximately 10% of the area available to the bison.  
113 Spring fires increase the protein concentrations of grass available to grazers by removing  
114 senesced material and increasing the protein concentrations of the new regrowth<sup>17,18</sup>.

115 In all it appears that when fire was removed from the South Dakota grassland, forage  
116 quality and weight gain declined. Cool-season grasses such as *Poa pratensis* and *Bromus*  
117 *inermis* have also increased in abundance (M.M., personal observation), likely in

118 response to the removal of May fires, but these species appear to be little utilized by  
119 bison. Although bison weight gain and reproduction is still greater at the South Dakota  
120 than the Kansas grassland, bison could not compensate for the decline in grass quality  
121 that accompanied the cessation of burning by increasing their consumption of high-  
122 protein forbs. Although there are still questions regarding the diet of bison on climatically  
123 cool grasslands that have recently experienced fire, the decline in bison performance with  
124 cessation of fire further evince the importance of forage quality and landscape-level  
125 community composition in limiting the resilience of herbivores to changes in grasslands  
126 that are likely to accompany climatic warming.

127

128

129

130



131

132

133 1 Caporaso, J. G. *et al.* Ultra-high-throughput microbial community analysis on the  
134 Illumina HiSeq and MiSeq platforms. *The ISME journal* **6**, 1621-1624, (2012).

135 2 Edgar, R. C. Search and clustering orders of magnitude faster than BLAST.  
136 *Bioinformatics* **26**, 2460-2461, (2010).

137 3 Sponheimer, M. *et al.* An experimental study of carbon-isotope fractionation  
138 between diet, hair, and feces of mammalian herbivores. *Can. J. Zool.* **81**, 871-876,  
139 (2003).

140 4 Codron, D., Codron, J., Lee-Thorp, J. A., Sponheimer, M. & de Ruiter, D. Animal  
141 diets in the Waterberg based on stable isotopic composition of faeces. *S. Afr. J.*  
142 *Wildl. Res.* **35**, 43-52, (2005).

143 5 Craine, J. M., Towne, E. G., Ocheltree, T. W. & Nippert, J. B. Community  
144 traitscape of foliar nitrogen isotopes reveals N availability patterns in a tallgrass  
145 prairie. *Plant Soil* **356**, 395-403, (2012).

146 6 Craine, J. M., Wolkovich, E. M., Towne, E. G. & Kembel, S. W. Flowering  
147 phenology as a functional trait in a tallgrass prairie. *New Phytol.* **193**, 673-682,  
148 (2012).

149 7 Van Soest, P. J. *Nutritional Ecology of the Ruminant*. 2nd edn, (O & B Books,  
150 1994).

151 8 Lyons, R. K. & Stuth, J. W. Fecal NIRS equations for predicting diet quality of  
152 free-ranging cattle. *J. Range Manag.* **45**, 238-244, (1992).

- 153 9 Roberts, C. A., Stuth, J. & Flinn, P. in *Near Infrared Spectroscopy in Agriculture*  
154 (eds C A Roberts, J Workman, & J B Reeves) 231-268 (American Society of  
155 Agronomy-Crop Science Society of America-Soil Science Society of America,  
156 2004).
- 157 10 Showers, S. E., Tolleson, D. R., Stuth, J. W., Kroll, J. C. & Koerth, B. H.  
158 Predicting diet quality of white-tailed deer via NIRS fecal profiling. *Rangeland*  
159 *Ecology and Management* **59**, 300-307, (2006).
- 160 11 Taberlet, P., Gielly, L., Pautou, G. & Bouvet, J. Universal primers for  
161 amplification of three non-coding regions of chloroplast DNA. *Plant Mol. Biol.*  
162 **17**, 1105-1109, (1991).
- 163 12 Evans, J. R. & Seeman, J. R. in *Photosynthesis* (ed W R Briggs) 183-205 (Liss,  
164 1989).
- 165 13 Craine, J. M., Elmore, A. J., Olson, K. C. & Tolleson, D. Climate change and  
166 nutritional stress in cattle. *Global Change Biology* **16**, 2901-2911, (2010).
- 167 14 Willerslev, E. *et al.* Fifty thousand years of Arctic vegetation and megafaunal diet.  
168 *Nature* **506**, 47-51, (2014).
- 169 15 Moore, J. E., Brant, M. H., Kunkle, W. E. & Hopkins, D. I. Effects of  
170 supplementation on voluntary forage intake, diet digestibility, and animal  
171 performance. *J. Anim. Sci.* **77**, 122-135, (1999).
- 172 16 Craine, J. M., Towne, E. G., Tolleson, D. & Nippert, J. B. Precipitation timing  
173 and grazer performance in a tallgrass prairie. *Oikos* **122**, 191-198, (2013).

174 17 Allred, B. W., Fuhlendorf, S. D., Engle, D. M. & Elmore, R. D. Ungulate  
175 preference for burned patches reveals strength of fire-grazing interaction. *Ecol*  
176 *Evol* **1**, 132-144, (2011).

177 18 Van de Vijver, C., Poot, P. & Prins, H. H. T. Causes of increased nutrient  
178 concentrations in post-fire regrowth in an East African savanna. *Plant Soil* **214**,  
179 173-185, (1999).

180

181

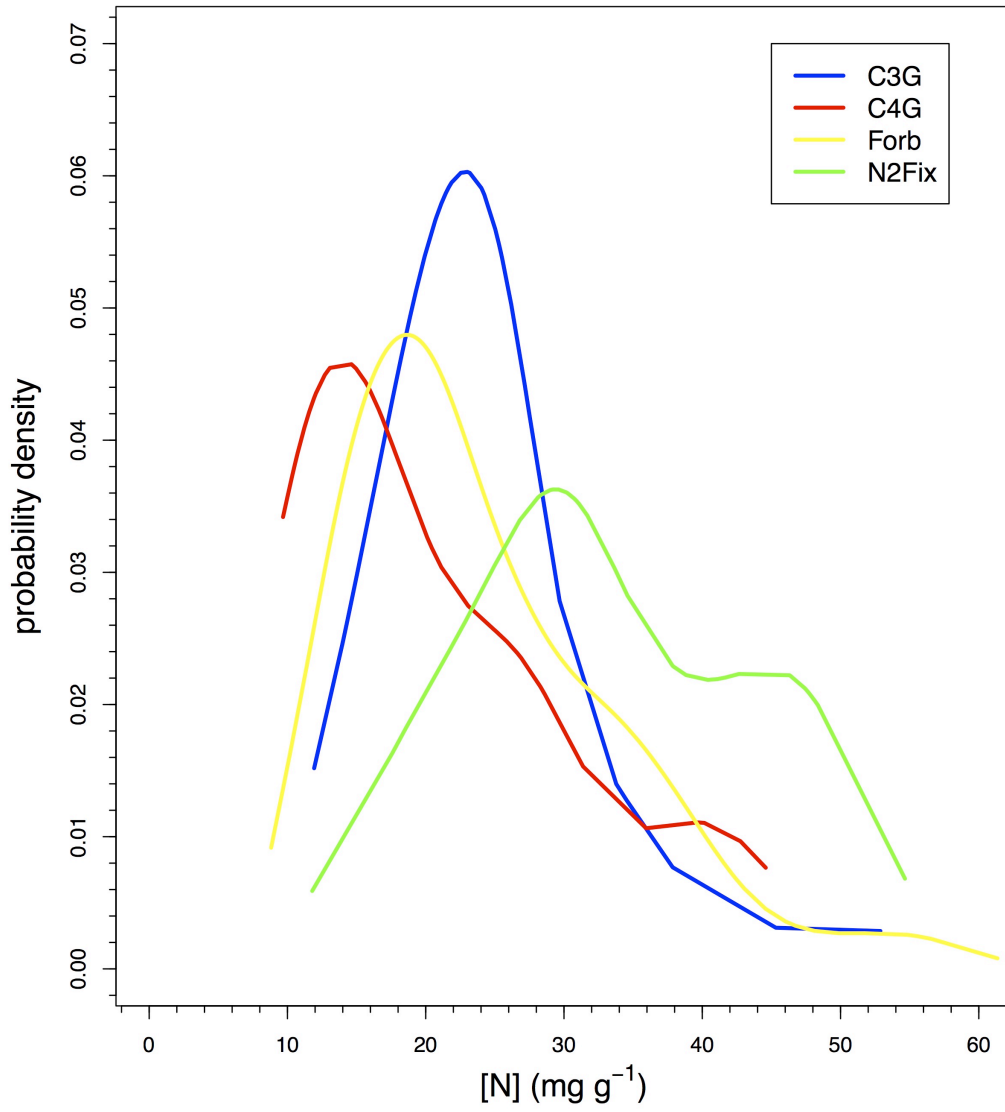
182 **Figure S1| Frequency among plant species of foliar N concentrations ([N]) in Kansas**  
183 **grassland by functional group.** Functional groups include C<sub>3</sub> grass [C3G], C<sub>4</sub> grass  
184 [C4G], non-N<sub>2</sub>-fixing herbaceous eudicots [Forb], and N<sub>2</sub>-fixing eudicots [N<sub>2</sub>-fix]. Foliar  
185 [N] for each species were measured on first day of flowering.

186 **Figure S2| Fall weights of male and female bison calves over time at South Dakota**  
187 **Prairie.**

188 **Figure S3| Crude protein (CP) and digestible organic matter (DOM) concentrations**  
189 **for the Kansas grassland (2011, 2012, 2013) and the South Dakota grassland (2003,**  
190 **2014).**

191

192 Figure S1.



193

194

