## **Electronic supplementary material for:**

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## **Appendix S1: Supplementary methods**

*Estimation of giant kelp fecundity.* We estimated giant kelp patch fecundity from canopy biomass by calibrating a relationship between the density of spore-bearing tissue (measured as sorus area per unit area of reef) and Landsat estimates of canopy biomass density at the San Clemente Artificial Reef (SCAR), located offshore of San Clemente, California, USA. Briefly, from March to June in each year from 2000 to 2004 at 7 artificial reef sites at SCAR, divers measured the mean sorus area of sporophytes (mean number of plants measured per site per year  $= 276 \pm 230$  [SD]; range = 10-933; data and detailed methods in ref. 1). We estimated canopy biomass density using Landsat 5 Thematic Mapper and Landsat 7 Enhanced Thematic Mapper Plus imagery collected from the same locations and similar dates (the average of all observations within one month before and after the diver measurement date). We estimated the relationship between pixel fecundity ( $cm^2$  sorus per  $m^2$ ) and the square root of pixel canopy biomass density (kg wet per m<sup>2</sup>) in the first semester of the year (January–June) using a zero intercept regression as fecundity density =  $1463 \times \sqrt{canopy}$  biomass density (electronic supplementary material, figure 2; 95% CI of model coefficient = 1234–1865). To accommodate the possibility of spore production in pixels lacking canopy biomass (due to measurement error or the presence of subsurface adult kelp), we also fit a linear regression with a nonzero intercept as *fecundity* 

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*density* =  $12.14 + 1596 \times \sqrt{canopy biomass density}$ . We fit these relationships in R 3.2.5 [2] using generalized least squares (GLS) linear regression with heterogeneous, first-order autoregressive covariance structure because ordinary linear regression residuals showed serial autocorrelation and heteroscedasticity [3,4]. Sample autocorrelation function analysis showed no evidence for temporal autocorrelation of GLS standardized residuals [3–5]. Semi-variograms and bubble plots showed no evidence of spatial autocorrelation of GLS standardized residuals [3,4,6].

Data to parameterize the relationship between canopy biomass density and fecundity were available at SCAR from the first semester of the year only. To adjust fecundity estimates for seasonality in biomass-specific fecundity, we used field data collected by Reed et al. [7]. Briefly, divers sampled giant kelp monthly from November 1990 to September 1994 at three sites in the Santa Barbara Channel, California, USA (7–14 m depth). During each sampling, divers measured the fecundity (i.e., sorus area per individual) and number of fronds of the first 30 mature giant kelp sporophytes (with  $\geq$  8 fronds and holdfast diameter  $\geq$  30 cm) encountered along each transect. We estimated biomass-specific fecundity as sorus area per frond because the density of fronds from mature sporophytes is a strong predictor of canopy biomass density [8]. Our analysis revealed that average giant kelp biomass-specific fecundity is 2.4 (95% CI = 1.73 - 1.73)2.96) times greater during the first half of the year than during the second half of the year (Welch's unequal variances *t*-test;  $t_{38} = 4.42$ , P < 0.001; January–June = 119.2 ± 84.9 [SD] cm<sup>2</sup> per frond; July–December =  $50.1 \pm 29.7$  cm<sup>2</sup> per frond). Therefore, to estimate the relationship between patch fecundity and canopy biomass density in the second semester of the year, we adjusted the relationship from the first semester by dividing the model coefficient by 2.4 (hence, for July–December, fecundity density =  $609.6 \times \sqrt{canopy}$  biomass density).

*Estimations of giant kelp spore dispersal.* Because the giant kelp patches in the study region are smaller than the ROMS connectivity cells (nominal alongshore distance = 8 km; electronic supplementary material, figure 1), minimum transport times between any two giant kelp patches cannot be calculated without making assumptions. Thus, for cases where two giant kelp patches were located in different ROMS cells, we assumed that for each semester, *t*, the transport time from kelp patch *i* to patch *j* ( $d_{ij,t}$ ) was proportional to the minimum transport time between the two ROMS cells *x* and *y* ( $D_{xy,t}$ ), respectively:

$$\frac{D_{xy,t}}{L_{xy}} = \frac{d_{ij,t}}{l_{ij}}$$

where  $L_{xy}$  is the alongshore distance between the centroids of ROMS cells x and y, and  $l_{ij}$  is the alongshore distance between the centroids of giant kelp patches *i* and *j*. For cases where two giant kelp patches were located within the same ROMS cell (y), we assumed that  $d_{ij,t}$  could be estimated using the minimum transport times between the adjacent ROMS cells (x and z):

$$\frac{(D_{xy,t} + D_{yz,t})}{(L_{xy} + L_{yz})} = \frac{d_{ij,t}}{l_{ij}}$$

*Assessment of model assumptions.* Prior to all analyses, we ensured that data conformed to model assumptions. To prevent extrapolation to extreme connectivity values with very few observed cases, we recoded the highest 0.1% of values to the 99.9% value for each connectivity measure. To directly compare regression coefficients of binary and continuous predictors within and among models as a measure of effect size, we centered (i.e., subtracted the mean from) all predictors and standardized continuous predictors by dividing by two standard deviations [9]. To assess data linearity, we examined the relationship between the log odds of patch occupancy and each continuous predictor using locally weighted scatterplot smoothing (bandwidth = 0.6-0.9; ref. 10). To determine multicollinearity, we measured the variance inflation factor (VIF =  $1/[1-R^2]$ ) among predictors within each model [3]; in all cases multicollinearity was very low (VIF <

1.27). We examined Pearson residuals for temporal and spatial autocorrelation using sample autocorrelation function analysis and spline correlograms (with bootstrapped 95% confidence intervals; n = 1000 randomizations), respectively. Temporal or spatial autocorrelation was not detected in any models (electronic supplementary material, figures S3 and S4).

## References

- 1. Reed, D. C., Schroeter, S. C. & Huang, D. J. S. 2006 An experimental investigation of the use of artificial reefs to mitigate the loss of giant kelp forest habitat: a case study of the San Onofre Nuclear Generating Station's artificial reef project. San Diego, California, USA: Sea Grant Extension Program, UC Cooperative Extension.
- 2. R Core Team 2016 R: A language and environment for statistical computing. R Foundation for Statistical Computing, Vienna, Austria. http://www.R-project.org/.
- 3. Zuur, A. F., Ieno, E. N., Walker, N. J., Saveliev, A. A. & Smith, G. M. 2009 *Mixed Effects Models and Extensions in Ecology with R.* New York, New York, USA: Springer.
- 4. Zuur, A. F. & Ieno, E. N. 2016 A protocol for conducting and presenting results of regression-type analyses. *Methods Ecol. Evol.* **7**, 636–645. (doi:10.1111/2041-210X.12577)
- 5. Diggle, P. J., Heagerty, P., Liang, K.-Y. & Zeger, S. L. 2013 *Analysis of Longitudinal Data*. Second ed. New York, New York, USA: Oxford University Press.
- Dormann, C. F. et al. 2007 Methods to account for spatial autocorrelation in the analysis of species distributional data: a review. *Ecography* 30, 609–628. (doi:10.1111/j.2007.0906-7590.05171.x)
- 7. Reed, D. C., Ebeling, A. W., Anderson, T. W. & Anghera, M. 1996 Differential reproductive responses to fluctuating resources in two seaweeds with different reproductive strategies. *Ecology* **77**, 300–316. (doi:10.2307/2265679)
- 8. Rassweiler, A., Arkema, K. K., Reed, D. C., Zimmerman, R. C. & Brzezinski, M. A. 2008 Net primary production, growth, and standing crop of *Macrocystis pyrifera* in southern California. *Ecology* **89**, 2068. (doi:10.1890/07-1109.1)
- 9. Gelman, A. 2008 Scaling regression inputs by dividing by two standard deviations. *Stat. Med.* **27**, 2865–2873. (doi:10.1002/sim)
- 10. Hosmer, D. W., Lemeshow, S. & Sturdivant, R. X. 2013 *Applied Logistic Regression*. Third ed. New York, New York, USA: John Wiley & Sons.