Supporting Information

Constrained Mixed-Effect Models with Ensemble Learning for Prediction of Nitrogen Oxide Concentrations at a High Spatiotemporal Resolution

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1. Overview

 This paper proposed a three-stage spatiotemporal model that can reliably predict nitrogen oxide concentrations with a high spatiotemporal resolution over a long time span (>20 years). The spatially extensive highly-clustered exposure data from short-term measurement campaigns across 1-2 years and long-term central site monitoring in 1992-2013 were leveraged to develop the first stage mixed-effect models and the second stage ensemble learning with uncertainty estimates. Then at the third stage, constrained optimization was designed and implemented based on the point estimates from the first and second stages to simulate the long-term series of pollutant concentrations for any target location in the study region. The following sections provide the supplemental information about concentration measurements, the covariates selected, modeling method, and results to support the formal paper.

2. Measurements of NO² and NO^x concentrations

 Besides the measurements of routine monitoring stations, additional data were generated in intensive field measurement campaigns conducted by the University of Southern California (USC), University of California Los Angeles (UCLA), and University of California Irvine (UCI), respectively. Passive diffusion-based Ogawa 18 samplers ¹ were used to measure NO_2 and NO_x at different time periods and at different locations (e.g. outside homes, schools, strategic outdoor sampling locations, and central monitoring sites). Our previous papers respectively provide more details 21 about the measurement methods that generated the USC², UCLA³ and UCI⁴ samples. Figure S1 also shows the locations for the routine and USC sampling sites [the UCI and UCLA sampling locations are concealed to comply with specific requirements by their Institutional Review Boards].

 To minimize systematic bias in the field campaign data, we compared the passive data with the active data from the routine government monitors at the co-located sites and made small adjustments to standardize the passive measurements to Federal 28 Reference Method equivalent values. For details, please refer to Table S1.

3. The Covariates Selected

3.1 CALINE4-estimated concentrations from local traffic emissions

CALINE4 is a line source dispersion model that was used to assess the

32 contribution of local motor vehicle emissions to ambient concentrations $5, 6$. 33 CALINE4 was used to compute mean NO_x concentration from emissions on freeways [coded using Feature Class Codes (FCC) as FCC1] and non-freeways (FCC2, FCC3 and FCC4). Traffic count data were obtained from Caltrans and TeleAtlas/GDT and assigned to ERSI Premium Street Map roadway geometry. Emission strength was estimated using quarterly average daily traffic volumes and EMFAC2011 (for 38 1992-2012)⁷ and EMFAC2014 (for 2013)⁸, which generated air basin emission factors that were based on average vehicle speed and heavy-duty truck fraction (Caltrans post-mile truck count data by year). Wind speeds and directions were based on hourly observations of these surface meteorological variables from 72 42 monitoring stations of California⁸.

3.2 Traffic density

 Traffic density represents distance-decayed annual average daily traffic (AADT) volume in both directions from all roads (FCC1-FCC4) within a circular buffer. Traffic density is symmetric on both sides of each roadway, computed as if the wind directions were uniformly distributed around the compass. The values of traffic density were computed by the ESRI ArcGIS density function using a kernel with a 300 m search radius and 5 m grid resolution. Annual traffic density estimates were provided for the regulatory monitoring sites and local sampling locations. In addition, because these cover a long time period, the traffic densities were scaled by 52 the South Coast Air Basin (SoCAB) EMFAC2011 vehicle fleet average NO_x emission factor for 50 mph and 6% heavy-duty vehicle fraction (normalized to 1.00 in 2002 since we used the 2002 AADT as the baseline data) to reflect the composite trend in traffic volumes and emissions over time.

4. Modeling Approach

4.1 Non-parametric additive methods

 For spatiotemporal factors, we adopted non-parametric additive methods to model non-linear effects. Specifically, we used non-parametric trend functions to quantify the association, *s*(*…*), i.e. approximated by the weighted sum of polynomial spline (B-spline basis) functions

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$$
x_{min} = \zeta_0 < \zeta_1 < \zeta_2 \dots < \zeta_{m-1} < \zeta_m = x_{max} \tag{S1}
$$

$$
62
$$

63
$$
f(x_i) = \sum_j \beta_{ij} B_{ij}(x_i)
$$
 (S2)

64 where ζ_i is the split for an interval for the covariate x_i (ζ_0 and ζ_m are respectively the 65 minimum and maximum values of x_i), B_{ij} and β_{ij} respectively represent the basis 66 function and parameter for the interval *j* for the covariate x_i . Penalized maximum

67 likelihood was used to solve (eq. S2) to estimate β_{ij} .

68 **4.2 Modeling of spatial random effects**

 By estimating a structured component and an unstructured component, we can 70 distinguish between the two sources of spatial autocorrelations⁹. The spatial effects were modeled using the following formulas for both structured (S3) and unstructured spatial effects (S4).

$$
^{73}
$$

73
$$
f_{s}(r_{s})|f_{s}(r'), r' \neq r_{s}, \tau_{str}^{2} \sim N(\frac{1}{N_{r_{s}}} \sum_{r' \in \delta_{r_{s}}} f_{s}(r'), \frac{\tau_{str}^{2}}{N_{r_{s}}})
$$
(S3)

$$
f_{re}(r_s)|\tau_{unstr}^2 \sim N(0, \tau_{unstr}^2)
$$
\n
$$
\tag{S4}
$$

75 where r_s is the region where the observation $y(s,t)$ is located, δ_{r_s} represents a set of 76 neighbors (r') of the polygon r_s , N_{r_s} is the number of neighboring polygons for r_s , 77 τ_{str}^2 is the total variance for the structured component, $\tau_{str}^2 \sim IG(a,b)$. $f_s(r')$ in (eq. 78 S3) represents the spatial influence from neighboring polygons (*r'*) on r_s ; $f_{re}(r_s)$ in (eq. 79 S4) represents the unstructured spatial effect with zero mean and standard deviation 80 (τ_{unstr}^2) for r_s .

 Thiessen polygons were constructed around the central points (derived by averaging the coordinates of all the routine or/and campaigns sampling locations within a certain distance) to simulate spatial effects. Rook adjacency was used for spatial adjacency: two polygons were assumed to be neighbors if they share a common border. We conducted sensitivity tests for a series of aggregation distances (100 m, 300 m, 500 m and 3 km) and finally selected an optimal aggregate distance (500 m) that provided a good balance between model accuracy and computing efficiency. We used the packages of rgdal and spdep in the statistics software R (Version 3.3) for generation of the Thiessen polygons with their spatial weight matrix.

90 **4.3 Aggregated predictions by ensemble learning**

91 The aggregated predictions (mean and standard deviation) are the weighted 92 summary of all trained models, where the weighting is the square of each model's R^2 .

$$
m_f(s,t) = \sum_i h_i(d_b, f_r) w_i \tag{S5}
$$

94
$$
\sigma_f(s,t) = \sqrt{\frac{\sum_i w_i (h_i(d_b, f_r) - m_f(s,t))^2}{\frac{M-1}{M} \sum_i w_i}}
$$
(S6)

95
$$
w_i = R_i^{2^2} / \sum_i R_i^{2^2}
$$
 (S7)

96 where $m_f(s, t)$ is the aggregated prediction (the weighted mean), $h_i(d_b, f_r)$ is the 97 prediction by the ith spatiotemporal model (eq. 1) trained using the bootstrap sample 98 (*d_b*) and selected set of predictors (*f_r*); *w_i* is the normalized weight derived from the *i*th

S3

- 99 model's performance measure; $\sigma_f(s,t)$ is the standard deviation from the output of
- 100 multiple models, *M* is the number of nonzero weights.

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^a. ICV: The Intra-Community Variability study

Table S2. Variance explained by the predictors included in the Stage 1 mixed effects

model

^a: threshold defined to remove the outliers for the covariate;

^{*b*}: bold font highlights the variance explained \geq 10% by the variable;

c : Feature Class Codes for freeways and highways

| ICV 1^a | Samples | Mean (ppb) | Correlation | RMSE ^b | $\ensuremath{\mathsf{NRMSE}^c}\xspace$ | CVRMSE ^d |
|-------------------|---------|------------|-------------|-------------------|--|---------------------|
| Alpine | 156 | 8.71 | 0.89 | 2.51 | 0.1 | 0.29 |
| Anaheim | 128 | 30.07 | 0.69 | 4.02 | 0.13 | 0.13 |
| Glendora | 221 | 20.84 | 0.88 | 3.05 | 0.09 | 0.15 |
| Lake Arrowhead | 121 | 8.93 | 0.61 | 2.36 | 0.15 | 0.28 |
| Lake Elsinore | 165 | 10.93 | 0.92 | 1.46 | 0.08 | 0.13 |
| Long Beach | 162 | 20.2 | 0.98 | 2.59 | 0.07 | 0.13 |
| Mira Loma | 189 | 13.66 | 0.98 | 1.46 | 0.06 | 0.11 |
| Riverside | 215 | 16.55 | 0.93 | 2.84 | 0.1 | 0.17 |
| San Bernardino | 138 | 15.17 | 0.97 | 2.13 | 0.08 | 0.14 |
| San Dimas | 174 | 25.17 | 0.9 | 2.7 | 0.08 | 0.11 |
| Santa Barbara | 147 | 11.29 | 0.93 | 2.34 | 0.1 | 0.21 |
| Santa Maria | 153 | 8.39 | 0.64 | 1.02 | 0.14 | 0.12 |
| Upland | 232 | 20.14 | 0.95 | 2.73 | 0.08 | 0.14 |
| | | | | | | |
| $ICV2^a$ | Samples | Mean (ppb) | Correlation | RMSE ^b | $\ensuremath{\mathsf{NRMSE}^c}\xspace$ | CVRMSE ^d |
| Anaheim | 26 | 19.56 | 0.77 | 3.32 | 0.15 | 0.17 |
| Glendora | 30 | 17.03 | 0.75 | 2.45 | 0.19 | 0.14 |
| Long Beach | $27\,$ | 20.42 | 0.92 | 3.15 | 0.13 | 0.15 |
| Mira Loma | 26 | 18.24 | 0.98 | 1.68 | 0.08 | 0.09 |
| Riverside | 28 | 13.82 | 0.51 | 2.35 | 0.28 | 0.17 |
| San Bernardino | 40 | 11.07 | 0.71 | 2.41 | 0.16 | 0.22 |
| San Dimas | $28\,$ | 20.16 | 0.83 | 2.66 | 0.19 | 0.13 |

Table S3. NO₂ model performance by CHS community in cross validation

^a: ICV, The Intra-Community Variability study; ^{*b*}: RMSE, root mean square error; ^{*c*}: RMSE, root mean square error; NRMSE, normalized RMSE; *^d* :CV RMSE, coefficient of variation of the RMSE.

Upland 28 15.79 0.75 2.56 0.18 0.16

l,

| ICV 1^a | Samples | Mean (ppb) | Correlation | \mathbf{RMSE}^b | NRMSE ^c | CVRMSE ^d |
|-------------------|---------|------------|-------------|-------------------|--|----------------------------------|
| Alpine | 156 | 18.8 | 0.9 | 4.29 | 0.08 | 0.23 |
| Anaheim | 128 | 67.25 | 0.95 | 6.59 | 0.08 | 0.1 |
| Glendora | 221 | 41.82 | 0.9 | 5.34 | 0.07 | 0.13 |
| Lake Arrowhead | 121 | 16.12 | 0.68 | 3.7 | 0.13 | 0.23 |
| Lake Elsinore | 165 | 20.34 | 0.91 | $2.5\,$ | 0.08 | 0.12 |
| Long Beach | 162 | 63.01 | 0.98 | 8.78 | 0.07 | 0.14 |
| Mira Loma | 189 | 31.64 | 0.91 | 3.17 | 0.07 | 0.1 |
| Riverside | 215 | 35.86 | 0.91 | 5.99 | 0.08 | 0.17 |
| San Bernardino | 138 | 39.42 | 0.91 | 4.74 | 0.07 | 0.12 |
| San Dimas | 174 | 52.64 | 0.94 | 5.73 | 0.06 | 0.11 |
| Santa Barbara | 147 | 26.98 | 0.94 | 6.65 | 0.09 | 0.25 |
| Santa Maria | 153 | 14.21 | 0.67 | 1.43 | 0.13 | 0.1 |
| Upland | 232 | 41.37 | 0.93 | 5.83 | 0.06 | 0.14 |
| | | | | | | |
| $ICV2^a$ | Samples | Mean (ppb) | Correlation | RMSE ^b | $\ensuremath{\mathsf{NRMSE}^c}\xspace$ | $\ensuremath{\mathrm{CVRMSE}}^d$ |
| Anaheim | 26 | 33.4 | 0.76 | 8.86 | 0.17 | 0.27 |
| Glendora | 30 | 24.47 | 0.85 | 3.05 | 0.14 | 0.12 |
| Long Beach | 27 | 48.92 | 0.97 | 8.09 | 0.08 | 0.17 |
| Mira Loma | 26 | 33.9 | 0.93 | 6.98 | 0.12 | 0.21 |
| Riverside | $28\,$ | 19.29 | 0.83 | 2.78 | 0.14 | 0.14 |
| San Bernardino | 40 | 17.58 | 0.83 | 4.21 | $0.11\,$ | 0.24 |
| San Dimas | 28 | 29.51 | 0.83 | 4.43 | 0.18 | 0.15 |
| Upland | 28 | 22.59 | 0.89 | 3.05 | 0.12 | 0.14 |

Table S4. NO_x model performance by CHS community in cross validation

^a: ICV, The Intra-Community Variability study; ^{*b*}: RMSE, root mean square error; ^{*c*}: RMSE, root mean square error; NRMSE, normalized RMSE; *^d* :CV RMSE, coefficient of variation of the RMSE.

Figure S1. Study region with routine monitoring locations and USC sampling locations (the UCLA and UCI data are not shown due to IRB restrictions)

b. Histogram of NO_x (left: the original NO_x measurements with a big right skewness of 2.0; right: log-transformation with a small skewness of -0.25)

Figure S2. Histograms for NO₂ and NO_x to determine log-transformation (no log transformation for NO₂; log transformation for NO_x)

Figure S4. Non-linear association between predictive variables and concentrations by mixed models

Figure S5. Spatial topology for spatial effect modeling by Thiessen polygons (aggregate distance: 500 m)

b

Figure S6. Residual plot between observed values vs. residuals (a. NO_2 ; b. NO_x) for the sample selected by bootstrap aggregating

 $b. NO_x$

Figure S7. Spatial distribution of the predicted and observed $NO₂$ and NO_x concentration means across the CHS communities

a. NO_2 b. NO_x Figure S8. Boxplot for correlation between constrained prediction and observed values for the time series of 51 routine monitoring stations

Figure S9. Time series simulated for the routine monitoring sites with the minimum correlation between constrained prediction and observed values (a.0.55 for NO_2 ; 0.70 for NO_x)

Figure S10. Summer (a) and winter (b) averages of the 2005-2006 biweekly NO_x at USC ICV1 sampling locations for San Dimas

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