# Antibiotic Pollution in Marine Food Webs in Laizhou Bay, North

#### **China: Trophodynamics and Human Exposure Implication**

Sisi Liu,<sup>†</sup> Hongxia Zhao,<sup>\*†</sup> Hans-Joachim Lehmler,<sup>§</sup> Xiyun Cai,<sup>†</sup> Jingwen Chen<sup>\*†</sup>

<sup>†</sup> Key Laboratory of Industrial Ecology and Environmental Engineering (Ministry of Education), School of Environmental Science and Technology, Dalian University of Technology, Linggong Road 2, Dalian 116024, China

<sup>§</sup> Department of Occupational and Environmental Health, College of Public Health, The University of Iowa, IA 52242, USA

#### **Corresponding Authors**

\* Hongxia Zhao, phone/fax: +86-411-8470 7965, e-mail: <u>hxzhao@dlut.edu.cn</u>, address: Linggong Road 2, Ganjingzi District, Dalian 116024, China; Jingwen Chen, phone/fax: +86-411-8470 6269, e-mail: <u>jwchen@dlut.edu.cn</u>, address: Linggong Road 2, Ganjingzi District, Dalian 116024, China.

Number of pages: 40

Number of tables: 10

Number of figures: 4

#### **Table of Contents**

Chemicals and Reagents	S4
Sample Extraction and Cleanup	S5
Stable Isotope Analysis	S6
Estimation of Trophic Magnification Factors ( <i>TMF</i> )	S7
Determination of Relative Carbon Sources	<b>S</b> 8
Determination of Estimated Daily Intakes (EDI)	S9
Determination of Hazard Quotients ( <i>HQ</i> )	10

## List of Tables

 
 Table S1. Physicochemical properties of antibiotics investigated in this study.
 Table S2. Summary of the average body lengths and body weights of the specimens collected from Laizhou Bay, China, and the corresponding sampling locations and Table S3. Optimized LC-MS/MS parameters and retention times of the antibiotics Table S4. Surrogate standards, linear range of matrix matched calibration curves, method detection limits (MLOD), method quantification limits (MLOQ), spiked average recoveries and relative standard deviations (RSD, n = 6) for each antibiotic.17 **Table S6.** Moisture content (%), estimated trophic level (TL), mean concentrations in ng/g ww (ng/g dw) of sulfonamide, fluoroquinolone, macrolide antibiotics and trimethoprim, and detection frequency (%) in different species collected from Laizhou Table S7. Comparison of antibiotic levels detected in aquatic animals in areas around 

Table S8. Analysis of variance (ANOVA) of the calculated trophic level (TL) values
at different sampling times
Table S9. Slope, correlation coefficients $(r)$ and $p$ -values of regression analyses
between logarithm concentrations and trophic levels (see Figure 3), and TMFs and 95%
confidence intervals (CIs) of target antibiotics and ΣSAs, ΣFQs, ΣMLs
Table S10. 90 <sup>th</sup> percentile antibiotic concentrations (C), estimation daily intake (EDI)
values and associated hazard quotients (HQ) of antibiotics for rural and urban
residents and acceptable daily intake (ADI) values obtained from the literature S28
List of Figures
Figure S1. Sampling locations of this study
<b>Figure S2.</b> $\delta^{15}$ N and adjusted $\delta^{13}$ C values for the species in the coastal food web from
Laizhou Bay, China
Figure S3. Significant difference ( $p < 0.05$ ) on fraction of neutral molecules ( $f_n$ )
values between SA antibiotics (including TMP) and the other two antibiotic groups.

References	5	5	5	5
------------	---	---	---	---

Chemicals and Reagents. Sulfadiazine (SDZ), sulfathiazole (ST), sulfapyridine (SPD), sulfamerazine (SMZ), sulfadimidine (SDM), sulfamethoxazole (SMX), sulfadimethoxine (SM2), sulfamonomethoxine (SMM), sulfachlorpyridazine (SCP), trimethoprim (TMP), enoxacin (ENX), ofloxacin (OFL), ciprofloxacin (CIP), norfloxacin (NOR), enrofloxacin (ENR), roxithromycin (ROX), and anhydroerythromycin (ERY-H<sub>2</sub>O) were purchased from Dr. Ehrenstorfer GmbH (Augsburg, Germany). Azithromycin (AZI) was obtained from Fluorochem Co., Ltd (Derbyshire, UK). Clarithromycin (CLA) were provided by TCI Co., Ltd (Shanghai, isotope-labeled compounds, China). Stable including sulfamethoxazole- $D_4$ , atrazine-D<sub>5</sub>, trimethoprim-D<sub>3</sub>, ciprofloxacin-D<sub>8</sub>, and caffeine- $^{15}N_2$ , were purchased from Dr. Ehrenstorfer GmbH. Stock solutions of 1 mg/mL were prepared in methanol and stored at -4°C for 6 months.

Methanol and acetonitrile were of HPLC grade and provided by Tedia Inc. (Fairfield, Ohio, USA). Ammonium formate (analytical purity) was obtained from Sinopharm Chemical Reagent Co., Ltd. (Beijing, China). Formic acid (analytical purity) was purchased from Kermel Chemical Reagent Co., Ltd. (Tianjin, China). Ammonium hydroxide (guaranteed purity) was provided by Guangfu Fine Chemical Research Institute (Tianjin, China). Diatomite was obtained from Dionex Co., Ltd. (Sunnyvale, California, USA) and combusted at 450 °C for 3 hours. Ultrapure water (18.2 M $\Omega$ ·cm, 25 °C) was obtained with an OKPURE water system (Laikie Instrument Co., Ltd., Shanghai, China).

**Sample Extraction and Cleanup.** Approximately 1 g dry weight of a pooled biota sample was spiked with surrogate standards (10 ng each of sulfamethoxazole-D<sub>4</sub>, atrazine-D<sub>5</sub>, trimethoprim-D<sub>3</sub>, and ciprofloxacin-D<sub>8</sub>) before being thoroughly mixed with 3 g of diatomite. The mixture was placed into a 10 mL stainless steel extraction cell and extracted by pressurized liquid extraction (ASE 350, Dionex, Sunnyvale, California, USA) using methanol as extraction solvent. The extraction conditions were as follows: extraction temperature, 80 °C; extraction pressure, 10.48 MPa; preheating period, 5 min; static extraction, 5 min; flush volume 60% of the cell volume; nitrogen purge, 120 s; and number of extraction cycles, 2.

After extraction, approximately 21.5 mL of extract were transferred to a 50 mL round-bottom flask and concentrated to near dryness at 40 °C using a RE-2000 rotary evaporator (Yarong, Shanghai, China). The extract was diluted immediately with 50 mL of ultrapure water. The Oasis HLB cartridges (60 mg, 3 mL; 30  $\mu$ m) were pre-conditioned with 5 mL of methanol and equilibrated with 2 mL of ultrapure water. After loading the samples, cartridges were washed with 12 mL of ultrapure water to remove matrix interferences and dried under vacuum for 30 min. The analytes and surrogate standards retained on the cartridges were eluted with 12 mL of methanol containing 5% ammonium hydroxide. The eluent was evaporated to near dryness at 40 °C and the residue was redissolved in 500  $\mu$ L of the initial mobile phase (95% phase A: 5% phase B). The internal standard (10 ng of caffeine-<sup>15</sup>N<sub>2</sub>) was added to each sample. Samples were briefly vortexed, and filtered using a 0.22- $\mu$ m syringe filter. Ten  $\mu$ L of extracts were injected for LC/MS/MS analysis.

Stable Isotope Analysis. An aliquot of approximately 1.50 mg of freeze-dried and homogenized samples were combusted in Sn capsules at 1020 °C. NO<sub>x</sub> was reduced at 650 °C, the mixed gas (N<sub>2</sub> and CO<sub>2</sub>) was separated and stable-carbon and stable-nitrogen isotope ratios were determined using an Isoprime 100 isotope ratio mass spectrometer interfaced to a vario PYRO cube elemental analyzer (Elementar, Hanau, Germany). The carbon stable-isotope ratio ( $\delta^{13}$ C) and nitrogen stable-isotope ratio ( $\delta^{15}$ N) of the samples are given in parts per thousand according to the formulas:

$$\delta^{15} N = \left(\frac{{}^{15} N/{}^{14} N_{\text{sample}}}{{}^{15} N/{}^{14} N_{\text{atmosphere}}} - 1\right) \times 1000 \,(\%)$$
(Eq. S1)

and

$$\delta^{13} C = \left(\frac{{}^{13}C/{}^{12}C_{sample}}{{}^{13}C/{}^{12}C_{PDB}} - 1\right) \times 1000\,(\%)$$
(Eq. S2)

where the  ${}^{15}N/{}^{14}N_{amosphere}$  standard value is based on atmospheric nitrogen (air) and the  ${}^{13}C/{}^{12}C_{PDB}$  standard value is based on PDB (Peedee Belemnite carbonate). Triplicate analyses of two international reference materials (IAEA-600 and USGS-43) indicate an accuracy of  $\pm$  0.15‰ and  $\pm$  0.2‰ for stable-carbon and stable-nitrogen isotope measurements, respectively. Estimation of Trophic Magnification Factors (*TMF*). *TMF* values were determined from the slope of a regression between log-transformed antibiotic concentrations (ww) in biota and the *TLs* of the sampled biota (Eq. S3 and Eq. S4). *TMF* value of each antibiotic is shown in Table S9. A *TMF* > 1 indicates biomagnification through the food web or from prey-to-predator; otherwise, trophic dilution is suggested.

$$log[chemical concentration]_{ww} = a + b \cdot TL$$
 (Eq. S3)

$$TMF = 10^b \tag{Eq. S4}$$

**Determination of Relative Carbon Sources.** Like  $\delta^{15}$ N,  $\delta^{13}$ C can be enriched through *TL*s. To eliminate the trophic influence of consumers on  $\delta^{13}$ C values, the relative contribution of benthic *vs.* pelagic carbon source (relative carbon source, Figure 2) was calculated using the following equation:<sup>1</sup>

$$Relative Carbon Source = 1 - \frac{\delta^{13}C_{Acetes chinensis} - \delta^{13}C_{consumer} + \Delta\delta^{13}C(TL_{consumer} - TL_{Acetes chinensis})}{\delta^{13}C_{Acetes chinensis} - \delta^{13}C_{Rapana venosa}}$$
(Eq. S5)

where the northern maoxia shrimp (*Acetes chinensis*) is used as the pelagic source, the veined rapa whelk (*Rapana venosa*) represents the benthic source, and  $\Delta \delta^{13}$ C is a constant trophic enrichment factor of 1.3‰ for consumers.

**Determination of Estimated Daily Intakes** (*EDI*). *EDI* values for the individual antibiotics (Table S10) were calculated using the following equation and assume that 100% of the ingested antibiotics are absorbed:

$$EDI = \frac{PR \times C}{BW}$$
 (Eq. S6)

where *EDI* is the estimated daily intake of an antibiotic in ng/kg body weight (bw)/day; *PR* is the annual per capita fish consumption of 18.63 g/person/day for rural residents and 34.45 g/person/day for urban residents in China;<sup>2</sup> *BW* is the average body weight of 60 kg for an adult;<sup>3</sup> and *C* is the 90<sup>th</sup> percentile concentration of each antibiotic (ng/g ww) detected in the present study (Table S10). The 90<sup>th</sup> percentile concentrations of FQs and MLs used to calculate *EDI* values were based on the concentrations detected in all samples collected from Laizhou Bay, China. 1/2MLOD was used to calculate the *EDI* values for antibiotic levels below the MLOD. To account for the biomagnification of SAs and trimethoprim, concentration used for calculation of the *EDI* value of each antibiotic were estimated as follows:

$$C_{\text{top consumer}} = 10^{\left[\log C_{Acetes chinensis} + (\log TMF) \times (TL_{\text{top consumer}} - TL_{Acetes chinensis})\right]}$$
(Eq. S7)

where  $C_{top consumer}$  is the concentration of a SA or trimethoprim in the top consumer (there are no more than 4 *TLs* in a coastal food web, *i.e.*, *TL*<sub>top consumer</sub> = 4); *C*<sub>Acetes</sub> <sub>chinensis</sub> is the 90<sup>th</sup> percentile concentration of a SA or trimethoprim in the northern maoxia shrimp (*Acetes chinensis*) (*TL*<sub>Acetes chinensis</sub> = 2); and *TMF* is the trophic magnification factor of the antibiotics estimated in the present study. **Determination of Hazard Quotients** (*HQ*). Human health risks of rural and urban residents associated with antibiotic exposure following the consumption of seafood were assessed using hazard quotients calculated for each antibiotic with the following equation:

$$HQ = \frac{EDI}{ADI}$$
(Eq. S8)

where HQ is the hazard quotient and ADI is the acceptable daily intake (µg/kg bw/day) of an antibiotic. ADI values are based on a review of the data published by the World Health Organization/Food and Agriculture Organization (WHO/FAO), the Joint Expert Committee on Food Additives (JECFA), the Australian Government, Department of Health, Office of Chemical Safety (OCS), or other relevant sources. ADIs and HQ values of all 19 antibiotics are summarized in Table S10. A comparison of HQ values of individual antibiotics for rural *versus* urban residents is shown in Figure S4.

Antibiotics	Abbreviation	CAS number	$\log K_{\rm OW}$ <sup>a</sup>	logD <sup>b</sup>	$pK_a^{\ a}$	$f_{ m n}{}^{b}$	Predominant ionized species at pH = 7.5
Macrolides (MLs)							
Azithromycin	AZI	83905-01-5	4.02	2.76	8.74	0.05	Cation
Clarithromycin	CLA	81103-11-9	3.16	1.66	8.99	0.03	Cation
Anhydroerythromycin	ERY-H <sub>2</sub> O	114-07-8	3.06	1.64	8.90	0.04	Cation
Roxithromycin	ROX	80214-83-1	2.75 <sup>j</sup>	2.00	8.16 <sup>c</sup>	0.18	Cation
Sulfonamides (SAs)							
Sulfadiazine	SDZ	68-35-9	0.39 <sup>d</sup>	-0.24	6.99 <sup>d</sup>	0.24	Anion
Sulfachlorpyridazine	SCP	80-32-0	1.30 <sup>e</sup>	0.26	6.50 <sup><i>f</i></sup>	0.09	Anion
Sulfamonomethoxine	SMM	1220-83-3	0.61 <sup>c</sup>	-0.28	6.67 <sup>c</sup>	0.13	Anion
Sulfadimethoxine	SM2	122-11-2	1.63 <sup>j</sup>	-0.02	5.86 <sup>g</sup>	0.02	Anion
Sulfamethoxazole	SMX	723-46-6	0.89	-0.24	6.40 <sup><i>h</i></sup>	0.07	Anion
Sulfapyridine	SPD	144-83-2	0.03 <sup>c</sup>	-0.01	8.54 <sup>c</sup>	0.92	Anion
Sulfathiazole	ST	72-14-0	0.05	-0.40	7.24	0.35	Anion
Sulfamerazine	SMZ	127-79-7	0.38 <sup>c</sup>	0.00	7.35 <sup>c</sup>	0.41	Anion
Sulfadimidine	SDM	57-68-1	0.14	-0.09	7.65	0.59	Anion
Trimethoprim	TMP	738-70-5	0.79 <sup>c</sup>	0.09	6.90 <sup>c</sup>	0.20	Anion

 Table S1. Physicochemical properties of antibiotics investigated in this study.

Antibiotics	Abbreviation	CAS number	$\log K_{\rm OW}$ <sup>a</sup>	logD <sup>b</sup>	pKa <sup>a</sup>	$f_{\rm n}{}^b$	Predominant ionized species at pH = 7.5
Fluoroquinolones (FQs)							
Norfloxacin	NOR	70458-96-7	0.46	-0.81	8.75	0.05	Cation
Enoxacin	ENX	74011-58-8	-0.20 <sup>c</sup>	-0.97	8.19 <sup>c</sup>	0.17	Cation
Ofloxacin	OFL	82419-36-1	-0.39	-0.87	7.81 <sup><i>i</i></sup>	0.33	Cation
Enrofloxacin	ENR	93106-60-6	0.44 <sup>c</sup>	-0.79	8.70 <sup>j</sup>	0.06	Cation
Ciprofloxacin	CIP	85721-33-1	0.28	-0.98	8.74	0.05	Cation

## **Continued Table S1.**

<sup>*a*</sup> Values were obtained from the Hazardous Substances Data Bank (HSDB) (available at <u>https://toxnet.nlm.nih.gov/newtoxnet/hsdb.htm</u>) unless indicated otherwise.

<sup>*b*</sup> pH dependent distribution coefficient (log*D*) and fraction of neutral molecules ( $f_n$ ) were calculated according to reference (4) at pH = 7.5, which is suggested to be the mean value of pH ranges in the intestinal tract of marine fish by reference (5).

<sup>c</sup> Values were calculated using Advanced Chemistry Development (ACD/I-Lab) Software 2.0 (available at <u>https://ilab.acdlabs.com/iLab2/)</u>. All

 $pK_a$  values are given for predominant ionized species at pH = 7.5.

<sup>d</sup> From reference (6). <sup>e</sup> From reference (7). <sup>f</sup> From reference (8). <sup>g</sup> From reference (9). <sup>h</sup> From reference (10). <sup>i</sup> From reference (11). <sup>j</sup> From

reference (12).<sup>j</sup> From ChemIDplus (available at https://chem.nlm.nih.gov/chemidplus/chemidlite.jsp)

and the corresponding sampling locations (Figure S1) and sampling times.						
Species <sup><i>a</i></sup>	Sampling location <sup>b</sup>	Sampling time	Length (cm) <sup><i>c</i></sup>	Weight (g) <sup>c</sup>		
Invertebrates (9)						
Veined rapa whelk (Rapana venosa)	S2, S4, S5	2014/10, 2015/4	8.2 ± 0.5	96 ± 6		
Mactra quadrangularis (Mactra veneriformis)	S1	2014/6	$3.3 \pm 0.2$	$8.4 \pm 1.6$		
Razor clam (Sinonovacula constricta)	S3	2014/6	$6.7 \pm 0.4$	$8.5 \pm 1.8$		
Octopus (Octopus ocellatus)	S2, S4	2014/10, 2015/4	$27.0 \pm 0.2$	$90\pm7$		
Mantis shrimp (Oratosquilla oratoria)	S2, S4	2014/10	$13.7 \pm 1.2$	$32 \pm 4$		
Northern maoxia shrimp (Acetes chinenisis)	S2, S5	2014/10, 2015/4	$3.3 \pm 1.6$	$1.7 \pm 1.0$		
Japanese stone crab (Charybdis japonica)	S2, S4	2014/10, 2015/4	$7.6 \pm 0.4$	$36 \pm 6$		
Swimming crab (Portunus trituberculatus)	S2, S4	2014/10, 2015/4	$10.3 \pm 1.1$	$52 \pm 10$		
Prawn (Fenneropenaeus chinensis)	S2	2014/10, 2015/4	$6.7 \pm 1.4$	$8.5 \pm 1.8$		
Benthic fish (5)						
Javeline goby (Acanthogobius hasta)	S1, S2	2014/3, 2014/6, 2014/10, 2015/4	27.4 ± 2.7	80 ± 8		
Eelgoby (Odontamblyopus rubicundus)	S4	2014/6	$30.7 \pm 2.8$	$62 \pm 6$		
Fat greenling (Hexagrammos otakii)	S4	2014/6	$21.5 \pm 0.9$	$121 \pm 7$		
Flathead (Platycephalus indlcus)	S1, S2, S4	2014/3, 2014/10	$42.8 \pm 3.7$	$430\pm30$		
Tongue sole (Cymoglossus robustus)	S2, S4	2014/10, 2015/4	$31.0 \pm 2.1$	$176 \pm 51$		

Table S2. Summary of the average body lengths and body weights of the specimens collected from Laizhou Bay, China,

## **Continued Table S2.**

Species <sup><i>a</i></sup>	Sampling locations <sup>b</sup>	Sampling time	Length (cm) <sup><i>c</i></sup>	Weight (g) <sup>c</sup>
Pelagic fish (5)				
Halfbeak (Hemirhamphus sajori)	S4	2014/10	$10.7 \pm 1.0$	$1.9 \pm 0.7$
Dotted gizzard shad (Konosirus punctatus)	S2	2015/4	$18.1 \pm 1.7$	$65 \pm 9$
Mullet (Liza haematocheilus)	S2	2014/3, 2014/6, 2014/10	$25.4 \pm 1.9$	$152 \pm 37$
Chinese sea perch (Lateolabrax maculatus)	S2	2014/10, 2015/4	$24.8\pm2.6$	$156 \pm 38$
Silvery pomfret (Pampus argenteus)	S5	2014/10	$16.5 \pm 1.2$	$134 \pm 15$

<sup>*a*</sup> The species were identified and classified using the SealifeBase database (available at <u>http://www.sealifebase.org</u>) and Checklist of Marine

biota of China Seas.<sup>13</sup>

<sup>b</sup> The coordinates of longitude and latitude for sampling sites are showed in Figure S1.

<sup>*c*</sup> Data are presented as the mean value  $\pm$  one standard deviation.

$C_{a} = 1^{a}$	MRM transition	Fragment voltage	Collision energy	Retention time
Compound <sup>a</sup>	(m/z)	(V)	(eV)	(min)
AZI	$749.9 \rightarrow 115.9$	215	49	14.7
ERY-H <sub>2</sub> O	$716.5 \rightarrow 158.2$	160	25	20.1
CLA	$748.9 \rightarrow 158.0$	170	29	21.1
ROX	$838.0 \rightarrow 158.0$	190	37	21.5
SDZ	$251.3 \rightarrow 92.1$	100	25	5.0
SCP	$285.0 \rightarrow 155.8$	110	13	10.0
SMM	$281.3 \rightarrow 92.1$	120	36	10.3
SM2	$311.0 \rightarrow 155.9$	125	17	13.9
SMX	$254.0 \rightarrow 108.0$	105	25	10.7
SPD	$250.3 \rightarrow 155.9$	110	13	6.7
ST	$256.0 \rightarrow 155.9$	110	13	6.3
SMZ	$265.0 \rightarrow 108.0$	122	29	7.1
SDM	$279.0 \rightarrow 92.0$	112	33	8.8
TMP	$291.0 \rightarrow 230.0$	165	21	8.5
ENX	$321.3 \rightarrow 303.2$	130	21	9.2
CIP	$332.2 \rightarrow 314.1$	124	17	9.8
ENR	$360.4 \rightarrow 342.1$	134	17	10.7

**Table S3.** Optimized LC-MS/MS parameters and retention times of the antibiotics analyzed in the present study.

Continued 7	Fable	<b>S3</b> .
-------------	-------	-------------

Common d <sup>a</sup>	MRM transition	Fragment voltage	Collision energy	Retention time
Compound <sup>a</sup>	(m/z)	(V)	(eV)	(min)
OFL	$362.0 \rightarrow 318.0$	130	15	9.5
NOR	$320.0 \rightarrow 302.0$	128	21	9.6
Sulfamethoxazole-D <sub>4</sub> (surrogate)	$258.0 \rightarrow 112.0$	105	25	10.7
Atrazine-D <sub>5</sub> (surrogate)	$221.5 \rightarrow 179.4$	80	12	17.9
Trimethoprim-D <sub>3</sub> (surrogate)	$294.4 \rightarrow 123.0$	152	21	8.5
Ciprofloxacin-D <sub>8</sub> (surrogate)	$340.2 \rightarrow 235.1$	125	44	9.9
Caffeine- <sup>15</sup> N <sub>2</sub> (IS)	$197.1 \rightarrow 139.1$	90	18	7.5

<sup>*a*</sup> The full names of the compounds can be found in Table S1.

Compound <sup><i>a</i></sup>	Spiked surregete	Range of calibration	14	MLOD $^{b}$	MLOQ <sup>b</sup>	Recovery <sup>c</sup>	RSD
Compound	Spiked surrogate	curve (ng/mL)	r	(ng/g ww)	(ng/g ww)	(%)	(%)
CIP	Ciprofloxacin-D <sub>8</sub>	2 - 2000	0.998	1.5	4.9	98	14
ENX	Ciprofloxacin-D <sub>8</sub>	2 - 2000	0.998	1.0	3.5	49	18
ENR	Ciprofloxacin-D <sub>8</sub>	2 - 2000	0.998	1.1	3.6	97	13
NOR	Ciprofloxacin-D <sub>8</sub>	2 - 2000	0.997	1.2	3.9	74	13
OFL	Ciprofloxacin-D <sub>8</sub>	2 - 2000	0.998	0.7	2.2	128	13
AZI	Atrazine-D <sub>5</sub>	0.5 - 100	0.997	0.3	0.9	111	9
CLA	Atrazine-D <sub>5</sub>	0.5 - 100	0.997	0.3	1.0	91	15
ERY-H <sub>2</sub> O	Atrazine-D <sub>5</sub>	0.5 - 100	0.995	0.1	0.3	82	11
ROX	Atrazine-D <sub>5</sub>	0.5 - 100	0.998	0.1	0.4	64	14
TMP	Trimethoprim-D <sub>3</sub>	0.1 - 100	0.998	0.02	0.08	90	14
SMX	Sulfamethoxazole-D <sub>4</sub>	1 - 1000	0.999	0.9	2.8	117	12
SM2	Sulfamethoxazole-D <sub>4</sub>	1 - 1000	0.996	0.2	0.7	80	18
SPD	Sulfamethoxazole-D <sub>4</sub>	1 - 1000	0.997	0.3	0.9	113	16

Table S4. Surrogate standards, linear range of matrix matched calibration curves, method detection limits (MLOD), method quantification limits

(MLOQ), spiked average recoveries and relative standard deviations (RSD, n = 6) for each antibiotic.

Continued 7	<b>Fable</b>	<b>S4</b> .
-------------	--------------	-------------

Compound <sup>a</sup>	Spiked surrogate	Range of calibration curve (ng/mL)	r	MLOD <sup>b</sup> (ng/g ww)	MLOQ <sup>b</sup> (ng/g ww)	Recovery <sup>c</sup> (%)	RSD (%)
ST	Sulfamethoxazole-D4	1 - 1000	0.996	0.5	1.8	67	17
SDM	Sulfamethoxazole-D4	1 - 1000	0.998	0.6	2.0	111	19
SMZ	Sulfamethoxazole-D4	1 - 1000	0.997	0.4	1.2	112	17
SDZ	Sulfamethoxazole-D4	1 - 1000	0.997	0.5	1.7	110	14
SCP	Sulfamethoxazole-D4	1 - 1000	0.997	0.4	1.2	102	18
SMM	Sulfamethoxazole-D4	1 - 1000	0.997	1.5	5.0	114	8

<sup>*a*</sup> The full names of the compounds can be found in Table S1.

<sup>b</sup> MLOD and MLOQ are defined as three and ten times the standard deviation (SD) of the mean procedure blanks (n = 6), respectively.

 $^{c}$  The recovery of each antibiotic was determined by analysis of six replicates of non-contaminated biotic matrix (*i.e.*, farm raised fish) spiked with a mixture of antibiotic standards. Different concentrations (5 to 250 ng/g) were spiked for each antibiotic to account for difference in response of all antibiotics during simultaneous analysis.

Compound <sup>a</sup>	Detection frequency (%)	Compound <sup><i>a</i></sup>	Detection frequency (%)
SCP	63	CIP	23
SDZ	47	ENX	33
SM2	23	ENR	60
SMZ	72	NOR	49
SDM	47	OFL	67
SMX	95	AZI	21
SMM	39	CLA	25
SPD	25	ERY-H <sub>2</sub> O	58
ST	47	ROX	23
TMP	91		

**Table S5.** Detection frequencies of the target antibiotics in all samples.

<sup>*a*</sup> The full names of the compounds can be found in Table S1.

Species	n <sup>a</sup>	Moisture content <sup>b</sup>	TL <sup>c</sup>	$\sum$ SAs <sup>d</sup>	TMP <sup>d</sup>	$\sum FQs^{d}$	$\sum$ MLs <sup>d</sup>	Total conc. <sup>d</sup>	Frequency
Mantis shrimp	16	76	$3.60 \pm 0.04$	$190 \pm 10$	$1.9 \pm 0.1$	$200 \pm 20$	$1.8 \pm 0.1$	$390 \pm 20$	47
Manus sinnip	10	70	$5.00 \pm 0.04$	$(36 \pm 2)^{e}$	$(0.35 \pm 0.02)$	$(37 \pm 3)$	$(0.34 \pm 0.03)$	$(74 \pm 3)$	4/
Swimming crab	8	72	$3.40 \pm 0.10$	$220\pm50$	$1.1 \pm 0.5$	$310 \pm 10$	$4.5 \pm 1.0$	$540 \pm 60$	90
Swinning crao	0	12	$5.40 \pm 0.10$	$(57 \pm 13)$	$(0.3 \pm 0.1)$	$(81 \pm 3)$	$(1.2 \pm 0.3)$	$(140 \pm 20)$	)) 90
Mactra quadrangularis	34	73	$2.60 \pm 0.04$	$110 \pm 5$	$17 \pm 2$	$1,200 \pm 90$	$3.5 \pm 0.4$	$1,300 \pm 90$	79
Macha quadrangularis	34	13	$2.00 \pm 0.04$	$(19 \pm 1)$	$(2.8 \pm 0.3)$	$(190 \pm 10)$	$(0.6 \pm 0.1)$	$(220 \pm 10)$	/9
Razor clam	11	81	$2.50 \pm 0.05$	$21 \pm 2$	$34 \pm 20$	$450 \pm 40$	nd	$500 \pm 50$	47
Kazor ciain	11	01	$2.30 \pm 0.03$	$(4.1 \pm 0.3)$	$(6.6 \pm 4.0)$	$(88 \pm 8)$	na	$(98 \pm 9)$	4/
Japanese stone crab	13	79	$3.40 \pm 0.10$	$170 \pm 13$	$4.0 \pm 3.0$	$160 \pm 40$	$3.9 \pm 4.0$	$340 \pm 30$	74
Japanese stone crau	15	19	$3.40 \pm 0.10$	$(35 \pm 3)$	$(0.8 \pm 0.7)$	$(33 \pm 8)$	$(0.8 \pm 0.7)$	$(70 \pm 6)$	/4
Veined rapa whelk	15	77	$2.30 \pm 0.06$	$82 \pm 7$	$8.9 \pm 2.0$	$330 \pm 40$	$0.7 \pm 1.0$	$420 \pm 40$	47
venieu rapa wneik	15	11	$2.30 \pm 0.00$	$(22 \pm 2)$	$(2.4 \pm 0.5)$	$(90 \pm 10)$	$(0.2 \pm 0.3)$	$(110 \pm 10)$	4/
Northern maoxia	10		• • • • • • •	$51 \pm 6$	$0.30 \pm 0.02$	• f	$13 \pm 3$	$64 \pm 10$	
shrimp	49	74	$2.00 \pm 0.10$	$(12 \pm 2)$	$(0.07 \pm 0.01)$	$\mathrm{nd}^{f}$	$(2.9 \pm 0.8)$	$(15 \pm 2)$	53
- -		01		$77 \pm 30$	$1.6 \pm 0.8$	$190 \pm 30$	$3.7 \pm 1.0$	$270 \pm 40$	
Prawn	11	81	$3.20 \pm 0.30$	$(22 \pm 7)$	$(0.2 \pm 0.3)$	$(54 \pm 9)$	$(1.1 \pm 0.3)$	$(77 \pm 10)$	47
	2	0.2	2.50 + 0.05	$350 \pm 60$	$0.5 \pm 0.2$	$68 \pm 4$	$4.4 \pm 2.0$	$420 \pm 50$	70
Octopus	3	83	$3.50 \pm 0.06$	$(83 \pm 14)$	$(0.12 \pm 0.04)$	$(16 \pm 1)$	$(1.1 \pm 0.5)$	$(100 \pm 10)$	79

**Table S6.** Moisture content (%), estimated trophic level (*TL*), mean concentrations in ng/g ww (ng/g dw) of sulfonamide, fluoroquinolone,

macrolide antibiotics and trimethoprim, and detection frequency (%) in different species collected from Laizhou Bay, China.

Species	n <sup>a</sup>	Moisture content <sup>b</sup>	TL <sup>c</sup>	$\sum$ SAs <sup>d</sup>	TMP <sup>d</sup>	$\sum FQ^{d}$	$\sum$ MLs <sup>d</sup>	Total conc. <sup>d</sup>	Frequency
Fat greenling	6	41	$3.50 \pm 0.02$	$290 \pm 30$ (69 ± 6)	2.4 <sup>g</sup> 0.6 <sup>g</sup>	$92 \pm 5$ (22 ± 1)	$\mathrm{nd}^{f}$	$380 \pm 20$ (91 ± 6)	21
Javeline goby	17	73	$4.00 \pm 0.03$	$110 \pm 30$ (22 ± 7)	$2.0 \pm 2.0$ ( $0.4 \pm 0.5$ )	$100 \pm 2$ (21 ± 0.4)	$0.2 \pm 0.3$ (0.05 ± 0.07)	$210 \pm 40$ (44 ± 8)	42
Flathead	3	79	3.90 ± 0.06	$(22 \pm 7)$ $340 \pm 40$ $(70 \pm 9)$	$(0.1 \pm 0.0)$ $2.0 \pm 2.0$ $(0.4 \pm 0.4)$	$(21 \pm 0.1)$ $48 \pm 30$ $(10 \pm 7)$	$(0.05 \pm 0.07)$ $1.4 \pm 1.0$ $(0.3 \pm 0.2)$	$(11 \pm 0)$ $390 \pm 60$ $(81 \pm 10)$	37
Eelgoby	7	78	$3.50 \pm 0.01$	$440 \pm 10$	$2.2 \pm 0.2$	$130 \pm 30$	(0.3 ± 0.2) nd	$580 \pm 40$	42
Tongue sole	15	77	$3.30 \pm 0.30$	$(100 \pm 2)$ $110 \pm 70$	$(0.51 \pm 0.04)$ $5.2 \pm 3.0$	$(30 \pm 8)$ $94 \pm 20$	$1.2 \pm 1.0$	$(130 \pm 10)$ $210 \pm 50$	47
Chinese sea perch	17	79	$3.90 \pm 0.30$	$(25 \pm 16)$ $310 \pm 140$	$(1.2 \pm 0.7)$ $5.4 \pm 4.0$	$(21 \pm 4)$ 16 ± 10	$(0.3 \pm 0.3)$ $7.5 \pm 4.0$	$(47 \pm 10)$ $340 \pm 150$	37
				$(74 \pm 30)$ $100 \pm 9$	$(1.3 \pm 1.0)$ $1.3 \pm 1.0$	$(3.8 \pm 2.0)$ $12 \pm 7$	$(1.8 \pm 1.0)$ $0.4 \pm 0.6$	$(81 \pm 30)$ $110 \pm 20$	
Mullet	11	76	$3.20 \pm 0.20$	$(22 \pm 2)$ $150 \pm 20$	$(0.3 \pm 0.2)$ $6.4 \pm 0.3$	$(3 \pm 2)$ $84 \pm 30$	$(0.1 \pm 0.1)$ $3.2 \pm 1.0$	$(26 \pm 4)$ 240 ± 20	32
Silvery pomfret	6	69	$4.10 \pm 0.02$	$(46 \pm 5)$	$(2.0 \pm 0.1)$	$(26 \pm 8)$	$(1.0 \pm 0.4)$	$(75 \pm 6)$	68
Dotted gizzard shad	6	78	$3.30 \pm 0.03$	$100 \pm 10$ (62 ± 7)	$20 \pm 20$ (7.7 ± 10.0)	$93 \pm 10$ (55 ± 6)	$4.2 \pm 2.0$ (2.5 ± 1.0)	$210 \pm 30$ (130 ± 20)	26
Halfbeak	32	76	$2.20\pm0.10$	$52 \pm 20$ (14 ± 6)	$5.0 \pm 1.0$ (1.3 ± 0.3)	$86 \pm 20$ (23 ± 5)	$5.7 \pm 1.0$ (1.5 ± 0.3)	$150 \pm 40$ (40 ± 10)	68

**Continued Table S6.** 

a *n* is the number of organisms sampled for each species, and three different homogenized samples were analyzed for each species.

<sup>b</sup> Water content was calculated based on the difference between wet and dry weights.

<sup>c</sup> TL values were calculated using equation (2) as described in data analysis of the text and presented as the mean value  $\pm$  one standard deviation.

 $^{d}\Sigma$ SAs: sum of sulfonamide antibiotics; TMP: trimethoprim;  $\Sigma$ FQs: sum of fluoroquinolone antibiotics;  $\Sigma$ MLs: sum of macrolide antibiotics;

Total conc.: sum of all antibiotics investigated in this study. All concentrations are presented as the mean value  $\pm$  one standard deviation.

<sup>e</sup> Concentrations on a dry weight basis for each species are given in parentheses.

<sup>*f*</sup> nd: value below MLOD.

<sup>g</sup> Analyte was detected in only one pooled sample.

Study region	Collected species <sup><i>a</i></sup>	Antibiotic class <sup>b</sup>	Conc. Range (ng/g ww)	Reference	
	Eich and invertable	$\sum SAs + TMP$	nd <sup>e</sup> - 453 (170 <sup>c</sup> )		
Laizhou Bay, North China	Fish and invertebrates $(n = 280)$	∑FQs	nd - 1,200 (190)	This study	
	(n - 280)	∑MLs	nd - 15 (2.9)		
Coastal areas,	Shrimps, crabs, mollusks, and fish	$\sum$ SAs + TMP	nd - 58.6		
Hailing Island, South China	(n = 39)	∑FQs	nd - 153.2	(14)	
Hannig Island, South China	(n  SY)	∑MLs	nd - 15,093		
Coastal areas, Dalian,	Crassostrea gigas and scallops	∑SAs	2.18 - 63.87	(15)	
North China	(n = 13)	∑CAPs	nd - 4.03	(15)	
		∑SAs	nd - $76.7(6.0)^d$		
Bohai Sea, North China	Mollusks $(n = 190)$	∑FQs	0.7 - 1,575.1(86.8) <sup>d</sup>	(16)	
	(n = 100)	∑MLs	nd - $36.2(2.6)^d$		
		∑SAs	nd - 98.3 <sup><i>d</i></sup>		
Baiyangdian Lake, North China	Crustacean and fish $(n = 23)$	∑FQs	17.8 - 167 <sup>d</sup>	(17)	
	(n-25)	∑MLs	nd - 182 <sup><i>d</i></sup>		
		∑SAs	nd - 9.3		
Rivers, Pearl River Delta, China	Fish $(n = 128)$	∑FQs	nd -19.0	(18)	
	(n  120)	∑MLs	nd - 2.36 (< 0.49)		

**Table S7.** Comparison of antibiotic levels detected in aquatic animals in areas around the world.

## **Continued Table S7.**

Study region	Collected species <sup>a</sup>	Antibiotic class <sup>b</sup>	Conc. Ranges (ng/g ww)	Reference	
		∑SAs	nd - 996 <sup>d</sup>		
Haihe River, North China	Fish $(n = 10)$	∑FQs	nd - 63.5 <sup><i>d</i></sup>	(19)	
	(n = 10)	∑MLs	nd - 45.1 $(7.8)^d$		
	Shrimps and fish	∑FQs	nd - 0.73		
Aquatic markets, Canada	(n = 30)	ΣCAPs	nd - 0.4	(20)	
Rivers, USA	Fish $(n = 144)$	SMX + TMP	nd	(21)	
Rivers, Germany	Bream $(n = 20)$	SMX + TMP	nd	(22)	
Coastal areas, Ireland	Mussels $(n = 45)$	TMP	nd - 9.22 <sup><i>d</i></sup>	(23)	

<sup>*a*</sup> Numbers in parenthesis are the number of analyzed samples.

 $^{b}\Sigma$ SAs: sum of sulfonamide antibiotics; SMX: sulfamethoxazole; TMP: trimethoprim;  $\Sigma$ FQs: sum of fluoroquinolone antibiotics;  $\Sigma$ MLs: sum of

macrolide antibiotics;  $\Sigma$ CAPs: sum of chloramphenicol antibiotics.

<sup>c</sup> Values in parentheses are the mean concentrations.

<sup>*d*</sup> Concentrations are reported by ng/g dry weight.

<sup>*e*</sup> nd: value below MLOD.

# Table S8. Analysis of variance (ANOVA) of the calculated trophic level (TL) values at different sampling times

	Sum of Squares	Df	Mean Square	F	Sig.
Between Groups	0.313	2	0.157	0.394	0.678
Within Groups	11.916	30	0.397		
Total	12.230	32			

(*i.e.*, June 2014, October 2014 and April 2015) by SPSS 19.0.<sup>*a*</sup>

 $\overline{a}$  TL values of a species derived from isotope determination with at least two different sampling times are used in this analysis.

Antibiotic <sup><i>a</i></sup>	slope	r	<i>p</i> -value	TMF <sup>c</sup>	95% CI
SCP	0.52	0.93	< <b>0.001</b> <sup>b</sup>	3.3	2.5 - 4.5
SDZ	0.57	0.83	0.003	3.7	1.8 - 7.6
SM2	0.14	_ e	0.624	1.4	0.2 - 9.2
SMZ	0.45	-	0.368	2.8	0.3 - 30.4
SDM	0.59	0.77	0.016	3.9	1.4 - 10.7
SMX	0.34	0.85	< 0.001	2.2	1.7 - 2.9
SMM	0.15	0.65	0.036	1.4	1.0 - 1.9
SPD	0.09	0.80	0.065	1.2	1.0 - 1.6
ST	0.09	0.54	0.061	1.2	1.0 - 1.5
$\sum$ SAs <sup>d</sup>	0.42	0.78	< 0.001	2.6	1.7 - 4.0
TMP	0.38	0.54	0.019	2.4	1.2 - 4.8
CIP	-0.36	-	0.459	0.4	0.0 - 9.5
ENX	-0.28	0.66	0.044	0.5	0.3 - 1.0
ENR	-0.40	0.76	0.003	0.4	0.2 - 0.7
NOR	-0.09	-	0.481	0.8	0.4 - 1.6

Table S9. Slope, correlation coefficients (r) and p-values of regression analyses between logarithm concentrations and trophic levels (see Figure

3), and *TMF*s and 95% confidence intervals (CIs) of target antibiotics and  $\Sigma$ SAs,  $\Sigma$ FQs,  $\Sigma$ MLs.

Antibiotic <sup><i>a</i></sup>	slope	r	<i>p</i> -value	TMF <sup>c</sup>	95% CI
OFL	-0.42	0.77	<b>0.001</b> <sup>b</sup>	0.4	0.2 - 0.6
$\sum$ FQs <sup>d</sup>	-0.48	0.69	0.002	0.3	0.2 - 0.6
AZI	-0.16	0.88	0.031	0.7	0.5 - 0.9
CLA	-0.02	-	0.900	1.0	0.3 - 2.6
ERY-H <sub>2</sub> O	-0.20	0.66	0.013	0.6	0.5 - 0.9
ROX	_ e	-	-	-	-
$\sum$ MLs <sup>d</sup>	-0.32	0.81	0.001	0.5	0.3 - 0.7
Total <sup>d</sup>	-0.14	0.22	0.205	0.7	0.4 - 1.2

**Continued Table S9.** 

<sup>*a*</sup> The full names of the compounds can be found in Table S1.

<sup>b</sup> p-values in bold print represent statistically significant increases or decreases of the wet weight concentration (*i. e.*, < 0.05).

<sup>c</sup> TMF values were calculated using Eq. S3 and Eq. S4 as described above.

 $^{d}$   $\Sigma$ SAs: sum of sulfonamide antibiotics;  $\Sigma$ FQs: sum of fluoroquinolone antibiotics;  $\Sigma$ MLs: sum of macrolide antibiotics; Total: sum of all antibiotics investigated in this study.

<sup>*e*</sup> -: not available.

Antibiotic <sup><i>a</i></sup>	$C^{b}$	EDI <sup>c</sup> (ng	/kg bw/d)	HQ <sup>d</sup>	$1 \times 10^{2}$	ADI	Reference	
Antibiotic	(ng/g ww)	urban	rural	urban rural		(µg/kg bw/d)	Kelelellee	
SCP	39.8	22.9	12.4	-	-	_ e	-	
SDZ	3.5	2.0	1.1	0.01	0.005	20	(24)	
SM2	1.8	1.1	0.6	0.01	0.006	10	(25)	
SMZ	38.1	21.9	11.8	-	-	-	-	
SDM	36.0	20.7	11.2	0.04	0.02	50	(26)	
SMX	160.9	92.4	50.0	0.2	0.09	57	(27)	
SMM	14.3	8.2	4.4	0.1	0.07	6.0	(24)	
SPD	1.9	1.1	0.6	0.03	0.02	3.3	(27)	
ST	4.0	2.3	1.3	0.005	0.003	50	(28)	
TMP	1.8	1.0	0.6	0.02	0.01	4.0	(28)	
$\sum SAs^{e}$	-	-	-	0.4	0.2	-	-	
CIP	27.8	16.0	8.6	1.0	0.5	1.6	(29)	
ENX	32.0	18.4	9.9	-	-	-	-	
ENR	250.9	144.1	77.9	7.2	3.9	2.0	(30)	
NOR	115.9	66.6	36.0	0.6	0.3	11.4	(27)	

**Table S10.** 90<sup>th</sup> percentile antibiotic concentrations (*C*), estimation daily intake (*EDI*) values and associated hazard quotients (*HQ*) of antibiotics

for rural and urban residents and acceptable daily intake (ADI) values obtained from the literature.

Antibiotic <sup><i>a</i></sup>	$C^{b}$	EDI <sup>c</sup> (ng.	/kg bw/d)	$HQ^{d} \times 10^{2}$		ADI	Reference
Anubiotic	(ng/g ww)	urban	rural	urban	rural	$(\mu g/kg bw/d)$	Reference
OFL	49.4	28.4	15.4	0.5	0.3	5.7	(27)
∑FQs	-	_ e	-	9.4	5.0	-	-
AZI	1.7	1.0	0.5	0.06	0.03	1.7	(31)
CLA	3.6	2.1	1.1	1.0	0.6	0.2	(32)
ERY-H <sub>2</sub> O	2.5	1.4	0.8	0.2	0.1	0.7	(33)
ROX	0.8	0.5	0.3	0.1	0.07	0.4	(34)
∑MLs	-	-	-	1.4	0.8	-	-

**Continued Table S10.** 

<sup>*a*</sup> The full names of the compounds can be found in Table S1.

<sup>*b*</sup> For SAs and TMP, concentration used to calculate the *EDI* value of each antibiotic (*C*) was estimated using Eq. S7 as described above. For FQs and MLs, *C* was the 90<sup>th</sup> percentile concentration of each antibiotic detected in the present study. 1/2MLOD was used to calculate the *EDI* values for antibiotic levels below the MLOD.

<sup>c</sup> EDIs were calculated using Eq. S6 as described above based on the consumption of a diet containing only seafood.

<sup>*d*</sup> *HQ* values were calculated using Eq. S8 as described above.

<sup>*e*</sup> -: not available.

 $^{f}\Sigma$ SAs: TMP was included based on similar mode of action (MOA) between SAs and TMP.

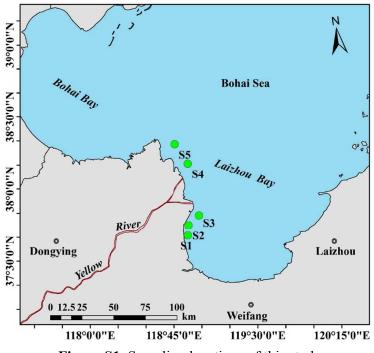
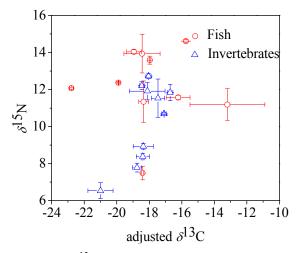
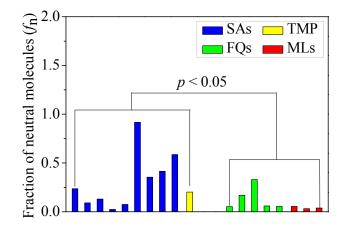


Figure S1. Sampling locations of this study.

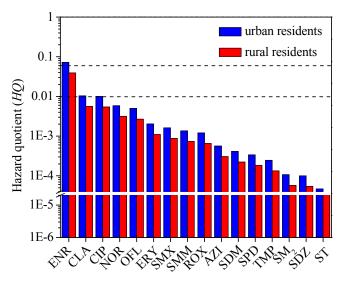


**Figure S2.**  $\delta^{15}$ N and adjusted  $\delta^{13}$ C values for the species in the coastal food web from Laizhou Bay, China.  $\delta^{15}$ N and  $\delta^{13}$ C values were calculated using Eq. S1 and Eq. S2, respectively, as described above. All  $\delta^{13}$ C values were mathematically adjusted for the C/N ratios obtained during the determination of the stable-isotope ratios as described

in equation (1).



**Figure S3.** Significant difference (p < 0.05) on fraction of neutral molecules ( $f_n$ ) values between SA antibiotics (including TMP) and the other two antibiotic groups (*i.e.*, FQs and MLs).  $f_n$  values of antibiotics (Table S1) were calculated using an equation described in Fu et al. at pH = 7.5, which is the mean value of pH ranges in the intestinal tract of marine fish reported by Ou et al.<sup>4,5</sup>



**Figure S4.** Hazard quotients (*HQ*) of antibiotics for rural and urban residents based on consumption of seafood only. Dotted lines represent threshold values of a considerable risk (*i.e.*, *HQ* = 0.01) and a distinct risk (*i.e.*, *HQ* = 0.05), respectively.

#### References

 McKinney, M. A.; McMeans, B. C.; Tomy, G. T.; Rosenberg, B.; Ferguson, S. H.; Morris, A.; Muir, D. C. G.; Fisk, A. T. Trophic transfer of contaminants in a changing arctic marine food web: Cumberland Sound, Nunavut, Canada. *Environ. Sci. Technol.* 2012, *46* (18), 9914-9922; DOI 10.1021/es302761p.

(2) National Burean of Statistics of China; <u>http://data.stats.gov.cn</u>. Access date: July 2016.

(3) Lee, M. M.; Wu-Williams, A.; Whittemore, A. S.; Zheng, S.; Gallagher, R.; Teh, C.
Z.; Zhou, L.; Wang, X. H.; Chen, K.; Ling, C. D.; Jiao, D. A.; Jung, D.;
Paffenbarger-Jr, R. S. Comparison of dietary habits, physical activity and body size among Chinese in North America and China. *Int. J. Epidemiol.* **1994**, *23* (5), 984- 990;
DOI 10.1093/ije/23.5.984.

(4) Fu, W. J.; Franco, A.; Trapp, S. Methods for estimating the bioconcentration factor of ionizable organic chemicals. *Environ. Toxicol. Chem.* **2009**, *28* (7), 1372-1379; DOI 10.1897/08-233.1.

(5) Ou, Y. J.; Luo, Q.; Li, J. E. On digestive enzyme activity of *Trachinotus ovatus* III Effects of pH on activity in young and adult fish. *Marine Fisheries* (Chinese). **2010**, *32* (4), 417-421; DOI 10.13233/j.cnki.mar.fish.2010.04.009.

(6) The Human Metabolome Database (HMDB); <u>http://www.hmdb.ca/</u>. Access date: Nov 2016.

(7) Boxall, A. B. A.; Kay, P.; Blackwell, P. A. Livestock Farming and the Environment: Assessing the environmental fate and effects of veterinary medicines.Conference workshop series: Proceedings of Workshop 4 on Sustainable Animal Production, held at Hannover, September 28, 2000.

S35

(8) Thomaidis, N. S.; Gago-Ferrero, P.; Ort, C.; Maragou, N. C.; Alygizakis, N. A.; Borova, V. L. Dasenaki, M. E. Reflection of socioeconomic changes in wastewater: licit and illicit drug use patterns. *Environ. Sci. Technol.* **2016**, *50* (18), 10065-10072; DOI 10.1021/acs.est.6b02417.

(9) Lin, C. E.; Chang, C. C.; Lin, W. C. Migration behavior and separation of sulfonamides in capillary zone electrophoresis III. Citrate buffer as a background electrolyte. *J. Chromatogr. A.* **1997**, *768* (1), 105-112; DOI 10.1016/S0021-9673(97)00010-1.

(10) Ye, S.; Yao, Z. W.; Na, G. S.; Wang, J. Y.; Ma, D. Y. Rapid simultaneous determination of 14 sulfonamides in wastewater by liquid chromatography tandem mass spectrometry. *J. Sep. Sci.* **2007**, *30* (15), 2360-2369; DOI 10.1002/jssc.200600539.

(11) Völgyi, G.; Ruiz, R.; Box, K.; Comer, J.; Bosch, E.; Takács-Nováka, K.
Potentiometric and spectrophotometric p*K*a determination of water-insoluble compounds: validation study in a new cosolvent system. *Anal. Chim. Acta.* 2007, *583*(2), 418-428; DOI 10.1016/j.aca.2006.10.015.

(12)Lizondo, M.; Pons, M.; Gallardo, M.; Estelrich, J. Physicochemical properties of enrofloxacin. *J. Pharm. Biomed. Anal.* **1997**, *15* (12), 1845-1849; DOI 10.1016/S0731-7085(96)02033-X.

(13)Liu, R. Y. *Checklist of Marine Biota of China Seas*. Science Press: Academia Sinica, Beijing, China, 2008.

(14)Chen, H.; Liu, S.; Xu, X. R.; Liu, S. S.; Zhou, G. J.; Sun, K. Y.; Zhao, J. L.; Ying, G. G. Antibiotics in typical marine aquaculture farms surrounding Hailing Island, South China: Occurrence, bioaccumulation and human dietary exposure. *Mar. Pollut. Bull.* 2015, *90* (1-2), 181-187; DOI 10.1016/j.marpolbul.2014.10.053.

(15)Na, G. S.; Fang, X. D.; Cai, Y. Q.; Ge, L. K.; Zong, H. M.; Yuan, X. T.; Yao, Z.
W.; Zhang, Z. F. Occurrence, distribution, and bioaccumulation of antibiotics in coastal environment of Dalian, China. *Mar. Pollut. Bull.* 2013, *69* (1-2), 233-237; DOI 10.1016/j.marpolbul.2012.12.028.

(16)Li, W. H.; Shi, Y. L.; Gao, L. H.; Liu, J. M.; Cai, Y. Q. Investigation of antibiotics in mollusks from coastal waters in the Bohai Sea of China. *Environ. Pollut.* **2012**, *162*, 56-62; DOI 10.1016/j.envpol.2011.10.022.

(17)Li, W. H.; Shi, Y. L.; Gao, L. H.; Liu, J. M.; Cai, Y. Q. Occurrence of antibiotics in water, sediments, aquatic plants, and animals from Baiyangdian Lake in North China. *Chemosphere*. 2012, 89 (11), 1307-1315; DOI 10.1016/j.chemosphere.2012.05.079.

(18)Zhao, J. L.; Liu, Y. S.; Liu, W. R.; Jiang, Y. X.; Su, H. C.; Zhang, Q. Q.; Chen, X. W.; Yang, Y. Y.; Chen, J.; Liu, S. S.; Pan, C. G.; Huang, G. Y.; Ying, G. G. Tissue-specific bioaccumulation of human and veterinary antibiotics in bile, plasma, liver and muscle tissues of wild fish from a highly urbanized region. *Environ. Pollut.* 2015, *198*, 15-24; DOI 10.1016/j.envpol.2014.12.026.

(19)Gao, L. H.; Shi, Y. L.; Li, W. H.; Liu, J. M.; Cai, Y. Q. Occurrence, distribution and bioaccumulation of antibiotics in the Haihe River in China. *J. Environ. Monitor.* **2012**, *14* (4), 1248-1255; DOI 10.1039/c2em10916f.

(20) Tittlemier, S. A.; Van de Riet, J.; Burns, G.; Potter, R.; Murphy, C.; Rourke, W.; Pearce, H.; Dufresne, G. Analysis of veterinary drug residues in fish and shrimp composites collected during the Canadian Total Diet Study, 1993-2004. *Food Addit. Contam.* **2007**, *24* (1), 14-20; DOI 10.1080/02652030600932937.

(21)Ramirez, A. J.; Brain, R. A.; Usenko, S.; Mottaleb, M. A.; O'Donnell, J. G.; Stahl, L. L.; Wathen, J. B.; Snyder, B. D.; Pitt, J. L.; Perez-Hurtado, P.; Dobbins, L. L.;

Brooks, B. W.; Chambliss, C. K. Occurrence of pharmaceuticals and personal care products in fish: Results of a national pilot study in the United States. *Environ. Toxicol. Chem.* **2009**, *28* (12), 2587-2597; DOI 10.1897/08-561.1.

(22)Subedi, B.; Du, B. W.; Chambliss, C. K.; Koschorreck, J.; Rudel, H.; Quack, M.; Brooks, B. W.; Usenko, S. Occurrence of pharmaceuticals and personal care products in German fish tissue: A national study. *Environ. Sci. Technol.* **2012**, *46* (16), 9047-9054; DOI 10.1021/es301359t.

(23) McEneff, G.; Barron, L.; Kelleher B.; Paull, B.; Quinn, B. A. A year-long study of the spatial occurrence and relative distribution of pharmaceutical residues in sewage effluent, receiving marine waters and marine bivalves. *Sci. Total Environ.* 2014, 476, 317-326; DOI 10.1016/j.scitotenv.2013.12.123.

(24) ADI List: Acceptable daily intakes for agricultural and veterinary chemicals;
ISSN 1446-1412; Australian Government, Department of Health, Office of Chemical Safety (OCS): Canberra, ACT, 2015;
http://www.health.gov.au/internet/main/publishing.nsf/Content/6279C451E3D11E89C
A257BF0001DAAE7/\$File/ADI%20List\_updated%20to%2031%20Dec%202015.pdf.
Access date: April 2016.

(25) Simazaki, D.; Hiramatsu, S.; Fujiwara, J.; Akiba, M.; Kunikane, S. Monitoring priority of residual pharmaceuticals in water sources and drinking water in Japan. *J. Water Environ. Technol.* **2014**, *12* (3), 275-283; DOI 10.2965/jwet.2014.275.

(26) Evaluation of certain veterinary drug residues in food: 48th meeting report of the Joint FAO/WHO Expert Committee on Food Additives; ISSN 0512-3054; Food and Agriculture Organization of the United Nations, World Health Organization (FAO/WHO): Geneva, Switzerland, 1995; http://apps.who.int/iris/bitstream/10665/37341/1/WHO\_TRS\_851.pdf. Access date: April 2016.

(27) *Residues of Some Veterinary Drugs in Foods and Animals*; Food and Agriculture Organization of the United Nations, World Health Organization (FAO/WHO): Rome, Italy, 2011.

(28) Schwab, B. W.; Hayes, E. P.; Fiori, J. M.; Mastrocco, F. J.; Roden, N. M.; Cragin, D.; Meyerhoff, R. D.; D'Aco, V. J.; Anderson, P. D. Human pharmaceuticals in US surface waters: a human health risk assessment. *Reg. Toxicol. Pharmacol.* **2005**, *42*, 296-312; DOI 10.1016/j.yrtph.2005.05.005.

(29) Younghee, K.; Jinyong, J.; Myunghyun, K.; Park, J.; Boxall, A. B. A.; Choi, K. Prioritizing veterinary pharmaceuticals for aquatic environment in Korea. *Environ. Toxicol. Pharmacol.* **2008**, *26* (2), 167-176; DOI 10.1016/j.etap.2008.03.006.

(30) Evaluation of certain veterinary drug residues in food: 48th meeting report of the Joint FAO/WHO Expert Committee on Food Additives; ISSN 0512-3054; Food and Agriculture Organization of the United Nations, World Health Organization (FAO/WHO): Geneva, Switzerland, 1998; http://apps.who.int/iris/bitstream/10665/42127/1/WHO\_TRS\_879.pdf. Access date: April 2016.

(31) Kitris, M. D.; Goldstein, F. W.; Miégi, M.; Acar, J. F. *In-vitro* activity of azithromycin against various Gram-negative bacilli and anaerobic bacteria. *J. Antimicrob. Chemother.* 1990, 25 (Suppl. A), 15-18; DOI 10.1093/jac/25.suppl\_A.15.
(32) Citron, D. M.; Appleman, M. D. 2001. Comparative *in vitro* activities of ABT-773 against 362 clinical isolates of anaerobic bacteria. *Antimicrob. Agents Chemother.* 2001, 45, 345-348; DOI 10.1128/AAC.45.1.345-348.2001.

(33) Evaluation of certain veterinary drug residues in food: 66th meeting report of the Joint FAO/WHO Expert Committee on Food Additives; ISSN 0512-3054; Food and Agriculture Organization of the United Nations, World Health Organization (FAO/WHO): Geneva, Switzerland, 2006; http://apps.who.int/iris/bitstream/10665/43464/1/9241209399\_eng.pdf. Access date: April 2016.

(34) Dubreuil, L. *In-vitro* comparison of roxithromycin and erythromycin against 900 anaerobic bacterial strains. *J. Antimicrob. Chemother.* **1987**, *20* (Suppl B), 13-19; DOI 10.1093/jac/20.suppl\_B.13.