

Context-dependent interactions and the regulation of species richness in freshwater fish

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Supplementary Table 1. Species list of (A) Ontario fish species included in this study, based on species captured by BSM surveys in at least 40 of the 721 lakes, and (B) species grouped into the category of ‘small prey’: these species are assumed to be prey for most other fish species, and mostly inhabit the littoral zone of lakes. Further, their small size means they tend to be under-sampled by the BSM net-based fish sampling. The remaining species (C) occurred in fewer than 40 lakes, and were excluded from the analysis due to their infrequency. The 721 lakes occur over an 11-degree latitudinal gradient (43.06 – 54.52 degrees), range in area from 21-90,484 ha, have maximum depths of 1-214 m, range in phosphorus levels from 1.7-52.5 units, and have a range of total degree days from 1059-2246. Average length for each species are from Ontario lakes, and taken from Holm et al (2010)¹. Thermal guilds (water temperature preferences) are taken from Hasnain et al. 2013².

Common Name	Species name	Family	Number of lakes (n = 721)	Average length (cm)	Thermal guild	Functional Group
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(A) *Large fishes species*

Northern Pike	<i>Esox lucius</i>	Esocidae	528	61.0	Cool	Predator
Muskellunge	<i>Esox</i>	Esocidae	43	96.5	Warm	Predator

	<i>masquinongy</i>					
Walleye	<i>Sander vitreus</i>	Percidae	499	42.0	Cool	Predator
Lake Trout	<i>Salvelinus</i>	Salmonidae	298	44.5	Cold	Predator
	<i>namaycush</i>					
Burbot	<i>Lota lota</i>	Gadidae	250	38.0	Cold/Cool	Predator
Cisco	<i>Coregonus artedi</i>	Salmonidae	385	25.0	Cold	Pelagic
Lake	<i>Coregonus</i>	Salmonidae	374	38.0	Cold	Pelagic
Whitefish	<i>clupeaformis</i>					
Brook Trout	<i>Salvelinus</i>	Salmonidae	77	28.0	Cold	Pelagic
	<i>fontinalis</i>					
Longnose	<i>Catostomus</i>	Catostomidae	82	33.5	Cold	Littoral
Sucker	<i>catostomus</i>					
White Sucker	<i>Catostomus</i>	Catostomidae	678	41	Cool	Littoral
	<i>commersonii</i>					
Shorthead	<i>Moxostoma</i>	Catostomidae	90	41.0	Warm	Littoral
Redhorse	<i>macrolepidotum</i>					
Rock Bass	<i>Ambloplites</i>	Centrarchidae	295	20.0	Cool	Littoral
	<i>rupestris</i>					
Pumpkinseed	<i>Lepomis gibbosus</i>	Centrarchidae	154	18.0	Warm	Littoral
Sunfish						
Bluegill	<i>Lepomis</i>	Centrarchidae	48	19.0	Warm	Littoral
	<i>macrochirus</i>					
Smallmouth	<i>Micropterus</i>	Centrarchidae	333	30.0	Warm	Predator
Bass	<i>dolomieu</i>					
Largemouth	<i>Micropterus</i>	Centrarchidae	85	30.0	Warm	Predator

Bass	<i>salmoides</i>					
Black Crappie	<i>Pomoxis nigromaculatus</i>	Centrarchidae	55	21.5	Cool	Littoral
Brown	<i>Ameiurus</i>	Ictaluridae	133	28.0	Warm	Littoral
Bullhead	<i>nebulosus</i>					
Yellow Perch	<i>Perca flavescens</i>	Percidae	526	18.0	Cool	Littoral
Sauger	<i>Sander canadensis</i>	Percidae	45	33.0	Cool	Predator
(B) <i>Small prey fishes</i>						
Mottled	<i>Cottus bairdii</i>	Cottidae	51	7.5	Cold	Small prey
Sculpin						
Slimy Sculpin	<i>Cottus cognatus</i>	Cottidae	51	7.5	Cold	Small prey
Spoonhead	<i>Cottus ricei</i>	Cottidae	22	5.5	Cold	Small prey
Sculpin						
Deepwater sculpin	<i>Myoxocephalus thompsonii</i>	Cottidae	11	8.0	Cold	Small prey
Lake Chub	<i>Couesius plumbeus</i>	Cyprinidae	104	10.0	Cold	Small prey
Common Shiner	<i>Luxilus cornutus</i>	Cyprinidae	155	9.0	Cool	Small prey
Northern Pearl Dace	<i>Margariscus margarita</i>	Cyprinidae	50	9.0	Cold/cool	Small prey
Golden Shiner	<i>Notemigonus crysoleucas</i>	Cyprinidae	8	10.0	Cool	Small prey

Emerald Shiner	<i>Notropis atherinoides</i>	Cyprinidae	159	7.5	Cool	Small prey
Blackchin Shiner	<i>Notropis heterodon</i>	Cyprinidae	49	6.0	Cool/warm	Small prey
Blacknose Shiner	<i>Notropis heterolepis</i>	Cyprinidae	127	6.5	Cool/warm	Small prey
Spottail Shiner	<i>Notropis hudsonius</i>	Cyprinidae	344	7.0	Cold/cool	Small prey
Mimic Shiner	<i>Notropis volucellus</i>	Cyprinidae	46	6.0	Warm	Small prey
Northern Redbelly Dace	<i>Phoxinus eos*</i>	Cyprinidae	42	5.5	Cool/warm	Small prey
Bluntnose Minnow	<i>Pimephales notatus</i>	Cyprinidae	113	6.5	Warm	Small prey
Fathead Minnow	<i>Pimephales promelas</i>	Cyprinidae	42	5.0	Warm	Small prey
Blacknose Dace	<i>Rhinichthys atratulus</i>	Cyprinidae	13	8.0	Cool	Small prey
Longnose Dace	<i>Rhinichthys cataractae</i>	Cyprinidae	12	7.5	Cool	Small prey
Creek Chub	<i>Semotilus atromaculatus</i>	Cyprinidae	39	10.0	Cool	Small prey
Banded Killifish	<i>Fundulus diaphanus</i>	Fundulidae	9	7.0	Cool	Small prey
Brook	<i>Culaea</i>	Gasterosteidae	40	5.0	Cool	Small prey

Stickleback	<i>inconstans</i>					
Ninespine	<i>Pungitius</i>	Gasterosteidae	64	6.5	Warm/cold*	Small prey
Stickleback	<i>pungitius</i>					
Iowa Darter	<i>Etheostoma exile</i>	Percidae	13	5.0	Cool	Small prey
Johnny Darter	<i>Etheostoma</i>	Percidae	39	5.0	Cool	Small prey
	<i>nigrum</i>					
Logperch	<i>Percina caprodes</i>	Percidae	180	9.0	Cool/warm	Small prey
Trout-perch	<i>Percopsis</i>	Percopsidae	302	9.0	Cold	Small prey
	<i>omiscomaycus</i>					
Central	<i>Umbra limi</i>	Umbridae	9	7.5	Cool/warm	Small prey
Mudminnow						
(C) <i>Less</i>						
<i>common</i>						
<i>species</i>						
Longnose gar	<i>Lepisosteus</i>	Lepisosteidae	10	76.0	Warm	Predator
	<i>osseus</i>					
Alewife	<i>Alosa</i>	Clupeidae	5	15.0	Cold	Pelagic
	<i>pseudoharengus</i>					
Goldeye	<i>Hiodon alosoides</i>	Hiodontidae	6	28.0	Warm	Pelagic
Mooneye	<i>Hiodon tergisus</i>	Hiodontidae	13	28.0	Cool/warm	Pelagic
Rainbow trout	<i>Oncorhynchus</i>	Salmonidae	4	53.0	Cold	Pelagic
	<i>mykiss</i>					
Round	<i>Prosopium</i>	Salmonidae	12	25.0	Cold	Pelagic
whitefish	<i>cylindraceum</i>					
Bowfin	<i>Amia calva</i>	Amiidae	9	54.0	Warm	Littoral

Silver	<i>Moxostoma</i>	Catostomidae	21	40.0	Cool	Littoral
Redhorse	<i>anisurum</i>					
Greater	<i>Moxostoma</i>	Catostomidae	10	41.0	Cool/warm	Littoral
Redhorse	<i>valenciennesi</i>					
Common Carp	<i>Cyprinus carpio</i>	Cyprinidae	14	37.0	Warm	Littoral
Fallfish	<i>Semotilus</i>	Cyprinidae	13	20.0	Cool	Littoral
	<i>corporalis</i>					
Yellow	<i>Ameiurus natalis</i>	Ictaluridae	20	25.5	Warm	Littoral
Bullhead						

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29 Supplementary Table 2. Species association values, measured as null-adjusted Pearson correlation
30 coefficients for abundances of 16 species identified as potential ‘interaction’ species by the SEM
31 analysis (Figs. 1, Suppl. Fig. 3). The association analysis tests the direction and statistical significance
32 of these ‘interactions’, correcting for bias in zero-inflated data arising because the correlation
33 coefficient is not double-zero symmetrical³. Significantly negative (32/120) pairings are grey or black
34 highlighted cells. These results cannot test whether the significant associations derive from species
35 interactions, or if negative associations derive from divergent habitat affinity (e.g., species x is absent
36 when species y is present, because they prefer different lake conditions). However, by organizing
37 these results by temperature affinity (BLUE = cold water species, RED = cool, or GREEN = warm),
38 most negative interactions occur among temperature groups (grey cells) – for example, warm and cold
39 water species tending not to co-occur. Only five negative species pairings were within- group (black
40 cells). There were 37/120 positive pairings and 51/120 non-significant pairings – we assumed these
41 associations to not reflect antagonistic species interactions, and thus could not generate restricted co-
42 occurrence between the species pairs.

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Species	LNSk	Whthf	BTrt	Burb	Cisco	YPer	Pike	Wall	Wsk	Red	SMB	Pmp	LMB	RBss	BBul
Lake Trout	0.07	0.11	-0.08	0.08	0.02	0.00	-0.20	-0.27	0.18	-0.09	0.08	-0.07	-0.08	0.12	-0.05
LNSucker		0.28	0.22	-0.02	0.00	-0.05	-0.09	-0.08	0.13	-0.03	-0.08	-0.03	-0.02	-0.03	-0.04
Whitefish			-0.01	0.22	0.10	0.03	0.08	0.11	-0.02	-0.04	-0.05	-0.11	-0.10	-0.03	-0.09
Brook Trout				-0.09	-0.06	-0.11	-0.17	-0.17	0.39	-0.06	-0.12	-0.05	-0.05	-0.05	-0.04
Burbot					0.15	0.02	0.00	0.16	-0.04	0.01	0.01	-0.07	-0.02	-0.01	-0.05
Cisco						0.16	0.30	0.25	0.05	0.02	0.04	-0.03	0.01	-0.02	0.06
Yellow Perch							0.20	-0.01	-0.01	-0.04	0.10	0.24	0.20	0.03	0.18
Pike								0.42	-0.12	0.07	-0.12	0.01	0.09	-0.12	0.09
Walleye									-0.15	0.16	-0.14	-0.11	-0.08	-0.16	-0.06
White Sucker										-0.07	-0.17	-0.05	-0.08	-0.08	-0.05
Redhorse											-0.03	-0.04	-0.04	-0.04	-0.04
SM Bass												0.31	0.18	0.30	0.12
Pumpkinseed													0.27	0.08	0.27
LM Bass														0.06	0.26
Rock Bass															0.03
Brown															
Bullhead															

45 Supplementary Table 3. Boosted regression tree (BRT) model results, for species with one or more
46 significantly negative species associations corrected for null expectation (Supplementary Table 2). The
47 results from the association analysis cannot distinguish between species interactions versus habitat
48 affinity – the BRT analysis tests the relative influence (%) of 22 possible drivers of variation affecting
49 the presence of each species, including species composition (the presence of each of the other species)
50 and abiotic factors including degree days, pH, phosphorus, minimum depth, maximum depth, lake
51 area, Secchi depth, and small prey species richness. Correlation refers to how much a predictor model
52 estimates the variation within the raw data (%). The seven grey-shaded cells are the only species
53 pairings where the explanatory species (top row) was one of the ten strongest factors negatively
54 associated (see Suppl. Fig. 5) with the presence of the response species (vertical column on the left
55 side of the table). The species are organized by temperature affinity: blue = cold, green = cool, red =
56 warm.
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Focal species	lake whitefish	cisco trout	lake trout	burbot	brook trout	walleye	pike	small mouth bass	yellow perch	rock bass	white sucker	brown bullhead	pumpkin seed	large mouth bass	Degree days	pH	Total P	Depth min (m)	Depth Max (m)	Area (ha)	secchi depth	small prey richness	No. of trees included in final model	Correlation Between observed and predicted response variables
lake whitefish		20	0.02	15	0	5	0.3	0.1	2	0.02	1	0	0.2	1	22	1	4	1	2	24	1	1	750	76
cisco	21		0	0.10	1	0.10	8	0.03	3	0.05	0.01	0.3	0.10	0	7	3	3	7	5	28	4	10	1000	77
lake trout	0	0		0.01	1	5	1	0	0.01	0	1	0	0.03	0	3	2	9	46	24	2	4	0.3	1050	85
burbot	20	0	0		0.1	0.01	1	0.3	0.3	0.1	0.1	0	0.02	0.1	4	3	1	7	34	23	3	2	900	77
brook trout	0	0.05	6	1		11	11	5	56	0.1	0	0	0	0	3	1	0.2	3	1	0.5	1	1	350	81
pike	1	4	3	0.2	11	24		0.10	22	0.5	0.4	1	0.10	0.2	12	3	2	4	2	6	3	1	1400	82
walleye	3	0	5	0	0.01		36	4	0.07	0.5	1	0	0.10	0.03	12	1	4	2	2	10	1	2	1000	79
small mouth bass	0.02	0.2	0.3	2	3	5	0.2		1	8	0.00	2	10	0.02	39	1	1	4	3	11	4	10	1050	80
yellow perch	2	2	0.2	1	44	0.2	14	3		1	0.2	0.01	0.06	0	10	2	4	4	3	6	2	3	1600	86
rock bass	0	4	0.3	0.2	0.3	2	1	10	3		0	0.05	14	4	39	3	2	3	3	6	3	4	1450	79
white sucker	7	0	5	0	0.2	4	0	0	0.4	0		0	3	0.3	31	2	1	4	3	37	6	1	300	49
brown bullhead	0.1	0.02	0	0.05	0.04	0	2	1	0.01	0.1	0		17	8	60	1	3	1	0.4	1	5	0.3	700	79
Pumpkin seed	4	0	0	0	0.1	0.1	0	2	0.02	4	0	11		11	61	1	1	2	1	1	2	1	650	81
large mouth bass	0	0	0	2	0	0	0	0	0	2	3	6	7		73	2	2	2	2	3	0	0	550	83

59 Supplementary Table 4. Environmental range comparisons of the negatively associated species pairs
 60 identified in Supplementary Table 2. Species AB refers to lakes where both species co-occur. Ten
 61 pairs had range overlap in environmental factors yet rarely co-occur. Nine pairs had range
 62 compression in lakes of co-occurrence, where the range of environmental conditions in lakes where a
 63 species occurs in the absence of the other is wider than when the two species co-occur. BRT refers to
 64 pairs where the presence of one species significantly and directly predicts the absence of another (see
 65 Suppl. Table 3, Suppl. Fig. 5). **Cold water species**, **cool water species**, **warm water species** - most
 66 rarely co-occurring pairs or those with compressed co-occurrence involve species of different
 67 temperature groupings.

species pair	% lakes	RANGE OVERLAP*	RANGE	RANGE	BRT	CO-OCCURRENCE
	co- occur	Species A vs Species B (ANOSIM)	COMPRESSION† Species A vs AB (PERMDISP)	COMPRESSION Species B vs AB (PERMDISP)	The presence of A predicts B and vice versa	RANKING
Brooktrout/Pike	0.50%	0.138 (0.02%)	ns	ns	Yes	<i>species rarely co- occurring (Fig. 2A)</i>
Brooktrout/Walleye	0.30%	0.150 (0.02%)	ns	ns	Yes	
Brooktrout/SMBass	0.60%	0.193 (0.02%)	ns	ns	Yes	
SMBass/LongnoseSucker	3.50%	0.203 (0.02%)	ns	ns		
Brooktrout/Burbot	1.40%	0.293 (0.02%)	ns	ns		
Brooktrout/Brownbullhead	1.08%	0.361 (0.02%)	ns	ns		
Brooktrout/LakeTrout	4.80%	0.381 (0.02%)	ns	ns	Yes	
LMBass/Whitefish	3.10%	0.448 (0.02%)	ns	ns		
Brooktrout/LMBass	0.15%	0.565 (0.02%)	NA^	NA^		
LakeTrout/LMBass	4.50%	0.641 (0.02%)	ns	ns		
LakeTrout/Pike	27.90%	0.318 (0.02%)	ns	compressed (p<0.001)	Yes	<i>one or both</i>
Pike/Rockbass	36.40%	0.383 (0.02%)	ns	compressed p=0.049)		<i>species occur in a</i>
Pike/SMBass	38.20%	0.389 (0.02%)	ns	compressed (p=0.008)		<i>compressed</i>
LakeTrout/Pumpkinseed	12.20%	0.475 (0.02%)	ns	compressed (p<0.001)		<i>range of</i>
LakeTrout/Walleye	23.70%	0.287 (0.02%)	compressed (p=0.01)	compressed (p<0.001)	Yes	<i>environments</i>

Rockbass/Whitefish	23.30%	0.305 (0.02%)	compressed ($p < 0.001$)	compressed ($p = 0.006$)		<i>when co-</i>
Burbot/Pumpkinseed	9.70%	0.348 (0.02%)	compressed ($p = 0.01$)	compressed ($p < 0.001$)		<i>occurring,</i>
Whitefish/Pumpkinseed	10.50%	0.363 (0.02%)	compressed ($p = 0.021$)	compressed ($p < 0.001$)		<i>compared to</i>
LakeTrout/ShrtnoseRedhorse	3.70%	0.428 (0.02%)	compressed ($p = 0.003$)	compressed ($p = 0.017$)		<i>when found alone</i> (Fig. 2B)
Pike/Whitesucker	71.50%	0.223 (0.10%)	ns	ns		<i>no compression</i>
Pumpkinseed/Walleye	7.90%	0.276 (0.02%)	ns	ns		<i>of environmental</i>
Rockbass/Walleye	36.50%	0.326 (0.02%)	ns	ns		<i>lake conditions</i>
LMBass/Walleye	12.20%	0.376 (0.02%)	ns	ns		<i>when together</i>
SMBass/Walleye	39.60%	0.407 (0.02%)	ns	ns		<i>compared to</i>
SMBass/Whitesucker	45.10%	0.448 (0.02%)	ns	ns		<i>when alone (Fig.</i>
Whitesucker/Rockbass	41.14%	0.428 (0.02%)	ns	ns		<i>2C)</i>
Brownbullhead/Whitefish	6.30%	0.425 (0.02%)	ns	ns		
LMBass/Whitesucker	12.90%	0.448 (0.02%)	ns	ns		
LongnoseSucker/Pike	7.60%	0.030 (3.2%)	expanded ($p < 0.001$)	ns		<i>species is found</i>
Brooktrout/YellowPerch	2.20%	0.104 (0.3%)	expanded ($p = 0.04$)	ns	Yes	<i>in a wider range</i>
LongnoseSucker/Walleye	6.90%	0.123 (0.5%)	expanded ($p = 0.017$)	ns		<i>of environmental</i>
WhiteSucker/Walleye	69.20%	0.311 (0.02%)	expanded ($p = 0.036$)	ns		<i>conditions when</i> <i>the other is</i> <i>present (Fig. 2D)</i>

68 * Overlap refers to the degree of overlap in the multivariate environmental conditions of lakes where Species A and

69 Species B occur without each other - R values closer to 0 indicate greater similarity in multivariate environmental

70 conditions. Numbers in brackets are the percentage of 5000 randomizations for which an R value as high or higher

71 occurred (i.e., a R value equal or greater to the calculated value less than 5% of the time is considered significant).

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73 †Compression refers to a significant change in the multivariate spread or range of environmental conditions, with

74 ‘compressed’ meaning that the range is NARROWER when the species is in the presence of the other species compared to

75 when alone; ‘expanded’ means the range is WIDER for lakes with both species present compared to when the species is

76 alone

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78 ^ The PERMDISP test is not applicable (NA) in this case, because only 1 lake contained both species.

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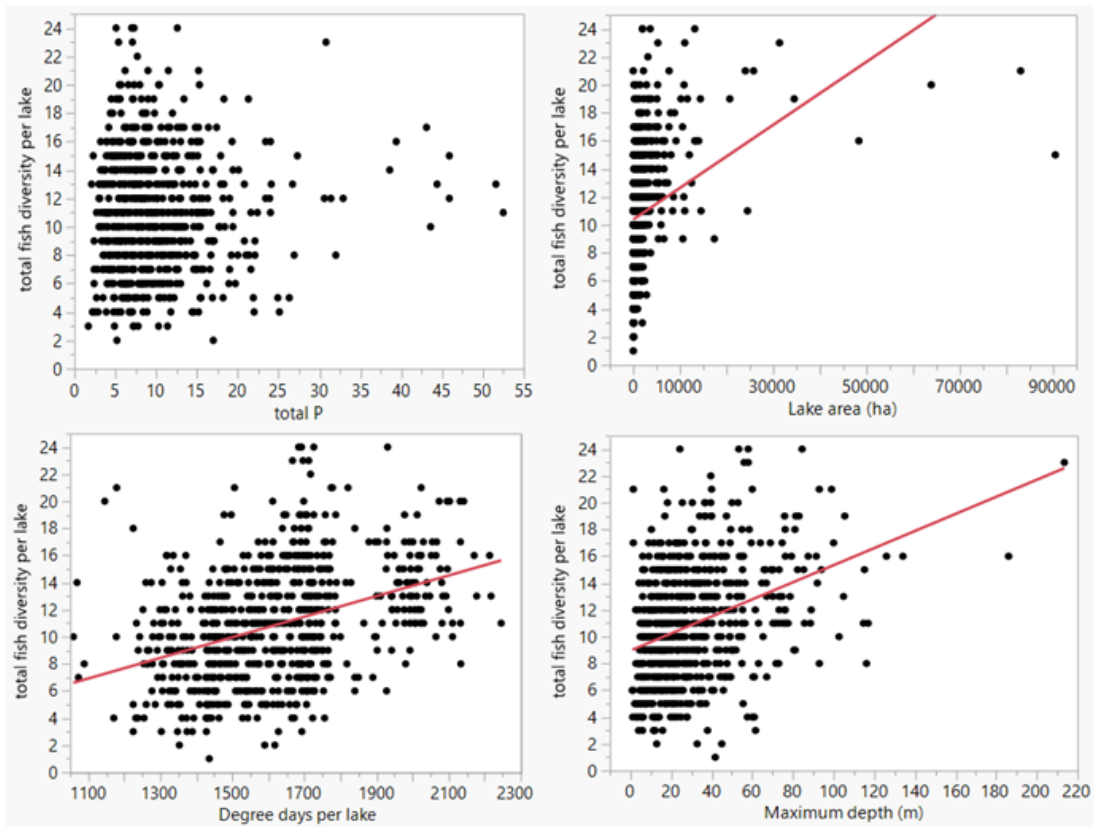
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82 Supplementary Figure 1: Bivariate relationships (with linear fits where significant) between total fish

83 diversity and major explanatory drivers of fish occurrences in lakes. These graphs reveal the potential

84 influence of each explanatory factor on fish.



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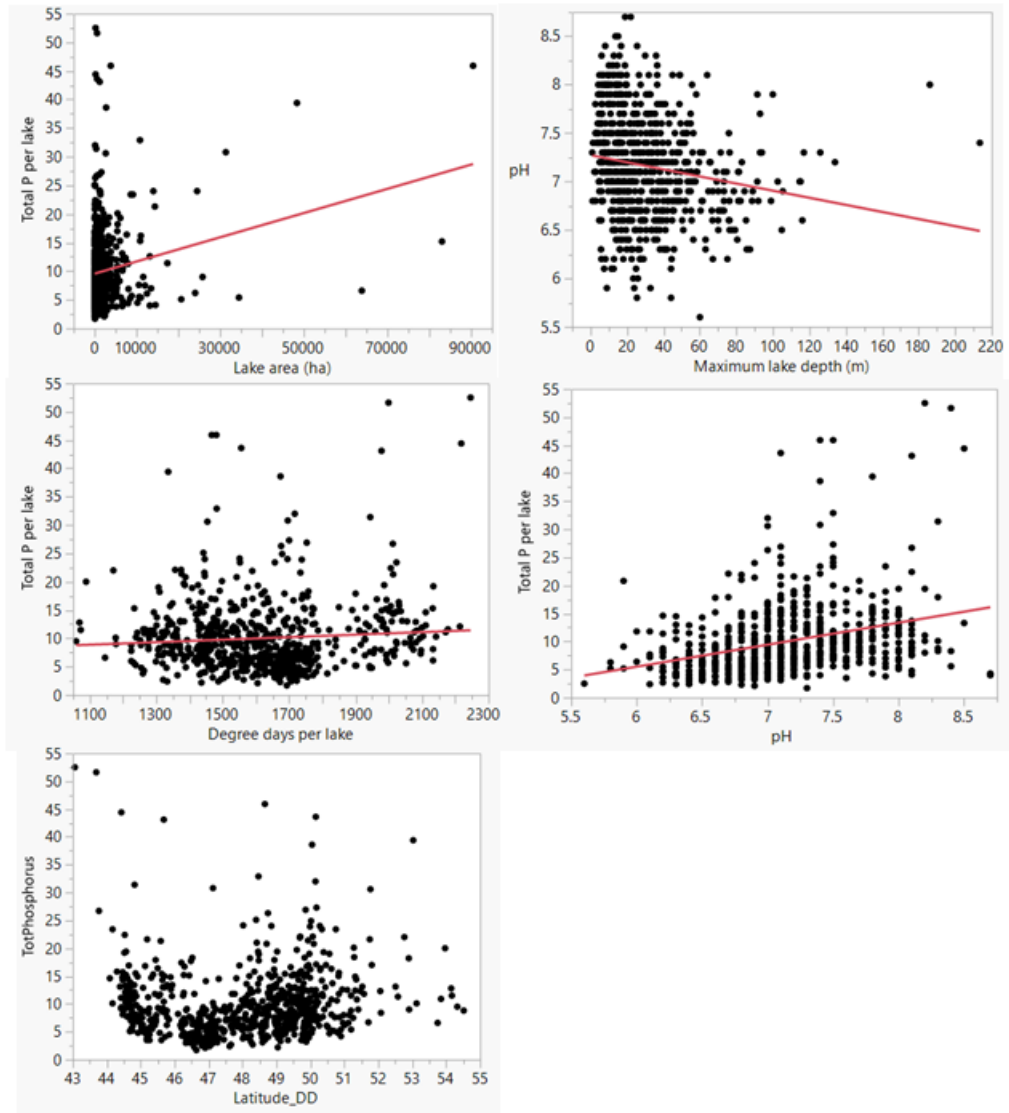
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90 Supplementary Figure 2: Bivariate relationships between abiotic factors in 721 lakes. All relationships
91 were significant with linear fits shown, except latitude despite the concentration of human populations
92 in the southern regions of Ontario. Again, these graphs reveal how any given combination of bivariate
93 factors can significantly interact in lakes, suggesting that integrative analyses testing all
94 simultaneously should better capture how they affect each other, and by extension, the fish
95 communities found within.

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99 **Supplementary Note 1 – Structural Equation Modeling**

100 Our model construction for the SEM analysis began with the entire presence/absence dataset, for
101 22 fish species in 721 lakes with ten associated environmental measures relating to climate (degree
102 days), lake morphometry (area [ha], maximum depth [m], mean depth [m]), and lake water quality
103 (total P concentration, ph, conductivity, dissolved solids, Secchi depth). To minimize strong co-
104 linearity in our model and maximize parsimony, we screened the correlation matrix among the
105 environmental variables (e.g., see some correlations displayed in Suppl. Fig. 2). This allowed us to
106 reduce the number of environmental variables in the SEM from 10 to 5, with these 5 factors
107 corresponding to our three hypothesized abiotic-based models of influence: climate (degree days), lake
108 morphometry (lake area, maximum lake depth), and water quality (P, pH). All SEM analyses were
109 conducted with R 3.1.2 (R Core Team 2016), using the lavaan package version 0.5-20⁴.

110 For the ‘biotic’ components of the SEM, we broadly classified the 22 large-sized species
111 (average length > 10 cm) into three functional groups based on habitat affinity and foraging strategy as
112 described in Holm et al.¹ – major predators, littoral species, and pelagic species (Suppl. Table 1). We
113 also grouped the 28 small-sized species (mean length < 10 cm) into a single functional group “small
114 prey species”, as described above in the Methods. We tested for influences on total numbers of species
115 per functional group (‘functional diversity’) and also shifts in species composition (see Methods). For
116 the latter (composition shifts), we identified which functional groups in the SEM were associated with
117 significant species composition changes among lakes using PCA analyses on Hellinger transformed
118 fish data (see Suppl. Fig. 3 and Methods for details). That is to say, which species were significantly
119 associated with the presence of other species. Finally, to satisfy the SEM assumption of multi-
120 normality we log-transformed variables with skewed distributions.

121 Our SEM analysis was based on the recognized combined importance of abiotic and biotic
122 factors in shaping fish diversity in freshwater lakes⁵. Starting from our initial model that incorporated
123 all levels of interaction among degree days, measures of lake morphometry and water quality, and
124 numbers of species per major functional group, we used modification indices⁴ to guide decisions about
125 adding missing paths to the model. We used the “modindices” function from the lavaan package⁴,
126 which provides a list of all missing linkages in the model and the expected effect of additions on
127 model fit (Maximum likelihood estimator [ML_{EST}]). We used the modification indices incrementally,
128 adding one path at a time, until no modification indices were higher than 2. After this process, we
129 scanned path regressions and incrementally pruned the less significant ones until there was no more
130 gain in model fit (decreasing ML_{EST} ⁶). Our final model converged after 171 iterations ($n = 648$,
131 $ML_{EST} = 4.91$, Degree of freedom = 13, $P=0,977$).

132 Our SEM analysis focused on the hypothesis that an integrated model capturing abiotic and
133 biotic interactions within and among functional groups would best capture the regulation of species
134 richness⁵. Yet, there is also evidence in the fish literature of more singular models shaping the
135 regulation of richness, specifically (I) top-down regulation by major predators, (II) bottom-up models
136 where richness in lower levels of the food chain influences diversity levels above, and (III) abiotic-
137 based models where climate and resources primarily affect species richness at all trophic levels. To
138 further disentangle the relative contribution of abiotic and biotic associations, we decomposed our
139 ‘Integrative’ model into its different components each representing a contrasting hypothesis on the
140 regulation of fish diversity (i.e., no biotic associations between functional groups, top-down
141 associations only, bottom-up associations only).

142 To make our models comparable we used the same variable structure for all candidate models
143 but altered the direction of effects among functional groups. The models of ‘Top-down’ effects of

144 consumers on prey (Suppl. Fig. 4b) and the ‘Bottom-up’ effects of prey on consumers (Suppl Fig. 4c)
145 were started from an initial model including all possible top-down or bottom-up associations among
146 fish functional groups. The abiotic factors were then added using modification indices and the less
147 significant regression paths were pruned (using a similar approach as described above). The model
148 without biotic associations (thereafter ‘Abiotic only’ model) included no effects (arrows) between
149 functional groups so that only direct effects of environmental factors were possible (Suppl Fig. 4a).
150 We compared the four models using Akaike Information Criterion (see Suppl. Table 5 and Suppl. Fig.
151 4). By examining our main ‘Integrative’ model against its reduced components, we were interested in
152 contrasting the magnitude of each factor in the SEM (effect size) versus its contribution to the overall
153 fit of the model (e.g., some factors might have large effect size, but not contribute much to our overall
154 understanding of the system).

155 Results from our model comparison indicate that the model including only abiotic factors
156 performed the weakest, suggesting that despite being important drivers in terms of effect size, abiotic
157 factors are insufficient to explain the system (i.e., lowest model fit, Suppl. Table 5). That is to say,
158 ignoring biotic associations among fish functional groups led to a significant decline in analytical
159 power (Suppl. Table 5). Alternatively, top-down factors had smaller effect sizes compared to some
160 abiotic factors yet strongly contributed to model fit. This contribution centered largely on two top-
161 down associations: from ‘predator PCA1’ to ‘small prey richness’ and from ‘predator richness’ to
162 ‘pelagic richness’. Removing either of those associations from the ‘Top-Down’ model lead to a jump
163 in the Maximum Likelihood estimator of 40 and 30 points respectively, sending the model at the
164 bottom of the ranking in the AICc comparisons. Overall, the ‘Integrative’ model balanced the factors
165 with the greatest effect sizes while maximizing model fit, including the described top down

166 associations, the strong effects of degree days and lake morphometry, and the bottom-up effect of
 167 littoral richness on predator richness (Fig. 1).

168 The R code for the SEM analysis is provided in full, in Supplementary Note 4.

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170 Supplementary Table 5. Comparison of our best model with more simple alternative hypotheses.

Model	K	AICc	Δ AICc	Difference support
Top-down	48	40119.91	0.00	
Integrative	50	40120.85	0.94	None
Bottom-up	52	40130.71	10.80	Strong
Abiotic	57	40135.64	15.73	Strong

171 Critical values for model difference support are: **none** Δ AICc < 2; **substantial** $4 < \Delta$ AICc < 7; and
 172 **strong**: Δ AICc > 10⁷.

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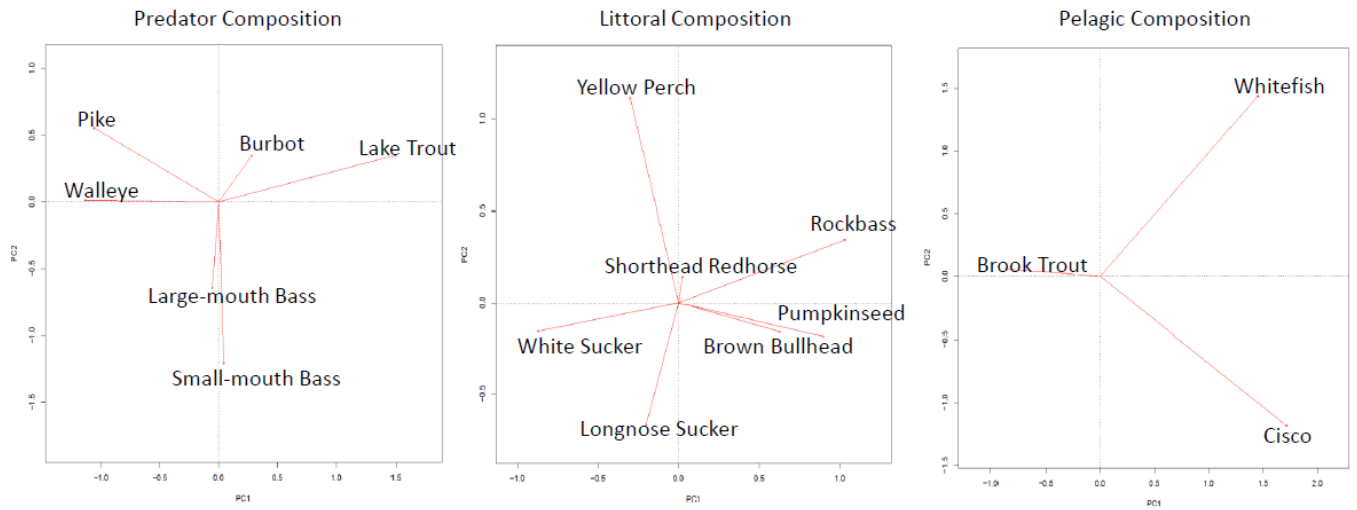
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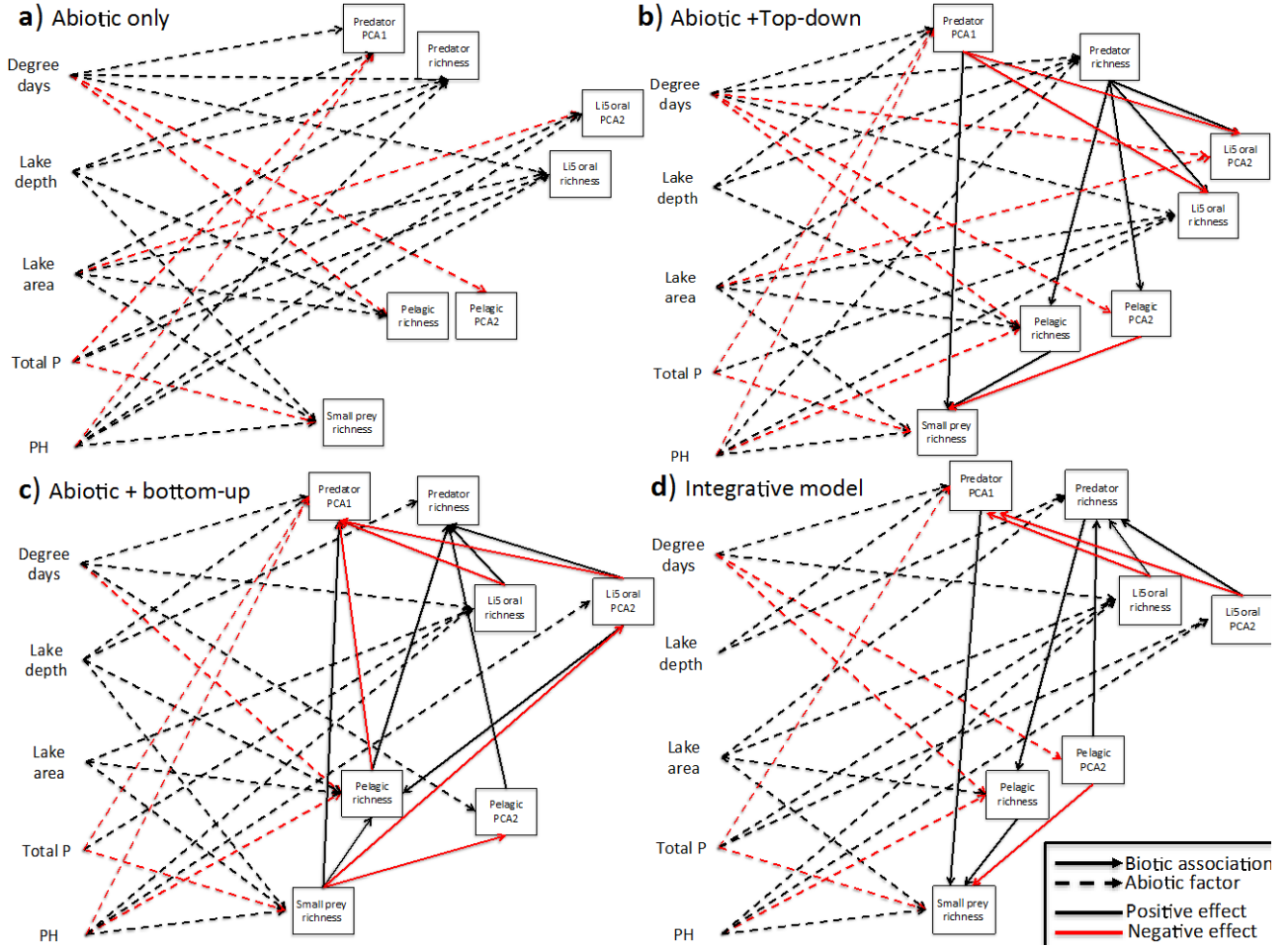
183

184 Supplementary Figure 3. PCA analysis of Hellinger transformed species associated with significant
185 shifts in species composition in the SEM model (Fig. 1), divided into three functional groups based on
186 broad lake habitat associations. The mechanisms underlying the association shifts cannot be
187 determined by the SEM, and may possibly reflect species interactions, habitat affinity not relating to
188 species interactions, or some combination of the two. All subsequent analyses tested the likelihood of
189 interactions versus habitat in explaining these association shifts, targeting these 16 species with the
190 greatest vector-length separation out of the initial 22 species used in the study.



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201 Supplementary Figure 4. Visual representation of all the alternative hypotheses considered in this
202 study to explain fish functional group richness and compositional shifts. Each model was built
203 iteratively with modifications indices⁴ starting from an initial model which structure was constrained
204 by specific rules as a function of the hypothesis tested (see Supplementary Notes 1). The a) Abiotic
205 only model represents the effects of abiotic factors of fish functional group richness and composition
206 in the absence of any biotic associations ($n = 648$, $ML_{EST} = 3.03$, Degree of freedom = 6, $P=0.805$), b)
207 Top-down model represents the effects of abiotic factors interacting with top-down associations on
208 richness and composition ($n = 648$, $ML_{EST} = 8.66$, Degree of freedom = 15, $P=0.895$), c) Bottom-up
209 model represents the effects of abiotic factors interacting with bottom-up associations on richness and
210 composition ($n = 648$, $ML_{EST} = 10.04$, Degree of freedom = 11, $P=0.527$), and d) Integrative model
211 represents the combined effects of abiotic factors, bottom-up and top-down associations on richness
212 and composition ($n = 648$, $ML_{EST} = 4.91$, Degree of freedom = 13, $P=0.977$).



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223 **Supplementary Note 2 - Individual fish species boosted regression tree (BRT) models**

224 We developed BRT models⁸ to assess the relative importance of 22 biotic and abiotic predictor
225 variables for explaining the variation in fish species presence in lakes for 14 of the fish species, i.e.
226 brown bullhead, pumpkin seed, walleye, lake whitefish, pike, cisco, yellow perch, rock bass, white
227 sucker, large-mouth bass, lake trout, small-mouth bass, burbot, and brook trout. In the model
228 developed for each species, we included presence data for the other 13 species to look for possible
229 species relationships, as well as a biotic predictor of the species richness of small prey and seven
230 abiotic characteristics of the lakes, i.e. degree days, pH, phosphorous levels, minimum depth of water,
231 maximum depth of water, area of the lake, and Secchi depth.

232 All models were fitted in R statistical computing program V. 3.2.1 using the gbm package
233 version 2.1.1⁹. BRT models are simple classification or rule-based models that partition observations
234 into groups based on similar values in the response variables using binary splits that are based on
235 predictor variables¹⁰. The boosting algorithm then iteratively develops a final model using a forward
236 stage-wise approach that progressively adds trees to the model. BRT models are advantageous for
237 analyzing our data because they can be used to assess different types of predictor variables in the same
238 model, which fits our questions well¹¹. We used a tree complexity rate of two (which allows for
239 possible two way interactions) and a learning rate of 0.05 to allow for a suitable predictive
240 performance¹⁰. Because response variables were based on presence and absence data from surveys
241 conducted within each of the lakes we used a Bernoulli distribution. We show the relative influence of
242 each of the 22 possible predictors of variation in the presence of the 14 fish species in Suppl. Table 3.
243 We also show the number of trees included in the final model and estimates of the correlation between
244 the observed and predicted response variables. Variation in the presence of the different fish species
245 predicted by each BRT model is shown in Suppl. Figure 5, where values are predicted for each

246 variable by keeping all other variables at mean values.

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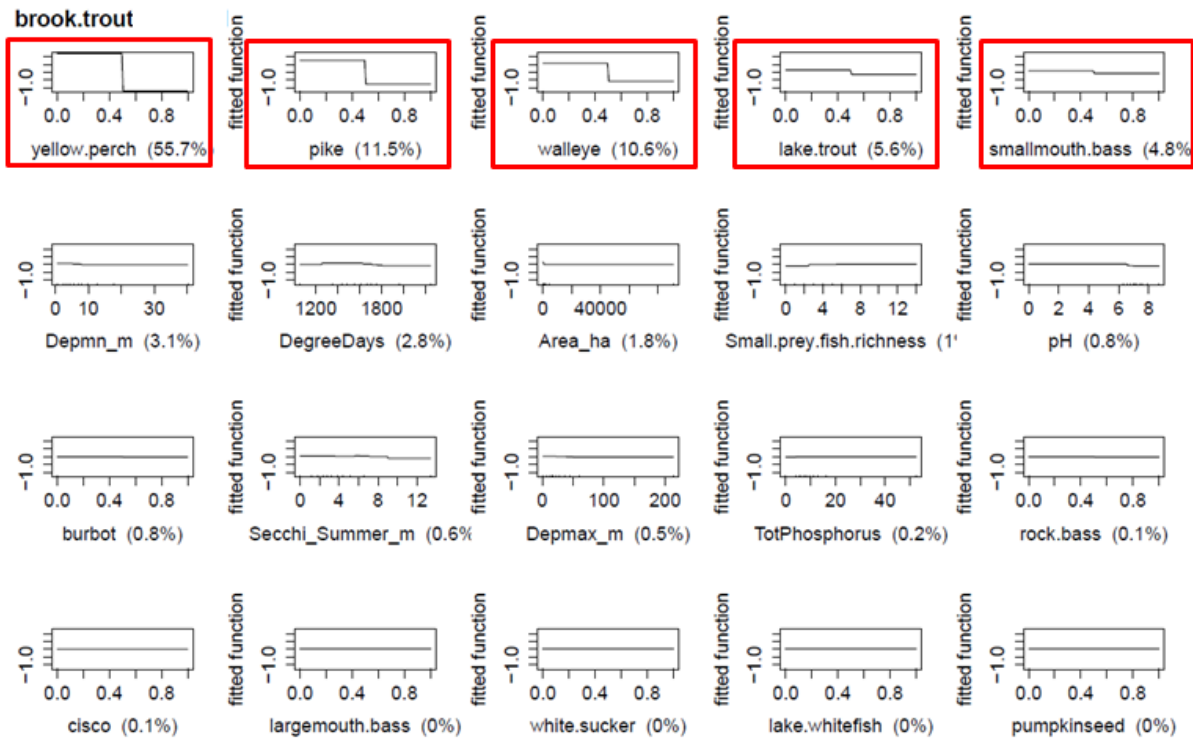
267

268 Supplementary Figure 5: The BRT tests if the occurrence of a species was associated with significant
 269 negative changes in the occurrence of the target species, versus all abiotic factors used in our analyses.
 270 Only two target species – brook trout (A) and lake trout (B) showed negative relationships versus the
 271 occurrence of other species, indicated by the red boxes. These data are the same as shown
 272 in Suppl. Fig. 3. Negative relationships are indicated by declining slopes, with the x-axis ranging from
 273 0 (absent) the 1 (present). For the remaining 14 species tested with the BRT (Suppl. Fig. 3), all major
 274 drivers of occurrence were abiotic.

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276 (A)

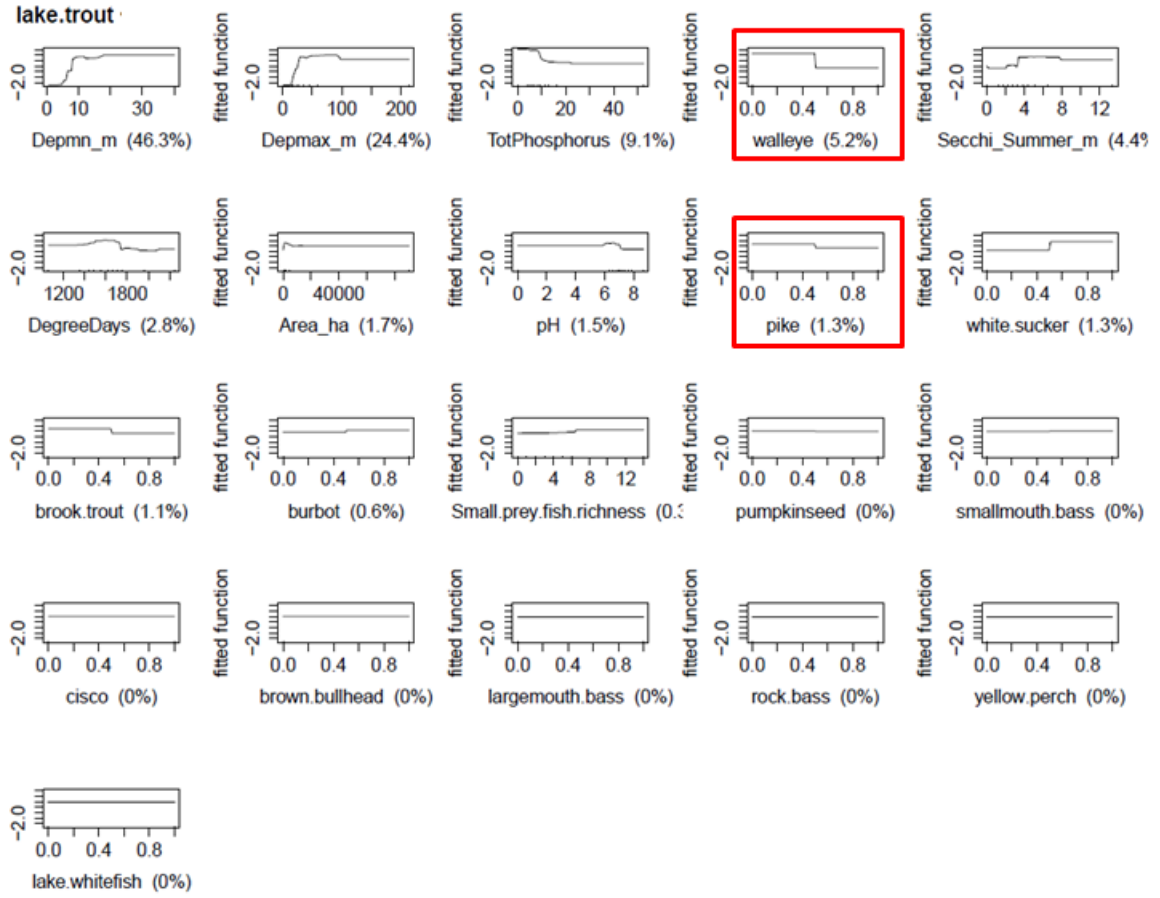
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280 (B)



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292 **Supplementary Note 3 - Multivariate analyses of environmental conditions for species pair**
293 **combinations**

294 For all species pairs, we calculated the Euclidian distance between three lake combinations
295 (Species A only, Species B only, both) based on seven of the measured abiotic/biogeographic factors
296 that were minimally inter-correlated. These factors are lake area, maximum depth of the lake, summer
297 Secchi depth, conductivity, pH, total phosphorus, and degree days. We eliminated lakes for which one
298 or more of these factors were not measured, leaving 649 of 721 lakes. We log transformed and
299 normalized all variables before the analysis.

300 To determine the overall degree of overlap in environmental space between lakes with species
301 A only versus lakes with species B only, we used an ANOSIM analysis¹². This analysis tests the null
302 hypothesis that there are no differences in terms of multivariate environmental conditions, and is
303 equivalent to a non-parametric MANOVA, with permutation to test for significance¹². The test
304 computes the R statistic, which ranges from 0 (complete overlap in environmental space) to 1
305 (complete separation in environmental space). As recommended by Clarke et al. (2014)¹², we
306 considered the pairwise test between lakes with species A only versus lakes with species B only after
307 confirming that the global ANOSIM test (considering all lake types: species A only, species B only,
308 both A and B, and neither) was significant. This was true in all cases. In all cases we used 5,000
309 random permutations to test for significant differences.

310 If a difference in multivariate environmental conditions was found via the ANOSIM analysis
311 above, it could be due to a shift in location (multivariate average of environmental conditions) and/or a
312 difference in dispersion (multivariate variation or range of environmental conditions). To test for the
313 latter possibility, we determined for each species pair whether lakes with both species co-occurring
314 had a reduced range of environmental conditions compared to lakes with species A only or species B

315 only. This could happen, for example, if a subordinate competitor or prey species can only co-occur
 316 with a dominant competitor or predator in deep and large lakes, with those lake conditions allowing
 317 the subordinate more possibilities to persist. Relatedly, we would predict that the subordinate could
 318 occur in smaller and shallower lakes, if the dominant species is absent from those lakes. To do this, we
 319 used a test of difference in multivariate dispersions (PERMDISP¹³). This is a multivariate extension of
 320 Levene's test, measuring the average distance from each lake to the centroid of the group of lakes in
 321 the same category, and comparing this average distance between different categories¹⁴. If coexistence
 322 of species A and B is limited to a smaller range of environmental conditions than lakes inhabited by
 323 species A without species B, then we would expect to find a significant decline in multivariate
 324 dispersion of lakes with both species compared to lakes with species A only. This test also uses
 325 permutation to determine significance of the test. We again set the number of permutations to 5,000.
 326 We note that for species pairs where the number of lakes containing both species was very small, our
 327 power to detect differences in multivariate dispersion was likely limited. For both the ANOSIM and
 328 PERMDISP analyses we used the program PRIMER¹⁴. These tests are also available in the 'vegan'
 329 package in R, but R does not allow pairwise tests in the ANOSIM function.

330 To visualize the similarity among lakes in terms of their environmental conditions, we used
 331 PCA to ordinate all lakes in environmental space. We used the 'vegan' package in R to create
 332 ordination plots, with vectors to visualize the correlation of each environmental factor with the first
 333 two PCA axes.

334

335 **Supplementary Note 4 – R code for SEM analysis**

336

```

337 #.....#
338 #.....BSM project (SEM).....#
339 #.....#
  
```

```

340 #.....
341 #
342 #... Collaborators: Andrew MacDougall et al#
343 #
344 #... Author of the script: Eric Harvey (http://ericharvey.weebly.com/)
345 #
346 #... Creation date: January 05 2016
347 #
348 #... Latest edits: November 07 2016
349 #
350 #.....
351 #
352 #... Clear variables
353 rm(list=ls())
354 #... Directories paths
355 datapath = "~/Documents/Research/1.Projects/3.BSM_project/3.Analysis/"
356 #...Load needed packages
357 library(vegan)
358 library(ade4)
359 library(lavaan)
360 library(corrplot)
361 #... Load the data
362 setwd(datapath)
363 presence.dat = read.delim("BSM_presence.txt")
364 presence.dat = na.omit(presence.dat)
365 #... Identify species of interest for each functional group
366 predator.of.interest =
367 c("walleye","lake.trout","pike","smallmouth.bass","largemouth.bass","burbot")
368 benthopelagic.species = c("cisco","brook.trout","round.whitefish","lake.whitefish")
369 bentholitoral.species =
370 c("shorthead.redhorse","silver.redhorse","goldeye","longnose.sucker","white.sucker","br
371 own.bullhead","channel.catfish","yellow.bullhead","black.crappie","bluegill","pumpkinsee
372 d","rock.bass","yellow.perch","mooneye","fallfish")
373 smallprey.species = c("blackchin.shiner","blacknose.shiner","bluntnose.minnow",
374 "brook.stickleback","central.mudminnow","common.shiner",
375 "creek.chub","deepwater.sculpin","eastern.blacknose.dace",
376 "emerald.shiner","fathead.minnow","finescale.dace",
377 "golden.shiner","iowa.darter","johnny.darter","lake.chub",
378 "logperch","longnose.dace","mimic.shiner","Mottled.Sculpin","ninespine.stickleback",
379 "northern.redbelly.dace","Pearl.dace","sculpin.spp","Slimy.Sculpin",
380 "spoonhead.sculpin","spottail.shiner","trout.perch")
381 #... Generate presence/absence matrix with species of interest
382 pred.comp.mat = presence.dat[,predator.of.interest]
383 pelagic.comp.mat = presence.dat[,benthopelagic.species]
384 littoral.comp.mat = presence.dat[,bentholitoral.species]
385 smallprey.comp.mat = presence.dat[,smallprey.species]
386 #####
387 #####

```

```

388 #...Extract variables for SEM
389 #.....Ordinations analyses for compositional shifts
390 #.....Change in species composition with PCA
391 pred.PCA = rda(decostand(pred.comp.mat,"hell"))
392 pelagic.PCA = rda(decostand(pelagic.comp.mat,"hell"))
393 litoral.PCA = rda(decostand(litoral.comp.mat,"hell"))
394 #.....PCA results
395 pred.PCA
396 cumsum(pred.PCA$CA$eig/sum(pred.PCA$CA$eig)*100)
397 pelagic.PCA
398 cumsum(pelagic.PCA$CA$eig/sum(pelagic.PCA$CA$eig)*100)
399 litoral.PCA
400 cumsum(litoral.PCA$CA$eig/sum(litoral.PCA$CA$eig)*100)
401 #.....PCA diagnosis plots
402 source ('http://www.davidzeleny.net/anadat-r/doku.php/en:numecolr:cleanplot.pca?
403 do=export_code&codeblock=0')
404 cleanplot.pca(pred.PCA)
405 cleanplot.pca(pelagic.PCA)
406 cleanplot.pca(litoral.PCA)
407 dev.off()
408 #define "evplot" function first:
409 source ("http://www.davidzeleny.net/anadat-r/doku.php/en:numecolr:evplot?
410 do=export_code&codeblock=0")
411 # select the data frame with eigenvalues of particular axes:
412 ev.pred <- pred.PCA$CA$eig
413 ev.pelagic <- pelagic.PCA$CA$eig
414 ev.litoral <- litoral.PCA$CA$eig
415 # calculate axis-importance and draw the barplots:
416 evplot (ev.pred)
417 evplot (ev.pelagic)
418 evplot (ev.litoral)
419 dev.off()
420 #.....Create variables for SEM
421 #.....Axes representing changes in species composition
422 pred.PCA1 = scores(pred.PCA)$sites[,1]
423 pred.PCA2 = scores(pred.PCA)$sites[,2]
424 pelagic.PCA1 = scores(pelagic.PCA)$sites[,1]
425 pelagic.PCA2 = scores(pelagic.PCA)$sites[,2]
426 litoral.PCA1 = scores(litoral.PCA,choices=c(1:4))$sites[,1]
427 litoral.PCA2 = scores(litoral.PCA,choices=c(1:4))$sites[,2]
428 #.....Species richness per functional group
429 pred.richness = specnumber(pred.comp.mat)
430 small.prey.richness = specnumber(small.prey.comp.mat)
431 pelagic.richness = specnumber(pelagic.comp.mat)
432 litoral.richness = specnumber(litoral.comp.mat)
433 #.....Environmental variables
434 #.....Geo
435 Latitude = presence.dat$Latitude_DD

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```

436 Longitude = presence.dat$Longitude_DD
437 DD = presence.dat$DegreeDays
438 #.....Morphology
439 Area = presence.dat$Area_ha
440 Perimeter = presence.dat$Perimeter_km
441 Depth = presence.dat$Depmn_m
442 Depth.max = presence.dat$Depmax_m
443 #.....Productivity
444 Secchi = presence.dat$Secchi_Summer_m
445 Conductivity = presence.dat$Conductivity..uS.cm.s.
446 TDSolids = presence.dat$TDSolids_mgL
447 TotPhosphorus = presence.dat$TotPhosphorus
448 PH = presence.dat$pH
449 #####
450 #####
451 #...Correlations among variables
452 #Among biotic factors
453 biotic.cor =
454 cor(cbind(litoral.PCA1,litoral.PCA2,pelagic.PCA1,pelagic.PCA2,pred.PCA1,pred.PCA2,
455 pred.richness,litoral.richness,pelagic.richness,small.prey.richness))
456 corrrplot(biotic.cor,method="number",type="full")
457 #Among environmental variables
458 env.cor =
459 cor(na.omit(cbind(Latitude,Longitude,DD,Area,Perimeter,Depth,Secchi,Conductivity,TD
460 Solids,TotPhosphorus,PH)))
461 corrrplot(env.cor,method="number",type="full")
462 #####
463 #####
464 #...Data frame for SEM ananalysis
465 #.....variables are log transformed if necessary
466 #.....variables need to be all at the same variance scale - to achieve this some are
467 divided/multiplied by a factor of 10 or 100
468 data.mod = data.frame(DD/100,Latitude,
469 Area/1000,Perimeter/100,Depth,
470 Secchi,TotPhosphorus,Conductivity/10,PH,TDSolids/10,
471 as.numeric(pred.richness),as.numeric(small.prey.richness),pelagic.richness,litoral.richne
472 ss,
473 pred.PCA1*100,pred.PCA2*100,pelagic.PCA1*100,pelagic.PCA2*100,litoral.PCA1*100,
474 litoral.PCA2*100)
475 colnames(data.mod) = c("DD","Latitude",
476 "Area","Perimeter","Depth",
477 "Secchi","TotPhosphorus","Conductivity","PH","TDSolids",
478 "pred.richness","small.prey.richness","pelagic.richness","litoral.richness",
479 "pred.PCA1","pred.PCA2","pelagic.PCA1","pelagic.PCA2","litoral.PCA1","litoral.PCA2")
480 #.....Correlation matrix of the data
481 round(cov2cor(cov(data.mod)),3)
482 #####
483 #####

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484 #...Structural equation models
485 #Each model is built by starting from a simplified model from which paths are
486 sequentially added based on the modification indices until no modification indices are
487 >2.
488 #The initial simplified model for each alternative model is built as follow:
489 #Model 1 - Abiotic factors only
490 #Model 2 - Top-down regulation: Initial model was only top-down interactions and then
491 abiotic factors are added sequentially
492 #Model 3 - Bottom-up regulation: Initial model was only bottom-up interactions and then
493 abiotic factors are added sequentially
494 #Model 4 - Integrative model (main model from the paper): Abiotic + Top-down +
495 Bottom-up (based on our initial meta-model - see Supp.Mat)
496 #If the modification indice proposed to add a path that is forbidden according to the
497 model (e.g., between trophic groups in the abiotic model), the proposition is ignored and
498 the next best proposition is implemented.
499 #If the modification indice value of adding a path between an abiotic factor and a
500 functional group is the same as the value of adding a path between said trophic group
501 and another trophic group, then it was interpreted as a signal that the effect of said
502 abiotic factor might be mediated by a trophic interaction
503 library(semTools)
504 #.....Model 1 --> Model with only abiotic factors
505 BSM.mod1 <- "
506 # Regression
507 small.prey.richness ~ DD + PH + Area + TotPhosphorus + Depth
508 pelagic.richness ~ DD + PH + Depth + Area + TotPhosphorus
509 pelagic.PCA2 ~ DD + TotPhosphorus
510 littoral.richness ~ DD + Area + TotPhosphorus + PH + Depth
511 littoral.PCA2 ~ DD + Area + PH + TotPhosphorus
512 pred.richness ~ PH + Depth + Area + DD
513 pred.PCA1 ~ DD + PH + TotPhosphorus + Depth
514 "
515 BSM.mod1.fit = sem(BSM.mod1,data=data.mod) #run the SEM
516 BSM.mod1.fit #visualize fitting results
517 summary(BSM.mod1.fit,stand=T,rsq=T) #visualize path coefficients
518 modindices(BSM.mod1.fit) #modification indices
519 # To inspect model estimates and parameters
520 varTable(BSM.mod1.fit) #Variance table
521 inspect(BSM.mod1.fit,"cov.ov") #co-variance matrix
522 inspect(BSM.mod1.fit,"est") #regression estimates
523 #
524 #.....Model 2 Top-down regulation --> paths only from higher to lower trophic groups +
525 abiotic factors
526 #Notice that Depth - a factor that was very important in Model 1 is now mainly important
527 for top-predators - it suggests that the effect of depth on other groups is mediated by
528 top-predators
529 BSM.mod2 <- "
530 # Regression
531 small.prey.richness ~ pred.PCA1 + pred.richness + pelagic.richness + pelagic.PCA2 +

```

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532 litoral.richness + PH + Area + TotPhosphorus
533 pelagic.richness ~ pred.richness + DD + Depth + PH + Area + TotPhosphorus +
534 pred.PCA1
535 pelagic.PCA2 ~ pred.richness + litoral.richness + DD
536 litoral.richness ~ pred.richness + pred.PCA1 + DD + TotPhosphorus + Area + PH
537 litoral.PCA2 ~ pred.richness + pred.PCA1 + DD + Area + pelagic.PCA2
538 pred.richness ~ PH + Depth + Area + DD
539 pred.PCA1 ~ DD + PH + TotPhosphorus + Depth
540 # Residual correlations
541 pred.richness ~~ pred.PCA1
542 pelagic.richness ~~ pelagic.PCA2
543 litoral.richness ~~ litoral.PCA2
544 "
545 BSM.mod2.fit = sem(BSM.mod2,data=data.mod)
546 BSM.mod2.fit
547 summary(BSM.mod2.fit,stand=T,rsq=T)
548 modindices(BSM.mod2.fit)
549 # To inspect model estimates and parameters
550 varTable(BSM.mod2.fit) #Variance table
551 inspect(BSM.mod2.fit,"cov.ov") #co-variance matrix
552 inspect(BSM.mod2.fit,"est") #regression estimates
553 #
554 #.....Model 3 Bottom-up regulation --> Ppaths only from lower to higher trohic groups+
555 abiotic factors
556 BSM.mod3 <- "
557 # Regression
558 small.prey.richness ~ PH + Depth + Area + TotPhosphorus
559 pelagic.richness ~ small.prey.richness + DD + Depth + PH + Area + litoral.PCA2
560 pelagic.PCA2 ~ small.prey.richness + litoral.richness + Area + DD + TotPhosphorus
561 litoral.richness ~ small.prey.richness + DD + TotPhosphorus + Area + PH
562 litoral.PCA2 ~ small.prey.richness + DD + PH
563 pred.richness ~ small.prey.richness + pelagic.richness + pelagic.PCA2 + litoral.richness
564 + litoral.PCA2 + Depth + DD + Area + TotPhosphorus
565 pred.PCA1 ~ small.prey.richness + pelagic.richness + pelagic.PCA2 + litoral.richness +
566 litoral.PCA2 + Depth + DD + TotPhosphorus + PH
567 # Residual correlations
568 pred.richness ~~ pred.PCA1
569 pelagic.richness ~~ pelagic.PCA2
570 litoral.richness ~~ litoral.PCA2
571 pelagic.richness ~~ litoral.richness
572 "
573 BSM.mod3.fit = sem(BSM.mod3,data=data.mod)
574 BSM.mod3.fit
575 summary(BSM.mod3.fit,stand=T,rsq=T)
576 modindices(BSM.mod3.fit)
577 # To inspect model estimates and parameters
578 varTable(BSM.mod3.fit) #Variance table
579 inspect(BSM.mod3.fit,"cov.ov") #co-variance matrix

```



```

580 inspect(BSM.mod3.fit,"est") #regression estimates
581 #
582 #.....Model 4 Integrative model (main model from our meta-model) --> Abiotic factors +
583 top-down + bottom-up
584 BSM.mod4 <- "
585 # Regression
586 small.prey.richness ~ PH + Area + TotPhosphorus + pred.PCA1 + pelagic.richness +
587 pelagic.PCA2
588 pelagic.richness ~ DD + PH + Depth + Area + TotPhosphorus + pred.richness
589 pelagic.PCA2 ~ DD + TotPhosphorus + pred.richness
590 littoral.richness ~ DD + Area + TotPhosphorus + PH + pred.richness
591 littoral.PCA2 ~ DD + Area + PH + TotPhosphorus
592 pred.richness ~ PH + Depth + littoral.richness + littoral.PCA2 + pelagic.richness +
593 pelagic.PCA2 + TotPhosphorus + small.prey.richness + Area
594 pred.PCA1 ~ DD + PH + TotPhosphorus + Depth + littoral.richness + littoral.PCA2
595 # Residual correlations
596 littoral.richness ~~ littoral.PCA2
597 pelagic.richness ~~ pelagic.PCA2
598 pred.richness ~~ pred.PCA1
599 littoral.richness ~~ small.prey.richness
600 "
601 BSM.mod4.fit = sem(BSM.mod4,data=data.mod)
602 BSM.mod4.fit
603 summary(BSM.mod4.fit,stand=T,rsq=T)
604 modindices(BSM.mod4.fit) #modification indices
605 resid(BSM.mod4.fit,type="raw")
606 # To inspect model estimates and parameters
607 varTable(BSM.mod4.fit) #Variance table
608 inspect(BSM.mod4.fit,"cov.ov") #co-variance matrix
609 inspect(BSM.mod4.fit,"est") #regression estimates
610 #
611 #####
612 #Model goodness of fit and comparison
613 #####
614 #AIC comparisons
615 library(AICcmodavg)
616 aictab(list(BSM.mod1.fit,BSM.mod2.fit,BSM.mod3.fit,BSM.mod4.fit),c("Abiotic","TD","BU
617 ","Integrative"))
618 # Model selection based on AICc:
619 #
620 # K AICc Delta_AICc AICcWt Cum.Wt LL
621 # TD 48 40119.91 0.00 0.61 0.61 -20008.03
622 # Integrative 50 40120.85 0.94 0.38 1.00 -20006.15
623 # BU 52 40130.71 10.80 0.00 1.00 -20008.72
624 # Abiotic 57 40135.64 15.73 0.00 1.00 -20005.22
625 #After removing 'Predator PCA1 --> 'Small prey richness' from the 'Top-down' model
626 # Model selection based on AICc:
627 #

```

628 # K AICc Delta_AICc AICcWt Cum.Wt LL
 629 # Integrative 50 40120.85 0.00 0.99 0.99 -20006.15
 630 # BU 52 40130.71 9.86 0.01 1.00 -20008.72
 631 # Abiotic 57 40135.64 14.79 0.00 1.00 -20005.22
 632 # TD 47 40157.95 37.10 0.00 1.00 -20028.22
 633 #After removing 'Predator richness' --> 'Pelagic richness' from the 'Top-down model'
 634 # K AICc Delta_AICc AICcWt Cum.Wt LL
 635 # Integrative 50 40120.85 0.00 0.99 0.99 -20006.15
 636 # BU 52 40130.71 9.86 0.01 1.00 -20008.72
 637 # Abiotic 57 40135.64 14.79 0.00 1.00 -20005.22
 638 # TD 47 40150.85 30.00 0.00 1.00 -20024.66
 639 #Observations on AIC comparisons
 640 # 1. Abiotic only is last
 641 # 2. The difference between TD and ALL is not substantial
 642 # 3. The difference between TD/ALL and BU/Abiotic is substantial
 643 # 4. It is important to note that no interactions in the TD model does not indicate
 644 of actual top-down interactions (none of them are negative!) except for one strong topdown
 645 interaction between top-predators and small prey richness
 646 # 5. In the context of observation #4 - It seems that the only reason why BU does not
 647 perform well is because it is lacking the very important top-down effect of top-predators
 648 on small prey richness
 649 # 6. Observation #5 can be verified easily by removing the top-down path from toppredators
 650 to small prey richness in the top-down model ('Predator PCA1' --> 'Small
 651 prey richness':ML increases by 40 points and the P-value = 0.0000, 'Predator richness' --
 652 > 'Pelagic richness': ML increase 33 points) - therefore the good fit of the dop-down
 653 model is driven by this keystone interaction in the model
 654 #Conclusion:
 655 # 1. A mixture of all variables is essential to understand the system. Both bottom-up
 656 and top-down effects are very important, and in many case they seem to mediate the
 657 effect of abiotic factors (e.g., Lake depth).
 658 # 2. Thus, Model 4 is confidently the best model to understand the system because it
 659 integrates all information and the mediated effects in the most parsimonious way
 660 # 3. Model 4 could be analyzed further to gain better insights about the system.
 661 #END
 662 #####

663

664 **Supplementary References**

665

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