SUPPORTING INFORMATION SECTION

Oxidation of Benzene by Fe(III)- and Mn(IV)-Containing Oxides: Stoichiometric Efficiency and Transformation Products

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Text S1: Experimental setup and material preparation

All chemicals used in this study were reagent grade or higher. Solutions were prepared using deionized water (resistivity >18.2 M Ω , Millipore system). Four types of pure minerals were employed in this study, *i.e.*, amorphous ferrihydrite (Fe(OH)_{3(s)}), goethite (α -FeOOH_(s)), pyrolusite (β -MnO_{2(s)}) and silica (SiO₂). Pyrolusite and ferrihydrite were directly obtained from Sigma-Aldrich. Pyrolusite was used without further processing. Ferrihydrite was aged in deionized water buffered at pH 8.0 with 50 mM borate for 2 weeks prior to experiments, with daily pH adjustment to 8.0 by adding 1 mM NaOH. After that, ferrihydrite suspension was centrifuged and the particles were washed three times with deionized water before finally dried with a freeze dry system. Silica obtained as pure sand (ACROS Organics) was rinsed in 0.1 M HClO₄ at a concentration of 300 g/L. After 24 hours, the solution was decanted and replaced. This step was repeated for 3 consecutive days. After that, silica particles were freeze-dried. Goethite was synthesized by aging freshly made ferrihydrite in a concentrated NaOH solution at 70 °C for 60 hours. Two clay materials, nontronite and montmorillonite were used. Five aquifer solids were used, denoted as AWBPH, AFTCS, CAROL, CADOU and AMTAL. The characterization of materials is listed in Table S1 and can be found in a recently published paper.¹.

Text S2: Control of dissolved O₂ concentration in samples

The dissolved O_2 concentration in the solution was adjusted by purging with air, N_2 , or pure O_2 . In most cases, the solution was saturated with air (*i.e.*, with a dissolved O_2 concentration of 250 μ M). In cases where O_2 -free conditions were needed, the suspension was purged with N_2 and kept in a glove box during the experiments. For those samples, the residual O_2 concentration was always less than 3 μ M. For experiments conducted at elevated O_2 concentrations, the solution was purged with pure O_2 in a 1-L volumetric flask without headspace at 4°C and was sealed before raising the temperature back to 23°C. This process yielded a dissolved O_2 concentration of 410 μ M.

Text S3: High-resolution mass spectrometry and its operating conditions

Chromatographic separation was achieved using a Hydro-RP column (150 x 4.6 mm) equipped with a guard column (both Phenomenex). 0.1% Acetic acid (A) and methanol (B) were used as eluents. The percentage of (A) was changed linearly as follows: 0-2 min, 100%; 8 min, 40%; 11 min, 5%; 12 min, 5%, 12.1 min 100%, 18 min, 100%. Injection volume was set to 50 μ L. QTOF analysis was performed in negative ESI-MS mode in both fullscan (80 – 400 m/z) and targeted MS/MS mode to obtain exact mass and structural information of the unknown compound. The source parameters were set to: gas temperature 300°C, gas flow: 12 L min⁻¹, nebulizer: 30 psi, capillary voltage: 3000 V.

Text S4: Nuclear magnetic resonance (NMR) and solid-phase extraction (SPE)

To prepare the sample for nuclear magnetic resonance (NMR), solid-phase extraction (SPE) was used to both separate the unknown compound from benzene and phenol and to concentrate the product. SPE cartridges (BondElut C18; 1g sorbent) were conditioned using 5 mL of heptane, 5 mL acetone and 5 mL methanol followed by 25 mL of DI water adjusted to pH 2 using 1 M H₂SO₄. To produce the unknown, 2 L of a 1mM benzene solution containing 10 mM persulfate were exposed to UV light from a low-pressure Hg lamp. Formation of the unknown compound was monitored by HPLC/UV and the experiments were terminated when its peak area reached a plateau (usually within 5 hours). Unbuffered ultrapure water was used to minimize potential interferences of salts in the NMR analysis. No significant differences between buffered and unbuffered water were observed with respect to the unknown compound. After loading the samples to SPE cartridges (50 mL per cartridge, sample pH was adjusted to pH 2), cartridges were first washed with 5 mL of DI water at pH 2. The unknown compound was then eluted using 10 mL of DI water at pH 7, leaving benzene and phenol on the cartridges for later elution with 10 mL of methanol. This procedure was repeated several times to achieve an overall enrichment factor of 200. The final sample was subjected once more to the SPE procedure described above but using D₂O (pH 2) for washing of the cartridge and CH₃CN-d₃ for elution of the unknown. NMR analysis (¹H-NMR, ¹H, ¹H-COSY, ¹H, ¹H-NOESY, ¹H, ¹³C-HSQC) was performed within 24 h to minimize degradation of the unknown compound. NMR spectra were acquired on a Bruker Avance 600 MHz instrument.

S6

Text S5: Proposed mechanism on the transformation of S_2O_8 .

Previous kinetics studies did not specifically examine the decomposition pathways of persulfate radical ($S_2O_8^{\bullet-}$). If $S_2O_8^{\bullet-}$ were to react with metals to generate peroxymonosulfate HSO₅⁻:

$$= M^{n+} + S_2 O_8^{\bullet-} + 2OH^- \to = M^{(n-1)+} + 2HSO_5^-$$
(S9)

 HSO_5 could reduce another metal, producing SO_4 and O_2 :^{9,2-4}

$$\equiv M^{n+} + HSO_5^- \rightarrow \equiv M^{(n-1)+} + SO_5^{\bullet-} + H^+$$
(S10)

$$2SO_5^{\bullet-} \rightarrow 2SO_4^{\bullet-} + O_2 \tag{S11}$$

A combination of Reactions S9-S11 becomes:

$$3 = M^{n+} + S_2 O_8^{\bullet-} + 2OH^- \rightarrow 3 = M^{(n-1)+} + 2SO_4^{\bullet-} + O_2 + 2H^+$$
(S12)

Considering Reactions S1, S2 and S12, every 2 moles of $S_2O_8^{2-}$ generate 3 moles of SO_4^{+-} , resulting in a maximum sulfate radical yeld of 150%. Alternatively, $S_2O_8^{+-}$ could undergo reactions that do not produce additional sulfate radical, in which case the maximum sulfate radical yield for the overall process would be 50%.

Therefore, the decomposition of S_2O_8 [•] through this mechanism could proceed through the pathway indicated below.



Text S6: Kinetics of SO4⁺ and persulfate radical calculation

Fate of SO_4^{\bullet} *in the absence of benzene:*

SO⁴ is generated from Fenton-like reactions:

$$\equiv M^{n+} + S_2 O_8^{2-} \implies \equiv M^{(n-1)+} + S_2 O_8^{*-}$$
(S1)

$$\equiv M^{(n-1)+} + S_2 O_8^{2-} \rightarrow \equiv M^{n+} + SO_4^{*-} + SO_4^{2-}$$
(S2)

In the absence of benzene, there are three reactions that act as sinks for SO₄::^{5,6,7}

$$SO_4^{\bullet-} + S_2O_8^{2-} \rightarrow S_2O_8^{\bullet-} + SO_4^{2-}$$
 $k_3 = 6.3 \times 10^5 \text{ M}^{-1}\text{s}^{-1}$ (S3)

$$SO_4^{\bullet-} + H_2 O \to SO_4^{2-} + HO^{\bullet} + H^+ \qquad k_4 = 660 \text{ s}^{-1}$$
 (S4)

$$SO_4^{-} + OH^{-} \rightarrow SO_4^{2-} + HO^{-}$$
 $k_5 = 7 \times 10^7 \text{ M}^{-1} \text{s}^{-1}$ (S5)

When $S_2O_8^{2^-}$ is activated (Reaction S1 to S2), one mole of $S_2O_8^{-}$ is generated for every mole of SO₄⁻ generated. In addition, $S_2O_8^{2^-}$ is also lost through reactions with SO₄⁻ (Reaction S3). The branching ratio for SO₄⁻ in the absence of benzene is $k_3[S_2O_8^{2^-}]/(k_4 + k_5[OH])$. At pH 8 and 1 mM S₂O₈²⁻, approximately half of the SO₄⁻ reacts with S₂O₈²⁻.

<u>Relative concentration of SO_4^{\bullet} :</u>

In the absence of benzene, the major reaction that acts as a sink for HO[•] is:^{8,9}

$$HO^{\bullet} + S_2 O_8^{2-} \rightarrow S_2 O_8^{\bullet-} + OH^- \qquad k_8 = 1.4 \times 10^7 \text{ M}^3 \text{ s}^3$$

$$\frac{d[HO^{\bullet}]}{dt} = k_4 [SO_4^{\bullet-}] + k_5 [OH^-] [SO_4^{\bullet-}] - k_6 [S_2 O_8^{2-}] [HO^{\bullet}]$$
(S6)

Under steady-state conditions: $[HO^{\bullet}]_{ss} = \frac{k_4 + k_5[OH^-]}{k_6[S_2O_8^{2-}]}[SO_4^{\bullet-}]_{ss}$

With 1 mM $S_2O_8^{2-}$ and at pH 8.0, $[HO^{\bullet}]_{ss}=0.05[SO_{4-}]_{ss}$. Therefore, the state-steady concentration of $SO_4^{\bullet-}$ is approximately 20 times higher than HO[•] under experimental conditions employed in this study.

Fate of SO_4^{\bullet} *in the presence of benzene:*

In the presence of 1 mM benzene, the following reaction becomes the major sink for SO₄:



The branching ratio for SO₄⁻ in the presence of benzene is k_2 [benzene]/(k_3 [S₂O₈⁻]+ k_4 + k_5 [OH⁻]). At pH 8, 1 mM S₂O₈⁻ and 1 mM benzene, essentially all SO₄⁻ reacts with benzene.

In the presence of benzene, there is an additional sink for HO:¹⁰



At steady-state conditions: $[HO^{\bullet}]_{ss} = \frac{k_4 + k_5[OH^-]}{k_6[S_2O_8^{2-}] + k_8[benzene]}[SO_4^{\bullet-}]_{ss}$

With 1 mM benzene, 1 mM $S_2O_8^{2-}$ and at pH 8.0, [HO[•]]_{ss}=0.01[SO₄⁻]_{ss}. Therefore, the state-steady concentration of SO₄^{•-} is still 2 orders of magnitude higher than HO[•] under experimental conditions in the presence of benzene in this study.

Text S7: Kinetics of organic radicals in the benzene/SO4⁺ system

The hydroxylcyclohexadienyl (HCHD, *i.e.*, C_6H_7O') radical can react with O_2 at the ortho position:¹⁰



In addition, HCHD radical can react with O₂ at the *para*-position as follows:



Based on the kinetics rate constants in Reactions 1 and 2,¹⁰ the steady-state concentration of organic peroxy radicals are:

$$\frac{d[o - C_6 H_7 O_3^{\bullet}]}{dt} = k_{12,f} [O_2] [HCHD] - k_{12,r} [o - C_6 H_7 O_3^{\bullet}] - k_{14} [o - C_6 H_7 O_3^{\bullet}] = 0$$

$$\frac{d[p - C_6 H_7 O_3^{\bullet}]}{dt} = k_{13,f} [O_2] [HCHD] - k_{13,r} [p - C_6 H_7 O_3^{\bullet}] - k_{15} [p - C_6 H_7 O_3^{\bullet}] - k_{16} [p - C_6 H_7 O_3^{\bullet}] = 0$$

Under air-saturated conditions (*i.e.*, $[O_2]=250 \mu M$), the steady-state radical concentrations are:

$$[o - C_6 H_7 O_3^{\bullet}]_{ss} = \frac{k_1 [O_2]}{k_{-1} + k_{1,f}} [HCHD]_{ss} = 0.045 [HCHD]_{ss}$$
$$[p - C_6 H_7 O_3^{\bullet}]_{ss} = \frac{k_2 [O_2]}{k_{-2} + k_{2,f} + k_{3,f}} [HCHD]_{ss} = 3.51 [HCHD]_{ss}$$

Therefore: $[p - C_6 H_7 O_3^{\bullet}]_{ss} = 80[o - C_6 H_7 O_3^{\bullet}]_{ss}$, and $p - C_6 H_7 O_3^{\bullet}$ accounts for 99% of $C_6 H_7 O_3^{\bullet}$.

The formation of phenol originates from both $o-C_6H_7O_3$ and $p-C_6H_7O_3$.

Because the branching ratio of $\frac{k_{1,f}[o - C_6H_7O_3^{\bullet}]_{ss}}{k_{2,f}[p - C_6H_7O_3^{\bullet}]_{ss}} = \frac{2 \times 10^4 \times 0.045[HCHD]_{ss}}{1.2 \times 10^{-1} \times 3.51[HCHD]_{ss}} = 2.1 \times 10^3$

Phenol is predominantly formed from $o-C_6H_7O_3^{\bullet}$.

Text S8: Estimate of phenol loss due to direct reaction of SO₄^{••} or HO[•] with phenol in the presence of benzene.

Phenol is produced from benzene reacting with SO.::



Phenol produced in this reaction can also react with SO₄:



At a given reaction time *t*, there is competitive kinetics between benzene and phenol:

$$\frac{d[benzene]_{t}}{dt} = -k_{7}[benzene]_{t}[SO_{4}^{\bullet-}]_{ss}$$
$$\frac{d[phenol]_{t}}{dt} = -k_{13}[phenol]_{t}[SO_{4}^{\bullet-}]_{ss}$$

Transform the above differential equation to numerical integration between reaction time t_i and t_i :

$$\Delta[benzene]_{t_2-t_1} = -k_7 \frac{([benzene]_{t_1} + [benzene]_{t_2})}{2} [SO_4^{\bullet-}]_{ss,t_1 \to t_2} (t_2 - t_1)$$

$$\Delta[phenol]_{t_2-t_1} = -k_{13} \frac{([phenol]_{t_1} + [phenol]_{t_2})}{2} [SO_4^{\bullet-}]_{ss,t_1 \to t_2}(t_2 - t_1)$$

Therefore,
$$\Delta[phenol]_{t_2-t_1} = \frac{k_{13}}{k_7} \bullet \frac{[phenol]_{t_1} + [phenol]_{t_2}}{[benzene]_{t_1} + [benzene]_{t_2}} [SO_4^{\bullet-}]_{ss,t_1 \to t_2}(t_2 - t_1)$$

 $\Delta[phenol]_{t_2-t_1}$ is the loss of phenol that should be accounted for as the oxidation product of benzene. Therefore, the total amount of phenol produced between the reaction time t_1 and t_2 is:

$$\Delta[phenol]_{produced,t_2-t_1} = \Delta[phenol]_{measured,t_2-t_1} + \Delta[phenol]_{t_2-t_1}$$

Based on the calculation, Δ [phenol]₁₂₄₁ is a very small fraction, *i.e.*, Δ [phenol]₁₂₄₁ is <0.1% of Δ [phenol]_{measured,1241}. Therefore, the loss of phenol due to reaction with SO₄⁻ is negligible.

In addition, since $[HO^{-}]_{a}$ is 2 orders of magnitude smaller than $[SO_{a}^{+}]_{a}$, and the rate constant of phenol reacting with HO $(k=6.6\times10^{9} \text{ M}^{-}\text{s}^{-})$ is similar to that of phenol reacting with SO₄⁺ $(k=8.8\times10^{9} \text{ M}^{+}\text{s}^{-})$, the loss of phenol due to reaction with HO is negligible as well.

Material Type	Material Name	BET Surface area (m ² /g)	Total Fe (wt %)	Total Mn (wt %)	Sand (wt %)	Silt (wt %)	Clay (wt%)
Aquifer Material	CADOU	3.9	0.77%	0.01%	84%	16%	4%
	CAROL	39.8	2.49%	0.02%	63%	18%	19%
	AWBPH	14.3	1.67%	0.03%	82%	10%	8%
	AFTCS	27.7	1.44%	0.03%	60%	22%	18%
	AMTAL	16.2	1.85%	0.12%	64%	22%	14%
Clay	Nontronite	69.0	26.2%	0.01%			100%
Material	Montmorillonite	32.0	2.6%	0.05%			100%
Pure Mineral	Geothite α -FeOOH _(s)	37.1	62.9%				
	Ferrihydrite Fe(OH) _{3(s)}	180.1	52.3%				
	Pyrolusite β -MnO _{2(s)}	0.11		62.8%			
	Silica SiO _{2(s)}	27.0					

Table S1 Characterization of aquifer materials, clay materials and pure minerals.¹

Chemical parameter	Concentration
Na ⁺	23 mg/L
Ca ²⁺	20 mg/L
Mg^{2+}	5 mg/L
SO ₄ ²⁻	20 mg/L
Cl	35.5 mg/L
Br	0.1 mg/L
NO ₃ -	1 mg/L
HCO ₃ -	1 mM
TDS	166 mg/L
Suwannee River NOM	1 mg C/L
pH	8

 Table S2 Chemical composition of synthetic groundwater used in this study.

The solution was buffered at pH 8.0 with 50 mM borate. The use of high buffer concentration was necessary to maintain a constant pH throughout the experiment.

• Experimental condition is with 50 g/L of minerals, initial persulfate concentration 1 mM and pH 8. In experiments with benzene, the initial concentration of benzene was 1 mM.

Table S3 Comparison of the production of aldehyde-like compound relative to phenol by $S_2O_8^{2-}$ and H_2O_2 .

	Relative ratio of aldehyde-like compound to phenol (10 ⁴ AU/ μ M) *		
	S ₂ O ₈ ²⁻ activation	H ₂ O ₂ activation	
Goethite	0.33 ± 0.11	0.23 ± 0.01	
Ferrihydrite	6.2 ± 2.4	Aldehyde not detected	
Pyrolousite	0.26 ± 0.09	Aldehyde not detected	

* The concentration of aldehyde-like compound is expressed as $\times 10^4$ adsorption units (AU) at the wavelength of 360 nm. The concentration of phenol is express as μ M.

Table S4 Comparison of oxidant yield in synthetic groundwater and MQ water from persulfate

 activation by different minerals*.

	Yield of SO ₄ ⁻ from S ₂ O ₈ ²⁻	Yield of SO4 ⁻ from S ₂ O ₈ ²⁻
	Synthetic Groundwater	MQ water
Goethite	155% ± 6%	167% ± 10%
Ferrihydrite	$20\% \pm 10\%$	$26\% \pm 8\%$
Pyrolousite	37% ± 5%	62% ± 12%

* Experimental condition: initial benzene=1 mM; initial persulfate=1 mM; mineral solids=50 g/L; pH=8.0; ionic strength=50 mM.



Figure S1 Measurement of total benzene (*i.e.*, aqueous benzene plus adsorbed benzene) in a 50 g/L pyrolusite suspension. Adsorbed benzene was recovered by acetonitrile extraction. Initial added benzene was 1000 μ M, pyrolusite concentration was 50 g/L, pH was buffered at 8.0 with 50 mM borate.



Figure S2 Changes in (A) persulfate concentration and (B) total benzene concentration via persulfate activation in the presence of ferrihydrite ($Fe(OH)_{3(s)}$). Initial benzene=1 mM, initial persulfate=1 mM, ferrihydrite concentration=50 g/L, pH=8.0.



Figure S3 Changes of (A) persulfate concentration and (B) total benzene concentration via persulfate activation in the presence of pyrolusite (β -MnO_{2(s)}). Initial benzene=1 mM, initial persulfate=1 mM, pyrolusite concentration=50 g/L, pH=8.0.



Figure S4 Changes of (A) persulfate concentration and (B) total benzene concentration via persulfate activation in the presence of montmorillonite. Initial benzene=1 mM, initial persulfate= 1 mM, montmorillonite concentration=50 g/L, pH=8.0.



Figure S5 Changes of (A) persulfate concentration and (B) total benzene concentration via persulfate activation in the presence of nontronite. Initial benzene=1 mM, initial persulfate 1 mM, nontronite concentration 50 g/L, pH 8.0.



Figure S6 Changes of (A) persulfate concentration and (B) total benzene concentration via persulfate activation in the presence of aquifer material AFTCS. Initial benzene=1 mM, initial persulfate=1 mM, AFTCS concentration=50 g/L, pH=8.0.



Figure S7 Changes of (A) persulfate concentration and (B) total benzene concentration via persulfate activation in the presence of aquifer material CADOU. Initial benzene=1 mM, initial persulfate=1 mM, CADOU concentration=50 g/L, pH=8.0.



Figure S8 Changes of (A) persulfate concentration and (B) total benzene concentration via persulfate activation in the presence of aquifer material AWBPH. Initial benzene=1 mM, initial persulfate=1 mM, AWBPH concentration=50 g/L, pH=8.0.



Figure S9 Changes of (A) persulfate concentration and (B) total benzene concentration via persulfate activation in the presence of aquifer material CAROL. Initial benzene=1 mM, initial persulfate=1 mM, CAROL concentration=50 g/L, pH=8.0.



Figure S10 Changes of (A) persulfate concentration and (B) total benzene concentration via persulfate activation in the presence of aquifer material AMTAL. Initial benzene=1 mM, initial persulfate=1 mM, AMTAL concentration=50 g/L, pH=8.0.





Figure S11 Identification of aldehyde-like compound as the oxidation product of benzene by SO_4 radicals. (A) HPLC-UV chromatogram of one sample obtained after persulfate was activated in the presence of benzene and minerals; (B) UV Spectra of the aldehyde-like product; (C) QTOF-LC-MS scan in negative mode of the aldehyde-like product. An exact mass of m/z 125.0246 was obtained, which corresponds to the sum formula $C_6H_5O_3$ (Δ ppm: 5.6).Fragmentation of the m/z 125 revealed cleavage of CO (-28 Da; fragments 97.0296 and 69.0350). At higher collision energies fragment 79.0180 increased indicating the cleavage of H₂O (-18 Da) from fragment 97.0296.



Figure S12 NMR spectra of the unknown compound confirm a ring cleavage product with a fragment identified as an enal. ¹H NMR (top left), ¹H, ¹H-NOESY (top right), ¹H, ¹H-COSY (bottom left), ¹H, ¹³C-HSQC (bottom right). ¹H NMR (600 MHz, CD₃CN) δ 9.32 (d, 1 H, *J* = 10), 8.63 (s, 2 H), 7.44 (d, 1 H, *J* = 15), 6.93 (dd, 1 H, *J* = 10, 15). ¹H-¹³C HSQC (600 MHz, 151 MHz, CD₃CN) δ (9.32, 198.5), (8.63, 185), (7.44, 143), (6.93, 115). MestReNova and iNMR (Mestrelab Research SL) were used to process the NMR spectra. Chemical shifts are reported in ppm and calibrated to residual solvent peaks. Coupling constants are reported in hertz. Unidentified peaks in the HSQC are due to contamination with phenol.



Figure S13 Formation of a bisulfite adduct of the unknown ring-cleavage product over time. The increase of the peak area for the bisulfite-adduct product indicated the presence of an aldehyde moiety. 10 mM bisulfite was mixed with SPE-enriched ring-cleavage product.



Figure S14 Formation of aldehyde-like product during persulfate activation by minerals. (A) Ferrihydrite; (B) Pyrolusite. Mineral mass loading=50 g/L, initial total benzene=1000 μ M, initial S₂O₈²⁻=1000 μ M, borate buffer=50 mM, pH=8.0.

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