Supplementary Material to:

An LCA of the Pelamis Wave Energy Converter

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S1 Input Data

S1.1 Key parameters

The analysis presented in this paper is an LCA of a single case-study manufacturing and installation scenario of the Pelamis Wave Energy Converter (WEC). In order to facilitate the use of this analysis in any "meta-models" of wave energy, it has been suggested by Astudillo et al. (2017) that a number of key parameters should be clearly reported. These are all detailed in the main text, but for clarity are summarised below.

S1.2 Data from manufacturer and detailed life cycle

The process of calculating the Life Cycle Inventory is described in [Figure S1.1.](#page-2-0) [Table S1.1,](#page-4-0) [Table S1.2,](#page-5-0) [Table S1.3](#page-6-0) an[d Table S1.4](#page-7-0) summarise the input data derived from information provided by Pelamis Wave Power Ltd (PWP), along with the selected process from ecoinvent v3.3 and the uncertainty indicator scores. The last refer to ratings used to estimate the uncertainty according to the same pedigree matrix used in the ecoinvent database, and described in Section 3.5 of the main report (Weidema, et al., 2013). [Figure S1.2](#page-3-0) describes the life cycle flows included/excluded from the study.

² Device design was superceded in 2010, but no data was gathered for an LCA of the later version of the machine before the manufacturer went into administration in 2014.

Figure S1.1 - Process of calculating life cycle inventory data from input data provided by Pelamis Wave Power Ltd.

Figure S1.2 - Flow chart describing the system evaluated. Processes framed in green are assumed to be included in the ecoinvent data. Processes framed in a red dashed line are excluded from the analysis

Table S1.1 - Input data for materials and manufacturing provided by PWP, with details of corresponding inventory processes and uncertainty indicator scores.

 3 Density 110 kg/m 3 from Trelleborg (2009).

⁴ With pipe extrusion, as detailed under Manufacturing Processes.

⁵ The PWP data for machining includes all small-scale precision removal of material, such as milling, grinding and drilling.

Table S1.2 - Information on electrical components provided by PWP, with details of corresponding inventory processes and uncertainty indicator scores.

 6 Generator for a gas cogeneration unit.

Table S1.3 - Input data for transport of components for power conversion modules provided by PWP or estimated, with details of corresponding inventory processes and uncertainty indicator scores.

⁸ Sum of all component estimates

 7 Assuming EURO3 standard, as this has the highest emissions and is therefore the most conservative

Table S1.4 - Input data for assembly and specialist sea vessel processes provided by PWP, with corresponding uncertainty indicator scores.

S1.3 Process approximations

The ecoinvent database does not contain detailed inventory information for some specialist materials, manufacturing processes and sea vessel operations. In order to assess the resource use and pollutant emissions associated with these, data on material quantities and fuel consumption were sourced elsewhere, and new processes were built using inventory data from ecoinvent. The selected materials, quantities and associated uncertainty are given in [Table S1.5,](#page-9-1) [Table S1.6,](#page-10-0) [Table S1.7,](#page-11-0) [Table S1.8](#page-12-0) and [Table S1.9.](#page-13-1)

S1.4 Waste disposal processes

[Table S1.10](#page-14-0) details the waste disposal processes selected from the ecoinvent dataset for each of the principal materials within the analysis.

S1.5 Alternative electricity generation

In order to compare the environmental impacts of the Pelamis with those from other types of power generation, selected average electricity generation data from a number of key energy sources was analysed with the ReCiPe Midpoint Hierarchist and Cumulative Energy Demand impact assessment methods. The processes selected from the ecoinvent database (v3.3) are detailed i[n Table S1.11.](#page-14-1)

Table S1.5 - Details of new processes created from data within ecoinvent for manufacturing processes and glass-flake paint, with corresponding uncertainty indicators

¹⁵ From Hempel (2007)

⁹ From Jiven et al. (2004).

¹⁰ Quantity derived from Axxiom (2008). Pressure from Kalpakjian et al. (2008). Compressed air sourced locally, so European average data selected.

¹¹ From data for abrasive blasting of aluminium in Classen et al. (2009).

¹² The paint is applied with an airless spray at 250 bar, providing a coverage of 3.9 m²/l with a thickness of 200μm (Hempel, 2007).

¹³ Parker et al. (2007) estimated an overall 1mm paint thickness requiring a base coat of primer, three layers of paint and a topcoat.

¹⁴ The paint application process was approximated from manufacturer's data for an airless spray pump (Graco, 2010), powered by 200 m³/min of compressed air to provide paint coverage of 12 l/min.

 16 Assumed to be the same as the glass flake paint, without the glass flakes.

Table S1.6 - Details of materials within glass-flake paint, selected ecoinvent data and corresponding uncertainty indicators

 ¹⁷ Uncertainty ranges taken from material data sheets

¹⁸ (Hempel, 2010b)

¹⁹ (Hempel, 2010a)

Table S1.7 - Details of materials in MV switch-disconnector cubicle, selected ecoinvent data and corresponding uncertainty indicators

 ²⁰ Second cubicle representing TSG control panel

²¹ (ABB, 2010)

Table S1.8 - Details of materials in SF6 breaker and transformer, selected ecoinvent data and corresponding uncertainty indicators

²³ (ABB, 2007)

Table S1.9 - Details of new processes created from data within ecoinvent for manufacturing processes and sea vessel operations, with corresponding uncertainty indicators

 ²⁴ (SWF, 2011)

²⁵ (Caterpillar, 2011)

²⁶ Scaled to match fuel consumption provided by PWP

Waste Material Selected Inventory Process

Table S1.10 - Waste processing datasets selected from ecoinvent.

²⁷ Only one nuclear power station in the UK is a pressurised water reactor. The remainder are advanced gas-cooled reactors, but as this is an old technology that is rarely used elsewhere, data for it is not included in ecoinvent v3.3.

S2 Additional Numerical Results

This section contains additional results not presented in the main article. [Table S2.1](#page-15-0) gives the breakdown of cumulative energy demand results for each primary energy carrier.

Table S2.1 - Breakdown of cumulative energy demand

The results of the uncertainty analysis are shown graphically in the paper, but for completeness, the numerical results are given i[n Table S2.2.](#page-16-0) Similarly, complete results of the sensitivity analysis are summarised in [Table S2.3.](#page-17-0)

Table S2.2 - Complete results of uncertainty analysis

Table S2.3 - Sensitivity analysis results. Highest values for each impact category are highlighted in orange, and lowest values in green.

Table S2.4 - Results of comparative uncertainty analysis of Pelamis with other types of generation. Values between 30 and 70% are highlighted, as these show a significant probability that the impacts of the Pelamis relative to the given type of generation may be reversed.

S3 Locational Adjustment Factors

The normalised impact potentials can be estimated for any given installation location, using the following equation:

 $E = \frac{(a + b \cdot l_{steel} + c \cdot l_{offshore})}{20W}$

where:

Note that this formula is a simplification of the results of this analysis, and cannot be used to determine the effect of a change in other factors. Furthermore, this model has been developed for an installation scenario in the UK, and therefore installation in other countries may not have the same impacts.

Table S3.1 - Constants for estimating the environmental impacts at alternative locations

S4 Recycling allocation

S4.1 Comparing recycled content method with APOS

Ecoinvent v3.3 includes data for two different allocation methods for attributional LCA: the recycled content method and the "allocation at the point of substitution" (APOS) method. In this study the former was chosen in order to enable consistency in application for foreground recycling processes. The latter is, however, considered by some to be the better approach for more consistent allocation (Schrijvers, et al., 2016). It is also the only available method in earlier versions of the ecoinvent dataset (v3 and v3.01) so will have been applied in other studies that also employ the recycled content method for the foreground data. The analysis was, therefore, re-run with the APOS approach applied to background processes, and results are given in [Table S4.1.](#page-21-0)

S4.2 Approximating the end-of-life recycling method

Section 5.2 of the main article describes how the analysis was re-run using an approximation of the end-of-life method for allocating recycling credit within the foreground data, in order to replicate the method applied by Parker et al. (2007). Although this method is no longer considered appropriate for use in an attributional LCA, it was tested here to explain the discrepancy in results between the two studies.

The end-of-life recycling method (also known as the avoided burdens or closed-loop approximation method) is a method of allocating credit for the avoided production of primary material in the future by producing recyclable material (Schrijvers, et al., 2016). Recycled material consumed in the product life cycle, therefore, does not give an environmental credit so has the same burdens has primary material. The underlying mathematical expression for this method from Schrijvers et al. can be rearranged to form Equation 1, assuming that the impacts of the substituted primary material will be the same as the impacts of the consumed primary material and the quality correction factor is one (as for closed-loop recycling of a material such as steel):

$$
E_{tot} = E_v + r(E_{RC} + E_{RRE} - E_v) + (1 - r)E_d
$$
\n(1)

where E_{tot} is embodied impacts per unit of material, E_v is embodied impact of primary material, E_{RC} is embodied impact of the recycling process, E_{RRE} is embodied impact of recovery and transport of the recyclable material, E_d is embodied impact of waste disposal and r is recycling rate at end-of-life. It can be seen that the first term E_v is the embodied impacts of all input material, which is considered to have the impacts of primary material. End-of-life impacts include the credit for recycling, described by $r(E_{RC} + E_{RRE} - E_v)$, which is a function of the difference between embodied impacts of the production of primary and recycled material. Disposal of non-recycled material is represented by $(1$ $r)E_d$.

In order to simulate the method applied by Parker et al., the above method was applied only to the foreground data for steel. All background data was still sourced from ecoinvent v3.3, using the recycled content allocation method, as with the main analysis. Modifications were made as follows:

• A new input steel dataset was created by copying the ecoinvent v3.3 data for the global steel market, but replacing all flows of recycled steel with data for primary steel for the same region. • Recycling credit was estimating by creating a waste flow with a global recycled steel market as input (as above, but with all primary steel replaced with recycled steel), and a global virgin steel market as avoided product.

Table S4.1 - Comparing results from the APOS and recycled content approaches to allocating for recycling

The result of running the analysis with this modification is a reduction in all impacts. Of the factors relevant for comparison with Parker et al.: climate change was found to be 28 g CO₂eq/kWh, cumulative energy demand 421 kJ eq/kWh and CO₂ emissions 26 g/kWh. This reduction is likely due to the recycling rate of 90% being much higher than the average recycled content of the global steel mix in the ecoinvent data (43%) (ecoinvent, 2016).

Errors may have been introduced to this analysis by using a mixture of allocation methods, so use of the method described here is not recommended.

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