

1 Increasing Isoprene Epoxydiol-to-Inorganic Sulfate Aerosol (IEPOX:Sulf_{inorg})
2 Ratio Results in Extensive Conversion of Inorganic Sulfate to Organosulfur Forms:
3 Implications for Aerosol Physicochemical Properties

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55 **Materials and Methods**

56 **Collection of PM_{2.5} Collection from SE-U.S.** PM_{2.5} samples were collected onto pre-baked 8 × 10 in
57 Tissuquartz™ Filters (Pall Life Sciences) using three high-volume PM_{2.5} samplers (Tisch Environmental)
58 operated at 1 m³ min⁻¹ during the 2013 Southern Oxidant and Aerosol Study (SOAS) campaign from 1 June
59 to 15 July 2013 at the CTR, AL ground site. One high-volume sampler collected PM_{2.5} for 23 h (08:00 to
60 07:00 the next day, local time), while the two remaining high-volume samplers followed two cycles. When
61 the sampling schedules were daytime (08:00 – 19:00, local time) and nighttime (20:00 – 07:00, local time),
62 the PM_{2.5} collected were defined as regular day or night samples, respectively. When atmospheric chemical
63 model simulations predicted a high level of BVOCs, sulfate and NO_x¹, PM_{2.5} were collected more frequently
64 (collection times were 08:00–11:00, 12:00–15:00, 16:00–19:00, and 20:00–07:00, local time) in order to
65 examine the potential enhancing effect of anthropogenic pollutants on BSOA formation, and are defined as
66 “intensive” samples. Field blanks were also collected weekly by placing pre-baked quartz filters into the 3
67 non-operating high-volume PM_{2.5} samplers for 15 min. Filters and field blanks were stored under dark
68 conditions at –20°C until analysis.

69 **Collection of PM_{2.5} Samples from Central Amazonia (Manaus downtown).** PM_{2.5} samples were collected
70 daily from 18 July through 1 August 2016 on filters similar to those used during the SOAS campaign using
71 a high-volume PM_{2.5} sampler operated at 1.13 m³ min⁻¹. The sampling site was located at 6 m above the
72 ground at the School of Technology of the State University of Amazonas in Manaus, Brazil. The high-
73 volume PM_{2.5} sampler was flow-calibrated and the filter holder was cleaned with the filter extraction solvent
74 each day before sampling to ensure no chemical carryover between samples. All filters were collected for
75 24 h. Immediately after collection, the 15 daily PM_{2.5} filter samples were stored at –18 °C under dark
76 conditions until chemical analysis. One field blank was collected during the 15-day campaign by placing a
77 pre-baked quartz filter into the high-volume PM_{2.5} sampler for 15 min, and then removing and storing it
78 under the same conditions as the field samples. Co-located measurements of total organic and inorganic
79 aerosol mass were not available for Manaus and the contribution of the OS could not be calculated.

80 ***Collection of PM₁ Samples from downwind Manaus.*** Filters were collected during the Green Ocean
81 Amazon (GoAmazon2014/5) field campaign² during intensive operating period 2 (IOP2) at the T3 site ~
82 70 km downwind of Manaus in August-October of 2014. The prevailing wind direction is from north-east,
83 transporting air masses over the Amazon forest to be mixed with the outflow from Manaus (to the east)
84 before reaching the sampling site. Transport time from city to T3 site was typically 4-5 h.³

85 Aerosol samples (PM₁) were collected on quartz fiber filters (101.6 mm diameter, QM-A Quartz,
86 Whatman) using a custom-built sequential filter sampler (Aerosol Dynamics, Inc.). Air was sampled 4 m
87 above ground level, passed through tubing kept at temperatures below the dew point for trapping excess
88 water, and then through a greaseless cyclone (with a 1 μm aerodynamic diameter cut-point), before
89 collection of particles. Filters were pre-treated by baking at 550°C for 12 hours. While samples were
90 collected with a 4-hour time resolution, 24-h average data are presented in this study. Field blanks were
91 collected each week by inserting pre-baked filters (same filters used for actual sample collection) into the
92 filter holders for at least a minute, and then removing following the same procedures as it is with field
93 samples. Particle and blank filters were kept frozen (-18°C) and transported on ice.

94 **Chemical Characterization**

95 ***Filter Extraction and Chemical Analysis of OS.*** Chemical characterization of the PM_{2.5} samples was
96 performed by UPLC/ESI-HR-Q-TOFMS (6520 Series, Agilent) in the negative ion mode, under operating
97 conditions described in detail elsewhere^{4,5}. High-volume quartz filter 37 mm punches were extracted with
98 22 mL of high-purity methanol (LC-MS CHROMASOLV-grade, Sigma-Aldrich) by sonication for 45 min,
99 filtered through 0.2-μm PTFE syringe filters (Pall Life Science, Acrodisc) to remove insoluble particles or
100 quartz filter fibers and blown dry under a gentle N₂ (g) stream at ambient temperature. The dried extracts
101 were reconstituted with 150 μL of a 50:50 (v/v) solvent mixture of methanol (LC-MS CHROMASOLV-
102 grade, Sigma-Aldrich) and high-purity water (Milli-Q, 18.2 M). Five μL aliquots were injected onto the
103 UPLC column (Waters ACQUITY UPLC HSS T3 column, 2.1 × 100 mm, 1.8 μm particle size) and eluted
104 at a flow rate of 0.3 mL min⁻¹ with a solvent mixture of methanol containing 0.1% acetic acid (LC-MS

105 CHROMASOLV-grade, Sigma-Aldrich) and water containing 0.1% acetic acid (LC-MS CHROMASOLV-
106 grade, Sigma-Aldrich).

107 Field, lab and blank filters from CTR, were spiked with known amounts of propyl and octyl sulfate
108 (electronic grade, City Chemical LLC) to estimate the extraction efficiency and potential losses throughout
109 the analytical protocol. Neither of these two OS have been identified in atmospheric PM_{2.5}, and thus can be
110 used as internal standards. Recovery was determined to be 89 ± 14 (1 std. dev.) %. To estimate potential
111 losses of MSA, control experiments were performed by analyzing five prebaked filters spiked with a known
112 amount of MSA. Extraction efficiency was above 70% and was used to correct the quantification of
113 samples. To optimize the quantification of identified BVOC-derived OS, a mixture of IEPOX-OS
114 (C₅H₁₁O₇S⁻)¹, 2-methylglyceric acid sulfate (C₄H₇O₇S⁻)⁶ and 2-oxopropyl sulfate (C₃H₅O₅S⁻)⁷ were
115 synthesized in-house as authentic standards. Propyl sulfate (C₃H₇O₄S⁻; electronic grade, City Chemical
116 LLC) and 3-pinanol-2-hydrogen sulfate (C₉H₁₃O₆S⁻; synthesized standard) served as surrogate standards to
117 quantify the remaining OS. Surrogate standards were selected to match as precisely as possible the retention
118 times of the compounds of interest, as summarized in Tables S2-S4. Due to the potential influence of marine
119 air masses at the CTR site, methanesulfonic acid formed from the oxidation of dimethyl sulfide ⁸ was also
120 quantified in all PM_{2.5} samples from CTR using an authentic standard (MSA, Sigma Aldrich, ≥ 99.5%) to
121 help with mass closure of total organosulfur compounds.

122 PILS vials were analyzed by UPLC/ESI-HR-Q-TOFMS operated in negative mode following the
123 analytical procedure described for the ambient filter extracts. PILS vials were analyzed for the data
124 presented in the main text (Figure 1) in order to provide time-resolved information for the conversion of
125 SO₄²⁻ to organosulfur forms.

126 ***PILS Operation and Dilution Correction.*** The aerosols are sampled through an impactor with a 2.5- μ m
127 cut-off at a flow rate of ~ 13 L min⁻¹ with a carbon strip denuder (Sunset Labs) upstream of the impactor to
128 remove organic vapor. Cool sample air flow was mixed adiabatically with a steam heated at 98.5-100°C in
129 the PILS condensation chamber, which allows aerosol particles to grow sufficiently for collection on a

130 quartz impactor plate. Impacted droplets were transferred by a wash-flow at 0.50 – 0.55 mL min⁻¹ into a
131 debubbler and the resulting bubble-free sample liquid was delivered through a tubing with an inline filter
132 into 2 mL glass vials in a rotating-carousel auto-collector (BMI). Air sampling rate and wash-flow rate were
133 checked and recorded before and after each experiment. Milli-Q water used in the wash-flow was spiked
134 with 25 μM of lithium bromide (LiBr) as an internal standard to correct for dilution caused by condensation
135 of water vapor during droplet collection inside the PILS. Ideally the concentrations of lithium bromide in
136 the samples should be identical to that in the wash-flow supply bottle without dilution. When taking into
137 account the addition of water due to condensations, the dilution factor can be calculated on a per sample
138 basis using bromide ion concentration measured by IC as such:

$$139 \quad \text{dilution factor} = \frac{[\text{Br}^-]_{\text{wash}}}{[\text{Br}^-]_i}$$

140 In this equation, subscript “i” denotes bromide concentration measured in each sample and subscript “wash”
141 denotes the bromide concentration measured for the liquid in the wash-flow supply bottle. Then the
142 correction was made for all SOA tracers in a sample by multiplying the measured tracer concentrations by
143 the dilution factor derived above.

144 ***Uncertainty Estimates for OS.*** The use of surrogate standards can lead to significant biases in the OS
145 quantification. The ESI can be highly impacted by the chemical composition of the mobile phase; for
146 example, when increasing the organic content of the eluent, the ionization efficiency increases significantly
147 ⁹. As presented Table S5, calibration factors were determined for several authentic OS standards to evaluate
148 the impact of the chemical structure as well as the mobile phase composition on the ionization efficiency
149 of the compounds of interest. During the first two minutes of the gradient elution, the mobile phase is
150 composed 100% water, and therefore effect of solvent composition on ionization efficiency is not expected
151 within this time period. We note that most of the isoprene-derived OS eluted within the first two minutes
152 and sensitivities towards standards eluting within this period are independent of structure. This suggests
153 that uncertainties in quantification using the surrogate standards for OS eluting in this time range are most

154 likely small (< 10%). For the later eluting compounds (e.g. monoterpene-derived OS) the composition of
155 the mobile phase changes and could impact the ionization efficiency. Propyl sulfate and 3-pinanol-2-
156 hydrogen sulfate were used as surrogate standards and selected to correspond as precisely as possible to the
157 retention times of the OS identified in the SOA collected during SOAS. Although the sensitivities of propyl
158 and 3-pinanol-2-hydrogen sulfates are of the same order, biases can be expected and might lead to an
159 inaccurate estimation of the concentration of the OS eluting later in the run.

160 We assume a small uncertainty in the calibration curve itself (5%). Uncertainties from the extraction
161 based on the recovery of propyl and octyl sulfates are 15%. In addition to extraction efficiencies and
162 calibration uncertainties, uncertainty related to the volume of air sampled could further contribute to overall
163 uncertainty. In the different campaigns, high-volume samplers were calibrated and the air volumes adjusted
164 based on the meteorological conditions. Therefore, this uncertainty was estimated as 5%. Overall
165 uncertainty in quantification is estimated to be $\pm 17\%$.

166 ***Characterization of Total Water-Soluble Organosulfur at CTR.*** Total water-soluble organosulfur
167 compound mass present in SOA collected in the SE-U.S. was determined from the difference between total
168 water-soluble sulfur measured by isotope ratio inductively coupled plasma mass spectrometry (IR-ICP-MS)
169 and sulfate-sulfur measured by ion chromatography (IC) in the same sample aliquot. Sample extractions,
170 sample spiking and standard preparation were performed gravimetrically, using a 5-place readout laboratory
171 balance (Mettler Toledo Model MS205DU). High-volume quartz filter punches (47 mm, or 37 mm
172 diameter) were extracted in Thermo-Fisher Nalgene wide mouth lab quality HDPE bottles (1 oz. 02-893-
173 5A) with 25 g distilled deionized water (DIW with 18.2 ± 0.2 M Ω cm resistivity). Each vessel was capped
174 by a polypropylene cap and sonicated without heat for 1 h. After sonication, the sample was allowed to
175 cool, shaken for uniformity, and allowed to equilibrate at 8°C overnight.

176 Each extract was analyzed in triplicate for sulfate via IC using a dual-channel Dionex ICS-3000 ion
177 chromatography system and Dionex AS-1 Autosampler. The IC was calibrated with NIST SRM 3181
178 (sulfate in water) at concentrations between 10 ngS g⁻¹ and 2,000 ngS g⁻¹. A gradient profile (5mM-30mM)

179 of KOH was used as eluent for sulfate-S separation (MSA for cations in other channel) with 4 micro-bore
180 IonPac analytical column AS18 (CS16 for cations). Each channel operated a self-regenerating SRS-ULTRA
181 suppressor in external DIW regeneration mode, a CD-1 conductivity detector, and a DP-1/SP-1 gradient
182 pump with built-in degassing. The applied micro-bore system allowed an analyte flow rate of 1 mL min⁻¹.
183 DIW was supplied directly to the eluent generator module. The instrument detection limit (IDL) for the
184 sulfate anion expressed in ambient concentrations from field blanks amounts to less than 2 ngS m⁻³. Average
185 precision determined from triplicate analysis of the samples collected during the study is 0.46 ± 0.40%.
186 Secondary source standards were used to routinely assess accuracy in terms of recovery, which averages
187 99.6 ± 0.4%.

188 Each extract was also analyzed in triplicate for total-S using a Perkin Elmer ELAN DRC-II IR-
189 ICP-MS operated in Dynamic Reaction Cell (DRC) mode with oxygen reaction gas (UHP 99.999% O₂,
190 Airgas, 4808 Nelson Rd. Morrisville, NC 27560). The IR-ICP-MS was calibrated with NIST SRM 3181 at
191 concentrations between 100 ngS g⁻¹ and 2000 ngS g⁻¹. Standards and sample aliquots were spiked with
192 Na₂³³SO₄ (Sigma-Aldrich cat. # 719374) to a final concentration of approximately 1000 ng³³S g⁻¹ and then
193 analyzed for ³²SO (48amu), ³³SO (49amu) and ³⁴SO (50amu). Total water-soluble S was calculated as
194 follows:

$$\text{Total-S [ngS g}^{-1}\text{]} = (\text{}^{32}\text{SO}/\text{}^{33}\text{SO})_{\text{sa}} * [\text{}^{33}\text{SO}]_{\text{sp}} / \text{IR}_{48/49}$$

196 with ³²SO and ³³SO = background-corrected counts at 48 amu and 49 amu, respectively; sa = sample; sp =
197 spike concentration [ngS g⁻¹]; IR_{48/49} = ion intensity ratio for masses 48 and 49. The instrument detection
198 limit of the IR-ICP-MS can be estimated conservatively at 2 ngS g⁻¹ or 50 ngS⁻¹ per sample, translating to
199 less than 2 ngS m⁻³ atmospheric PM_{2.5} concentration of total water-soluble S. Average precision for triplicate
200 analysis of samples was 0.35 ± 0.20%. Recoveries of a secondary sulfate standard (10,000 μgS mL⁻¹ ±
201 0.5%, 10M54-5 from HP Standards, Charleston, SC) were 100.3 ± 0.4%, and recoveries of camphorsulfonic
202 acid (CSA), methanesulfonic acid (MSA) and methionine were 99.9 ± 0.3%, 98.9 ± 0.3% and 98.2 ± 0.2%,
203 respectively (purity of CSA and MSA was 99% from ACROS Organics via Fisher-Scientific, purity of

204 methionine was >98.5% from Fisher BioReagents). All recovery checks were run at a 10% frequency or
205 every 10 samples.

206 Based on the above results (in particular the 0.16% uncertainty in the NIST SRM 3181, the linearity
207 of the IC and IR-ICP-MS calibrations, and the instruments' recoveries of the secondary standards), the
208 approximate 95% ($\sigma=2$) uncertainty for the difference between total-S and sulfate-S is $\pm 1.6\%$. In other
209 words, water-soluble organosulfur compounds are greater than 0 at the 95% confidence level when total-S
210 via IR-ICP-MS is at least 1.6% greater than sulfate-S via IC.

211 *Uncertainty Estimates for Water Soluble Sulfate-S by IC and Water Soluble Total-S by ICP-MS.*

212 Sulfate-S via Ion Chromatography

213 The mass of sulfate-S in a sample extract is determined as shown in eqn. 1.

$$214 \text{ Sulfate-S(ng)} = \text{Sulfate-S(ng-g}^{-1}\text{)} * \text{Extract Mass (g)} \quad (1)$$

215 The uncertainty (U) in Sulfate-S can then be estimated as shown in eqn. 2

$$216 U_{\text{Sulfate-S}} = ((\mu_{\text{Sulfate-S}})^2 + (\mu_{\text{Extract Mass}})^2)^{1/2} \quad (2)$$

217 Which can be expanded to

$$218 U_{\text{Sulfate-S}} (\%) = 100\% * ((\mu_{\text{Primary}})^2 + (\mu_{\text{Secondary}})^2 + (\mu_{\text{Rep}})^2 + (\mu_{\text{Dilution}})^2 + (\mu_{\text{Extract Mass}})^2)^{1/2} \quad (3)$$

219 Where,

220 μ_{Primary} = relative uncertainty in the primary sulfate standard (NIST SRM 3181) = 0.0016, from
221 NIST certificate.

222 $\mu_{\text{Secondary}}$ = relative recovery of the secondary sulfate standard = 0.0039, measured.

223 μ_{Rep} = relative standard deviation of sample replicates = 0.0046, measured.

224 μ_{Dilution} = relative standard deviation of primary standard diluent mass = 0.0003, based on balance
225 uncertainty of +/- 10 mg for 30 g standard.

226 $\mu_{\text{Extract Mass}}$ = relative standard deviation of extract mass = 0.0003, based on balance uncertainty of
227 +/- 10 mg for 30 g sample extract.

228 Estimated $U_{\text{Sulfate-S}} = 0.60\%$

229 Total water-soluble S (Total-S) via isotope ratio ICP-MS

230 The mass of Total-S in a sample extract is determined as shown in eqn. 4.

231
$$\text{Total-S(ng)} = \text{Total-S(ng-g}^{-1}\text{)} * \text{Extract Mass(g)} \quad (4)$$

232
$$= ({}^{32}\text{SO}/{}^{33}\text{SO})_{\text{sa}} * [{}^{33}\text{S}]_{\text{sp}} / \text{IR}_{48/49} * \text{Extract Mass} \quad (5)$$

233 Where,

234 $({}^{32}\text{SO}/{}^{33}\text{SO})_{\text{sa}}$ = ratio of background-corrected counts at 48 amu and 49 amu

235 $[{}^{33}\text{S}]_{\text{sp}}$ = concentration of ${}^{33}\text{S}$ in the spiked sample

236 $\text{IR}_{48/49}$ = ion ratio for 48 amu and 49 amu (counts per ng-g⁻¹/ counts per ng-g⁻¹)

237 $\text{IR}_{48/49}$ is determined by spiking approximately 1000 ngS-g⁻¹ of K₂³³SO₄ into solutions of NIST SRM 3181
238 containing 0, 100, 300, 500, 750, 1000 and 2000 ngS-g⁻¹ of total-S. Note that $\text{IR}_{48/49}$ decreases by about
239 1% over the observed calibration range, due to the presence of naturally-occurring ${}^{33}\text{S}$ in SRM 3181. To
240 partially compensate for decreasing IR, we bin IRs over successive pairs of calibration standards and use
241 the one that brackets an individual sample concentration to calculate Total-S. For example, if the initial
242 estimate of Total-S is between 100 and 300 ng-g⁻¹, we use the average of IRs obtained at calibration levels
243 100 and 300 ngS-g⁻¹.

244 The overall procedure assumes that the isotopic composition of SRM 3181 is at least roughly equal
245 to that of naturally-occurring sulfur (and, by inference, ambient samples). To confirm this, we asked a third-
246 party laboratory to measure $\delta^{34}\text{S}$ on a subsample of SRM 3181. Results showed $\delta^{34}\text{S}$ of +17 parts per
247 thousand, relative to Canyon Diablo Troilite. The $\delta^{33}\text{S}$ should be similar to this, based on mass dependent
248 fractionation (0.5 times $\delta^{34}\text{S}$) plus mass independent fractionation (0.6 to 1.0 times $\delta^{34}\text{S}$). The impact
249 of these enrichments is negligible, so we did not adjust SRM 3181 concentrations.

250 Based on the above, $U_{\text{Total-S}}$ can be estimated as shown in eqn. 6

251
$$U_{\text{Total-S}} = 100\% * ((\mu_{\text{CP-3181}})^2 + (\mu_{\text{IP-33S}})^2 + (\mu_{\text{CP33S}})^2 + (\mu_{\text{IR}})^2 + (\mu_{32\text{SO}/33\text{SO}})^2 + (\mu_{\text{Secondary}})^2 +$$

252
$$(\mu_{\text{Dilution}})^2 + (\mu_{\text{Spike Mass}})^2 + (\mu_{\text{Extract Mass}})^2)^{1/2} \quad (6)$$

253 Where,

254 $\mu_{\text{CP-3181}}$ = chemical purity of SRM 3181 = 0.004 (manufacturer's certificate).

255 μ_{IP-33S} = isotopic purity of ^{33}S standard = 0.0013 (ARA analysis of high concentration solutions).

256 μ_{CP33S} = chemical purity of ^{33}S standard, pooled relative standard deviation of for sulfate, CSA,
257 MSA and methionine = 0.006.

258 μ_{IR} = relative standard deviation of IR across full calibration curve = 0.0034.

259 $\mu_{32SO/33SO}$ = relative standard deviation of triplicate sample analyses = 0.0035.

260 $\mu_{Secondary}$ = relative standard deviation of secondary SO_4 standard recoveries = 0.0042.

261 $\mu_{Dilution}$ = relative standard deviation of primary standard diluent mass = 0.0003, based on balance
262 uncertainty of +/- 10 mg for 30 g standard.

263 $\mu_{Spike\ Mass}$ = relative standard deviation of primary standard diluent mass = 0.0016, based on balance
264 uncertainty of +/- 10 mg and 6 g sample aliquot.

265 $\mu_{Extract\ Mass}$ = relative standard deviation of extract mass = 0.0003, based on balance uncertainty of
266 +/- 10 mg for 30 g sample extract.

267 Estimated $U_{Total-S}$ = 0.99%.

268 Combining $U_{Sulfate-S}$ and $U_{Total-S}$ and applying a coverage factor (k) of 2 yields an estimated uncertainty for
269 the difference between Total-S and Sulfate-S of 2.3%, with a confidence interval $\geq 75\%$.

270 **Microscopy Imaging.** A 3-stage microanalysis particle sampler (MPS-3, California Measurements, Inc.)
271 with size cuts of 5.0, 2.5, and 0.4 μm was operated at 2.1 L min^{-1} . Particles were impacted onto carbon-
272 type-b Formvar-coated copper transmission electron microscopy (TEM) grids and silicon wafer substrates
273 (Ted Pella, Inc.). Samples were stored in sealed plastic vials at room temperature prior to analysis. Particles
274 collected onto silicon substrates were imaged in 5x5 μm regions by an AFM (Anasys Instruments) in
275 tapping mode with 75 \pm 15 kHz resonant frequency and 1-7 N m^{-1} spring constant at ambient laboratory
276 temperature (~ 23 $^{\circ}C$), pressure, and RH ($\sim 36\%$). SEM analysis of particles collected onto TEM grids was
277 performed by a FEI Helios 650 Nanolab-Dualbeam electron microscope equipped with a high angle annular
278 dark field (HAADF) detector operated at an accelerating voltage of 10.0 kV and a current of 0.80 nA under
279 vacuum conditions (10^{-3} to 10^{-5} Pa). SEM particles were analyzed orthogonal to the beam on TEM grids
280 and at a 55-degree angle (tilted) on the silicon wafer.

281 ***Biphasic Microfluidic Experiments with Samples Extracted from Filters.*** Microfluidic experiments were
282 also conducted with SOA samples (2-MT and IEPOX-OS) extracted from filters using methanol as the
283 solvent. It is observed (Figures S2A and S2B) that methanol itself causes a significant decrease in interfacial
284 tension compared to pure water. Furthermore, similar to the 2-MT and IEPOX-OS samples in water,
285 interfacial tension depression was observed with methanol as the solvent for SOA samples. Figure S2A
286 shows the slight decrease in interfacial tension measured with 2-MT in methanol, compared to the solvent
287 (methanol in water). At 50 mg/mL SOA, IEPOX-OS shows a larger depression in interfacial tension
288 compared to 2-MT (Figure S2B). Finally, when AS is added, significant salting out is observed at 25 mg/mL
289 of IEPOX-OS (Figure S2B).

290 ***Isotherm Model for Interfacial Tensions.*** A surface tension model using adsorption isotherms and based
291 on statistical mechanics was applied to the microfluidic measurements for the 2-MT (Figure S9) and
292 IEPOX-OS compounds (Figure 4A). Traditionally, Langmuir-like isotherm models are applied to surface
293 active organics, such as two parameter Szyszkowski equation (see eq. 1 in Schweir et al.¹⁰ for an example).
294 In the model applied here, unique features of the surface partition function are that the adsorption sites are
295 solvent molecules (waters) and that solutes can occupy a non-integer number of sites using a multifactorial
296 skip, implicitly accounting for the size of the solute. The model expression for binary solutions is $\sigma = \sigma_w -$
297 $\frac{KT}{rS_w \ln(1 + K a_s)}$, where r is the surface partition function multifactorial skip, K is a solute bulk energy
298 term, σ_w is the surface tension of pure water, S_w is the area of adsorption sites (the projected area of one
299 water molecule, 0.1 nm^2). The expression was normalized by the interfacial tension value of pure water from
300 biphasic microfluidic tensiometry and subsequently applied to the measurements of 2-MT and IEPOX OS
301 water. The solute bulk energy term (K) follows the expected size-dependence for each compound: $K =$
302 47.35 for 2-MT ($M_w = 136 \text{ g mol}^{-1}$) and $K = 66.00$ for IEPOX-OS ($M_w = 215 \text{ g mol}^{-1}$). The surface function
303 variable (r) also follows the size-dependence, as $r = 71.38$ for IEPOX-OS and $r = 27.24$ for 2-MT. For the
304 salty solutions, the concentrations, or solute activities, were modified using Setschenow constant ($K_s c_{\text{salt}} =$
305 $\log(\gamma/\gamma_0)$), where γ is the organic activity in salty water and γ_0 is the organic activity in pure water. Thus,

306 K_s shows enhancement or depletion of organic surface partitioning due to the presence of salt in solution.
307 An important assumption is that in pure water, the organic activity coefficient is unity at concentrations less
308 than 1.0 mol/L. Incorporating the Setschenow equation into the surface tension equation results in $\sigma = \sigma_w -$
309 $kT/rS_w \ln(1 + K_{a_s} * 10^{K_s c_{salt}})$. If parameters K and r are optimized from the pure water case and interfacial
310 tension is known for the same concentration of organic in both pure water and salty water, then K_s is the
311 single parameter. Recently, Toivola et al.¹¹ performed quantum chemistry computations of K_s for 2-MT
312 in 0.009 and 0.09 mole fraction of AS and found $K_s = 0.568$ kg-mol⁻¹ and $K_s = 0.866$ kg-mol⁻¹,
313 respectively. Our AS mole fraction is 0.03 and we find $K_s = 0.299$ kg-mol⁻¹ for 2-MT.

314 **Flow Tube Experiments.** The flow tube, CIMS, as well as aerosol and gas phase IEPOX generation have
315 been described in detail elsewhere.¹² Briefly, the flow tube consists of a 6 x 90 cm Pyrex cylinder having
316 inner walls coated with halocarbon wax to reduce the wall loss of IEPOX gas. ABS aerosol particles were
317 generated using a constant output atomizer (TSI Inc., Model 3076) from dilute solutions (0.1 wt %). The
318 atomizer output was diluted and conditioned to about 38% RH by mixing with a 3 L min⁻¹ flow of
319 humidified ultrahigh purity (UHP) N₂ before entering the flow tube. The IEPOX vapor was generated by a
320 flow of 30 standard cubic centimeters per minute (sccm) of UHP N₂ over ~200 μ l *trans*- β -IEPOX solution
321 in ethyl acetate in a glass bulb at room temperature (~23 °C), and was injected into the flow tube via a
322 movable injector downstream of the aerosol inlet. A constant 2 L min⁻¹ flow of conditioned aerosol was
323 drawn through the flow tube by the CIMS and a scanning mobility particle sizer (SMPS) connected to the
324 flow tube exit, which provided real-time measurements of the gas-phase IEPOX concentration and aerosol
325 surface area (Sa), respectively. The IEPOX injector was moved in 10 cm increments from the bottom (10
326 cm above the flow exit) to the top (70 cm above the flow exit) of the flow tube to vary the reaction time
327 between IEPOX and aerosol particles. The decay of gas phase IEPOX signal versus injector position
328 (reaction time) in the presence and absence of aerosol particles was measured to derive k_{obs} and k_{wall} ,
329 respectively, and the pseudo-first-order reaction rate constant k_{het} for IEPOX uptake onto particles was
330 determined as $k_{het} = k_{obs} - k_{wall}$. The IEPOX signals measured at every three adjacent injector positions (i.e.,

331 10-30, 20-40, 30-50, 40-60, 50-70 cm), which correspond to different average reaction times, were used to
332 determine the k_{het} and then γ_{IEPOX} as a function of reaction time. γ_{IEPOX} is calculated by the following
333 equation:

$$334 \quad \gamma_{IEPOX} = \frac{4k_{het}}{S_a\omega} \quad (7)$$

335 where ω is the mean molecular speed of *trans*- β -IEPOX under the experimental condition ($\omega_{IEPOX}=231$ m
336 s^{-1}).^{13,14}

337 **Viscosity Calculation of IEPOX-derived OS and water mixtures.** The improved parameterization by
338 Derieux et al.¹⁵ was designed to predict, Tg of pure compounds containing carbon, hydrogen and oxygen
339 using the number of carbon (nC), hydrogen (nH), and oxygen (nO), as shown in eqn. 8.

$$340 \quad T_{g,org} = \frac{n_C^0 + \ln(n_C)}{b_C} + \ln(n_H)b_H + \ln(n_C)\ln(n_H)b_{CH} + \ln(n_O)b_O + \ln(n_C)\ln(n_O)b_{CO} \quad (8)$$

341
342 where n_C^0 is the reference carbon number, b_C , b_H and b_O represent the contribution of each atom to T_g , and
343 b_{CH} and b_{CO} are coefficients that reflect contributions from carbon–hydrogen and carbon–oxygen bonds,
344 respectively. The values for these parameters used in this study were derived by DeRieux et al.¹⁵ based on
345 empirical fit and are shown in Table S6. However, the sulfur atom was not considered in the method. For a
346 conservative estimation of T_g of IEPOX-OS, the sulfur atom is assumed to have the same effect on T_g as an
347 oxygen atom. The T_g of IEPOX-OS, IEPOX-OS dimer and trimer are estimated to be 298 K, 333 K, and
348 360 K, respectively. The glass transition temperature of the sum of uncharacterized IEPOX-derived OS is
349 assumed to be the same as the IEPOX-OS dimers to give a more conservative estimation. Taking into
350 account the mass fraction of each OS compound based on the chemical data shown in Figure 1B (Amazon
351 ratio) at 120 min, the T_g of the IEPOX-OS mixtures is calculated using the Gordon-Taylor equation based
352 on the estimation of hygroscopicity, and the Gordon-Taylor constant among organic compounds ($k_{GT}=1$)
353 as well as the Gordon-Taylor constant between water and the organics ($k_{GT}=12.5$), as shown in eqn. 9. The
354 upper hygroscopicity value of 0.15 is based on the upper hygroscopicity range of organic compounds.

355 Gordon-Taylor constants between IEPOX-OS mixtures and water, and IEPOX-OS mixtures themselves are
 356 assumed to be 2.5, and 1, respectively based on previous studies.¹⁵⁻¹⁸

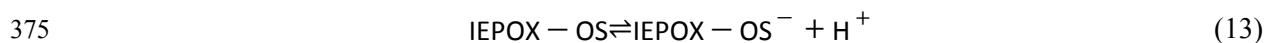
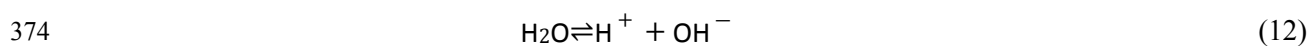
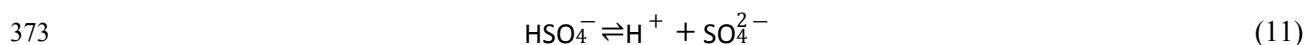
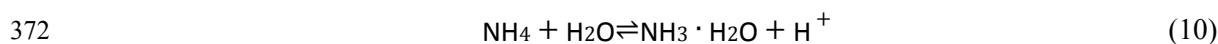
$$357 \quad T_{g,mix}(RH) = \frac{(1 - w_{org}(RH))T_{g,c} + \frac{1}{k_{GT}} \sum_{i=1}^n w_{i,org}(RH) T_{g,i,org}}{(1 - w_{org}(RH)) + \frac{1}{k_{GT}} \sum_{i=1}^n w_{i,org}(RH)} \quad (9)$$

358 where n denotes total number of organic compounds in the mixture and i denotes each organic compound;
 359 $w_{(i,org)}$ is the mass fraction of organic compound i. The dependency of viscosity on RH is reflected in
 360 mass fraction terms as variable RH changes the amount of water taken up by the SOA, which can be
 361 calculated using hygroscopicity (κ).^{15,19}

362 ***Thermodynamic Model of Aerosol pH.***

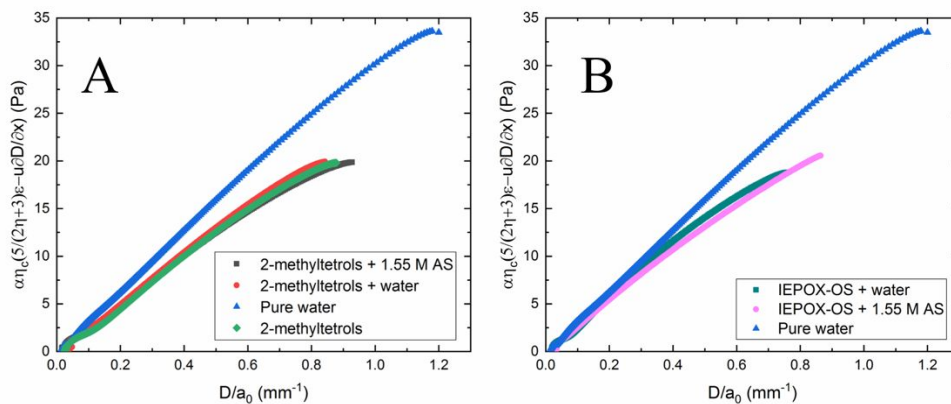
363 The pH of aerosol particles during the reaction processes are calculated using a thermodynamic model
 364 employing two methods. The first method uses measured SO_4^{2-} concentration as well as the NH_4^+
 365 concentration at the beginning of the experiment as the input values. The IEPOX-OS concentration was not
 366 considered. The second method takes into consideration all cations and anions in the aqueous aerosol
 367 particle.

368 The acidity model considers the acid disassociation of NH_4HSO_4 and H_2O . For scenario 1, the
 369 IEPOX-OSs were not considered. For scenario 2, IEPOX-OSs, including IEPOX-OS, IEPOX-OS dimer,
 370 and IEPOX-OS trimer, were taken into consideration in the model. The acid disassociation reactions of the
 371 above compounds are described below:

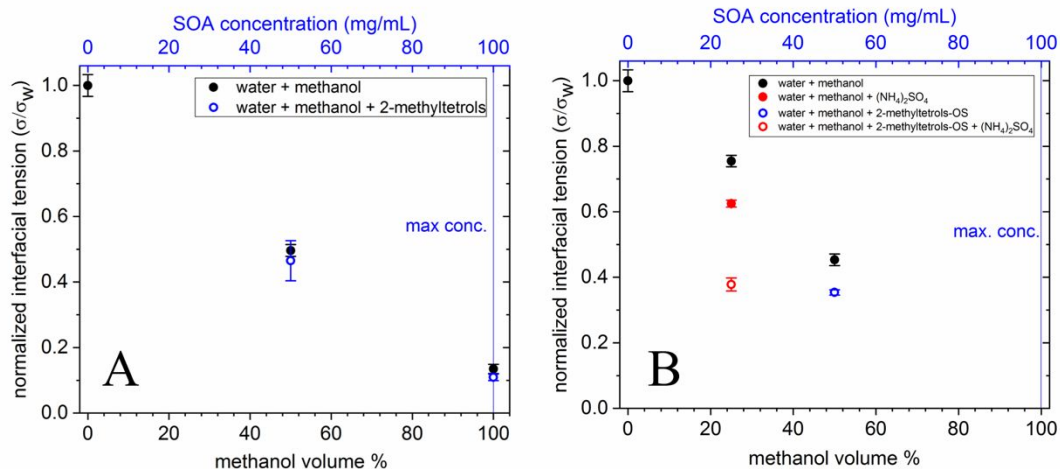


378 The acid disassociation equilibrium equations for (10)-(15) were constructed, with pK_a values for Eqs.
379 (7,8,9) were obtained from the literature. The pK_a values for Eqs. (13)-(15) were assumed to be ~ 2 , with a
380 range of 0-4. The proton balance equation together with the mass balance equations for NH_4^+ , HSO_4^- , H^+ ,
381 SO_4^{2-} , and IEPOX-OS monomer, dimer, and trimers were built to solve the H^+ ion concentration. The gas-
382 particle balance of the NH_3 , NH_4^+ was considered with a Henry's law constant of $0.0161 \text{ atm M}^{-1}$. A growth
383 factor of 1.3 was used to calculate the liquid water content of the particles. The activity coefficients for all
384 ions were assumed to be 1.

385



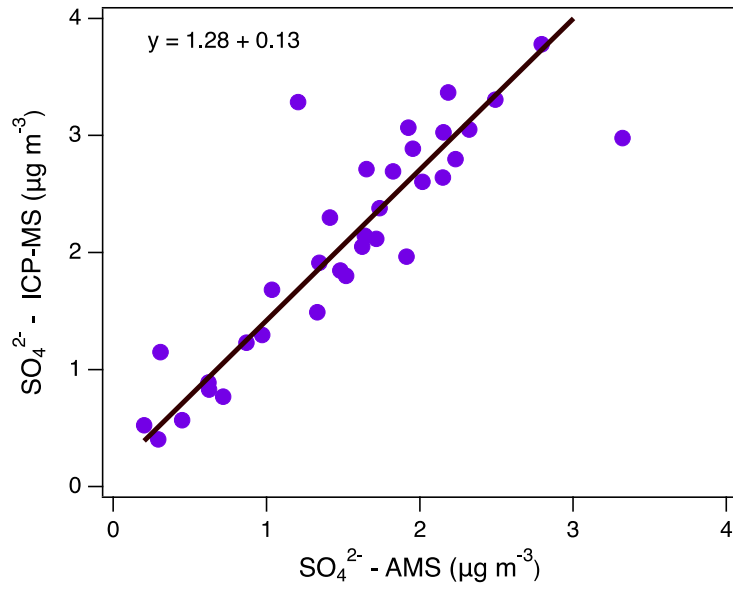
386
 387 **Figure S1.** Representative Taylor plots (single droplet) for (A) 2-MT and (B) IEPOX-OS samples (in
 388 water). The slope of the linear portion of the curves is equal to the interfacial tension. Interfacial tensions
 389 reported in this work are the median values of interfacial tension (over all droplets) for any given system.
 390 Normalized data points are shown in Figures 4A and S3.
 391



392

393 **Figure S2.** Interfacial tensions measured with microfluidics, for (A) 2-MT and (B) IEPOX-OS, with
 394 methanol as the solvent. For 2-MT, approximately 300 μ L of sample was used to prepare a 50 mg/mL
 395 sample of 2-MT in water, approximately 300 μ L was used to prepare a 50 mg/mL solution with AS and the
 396 remaining was used for the 100 mg/mL experiment. Similarly, two different solutions were prepared for
 397 the IEPOX-OS measurements (with and without AS) at 50 mg/mL.

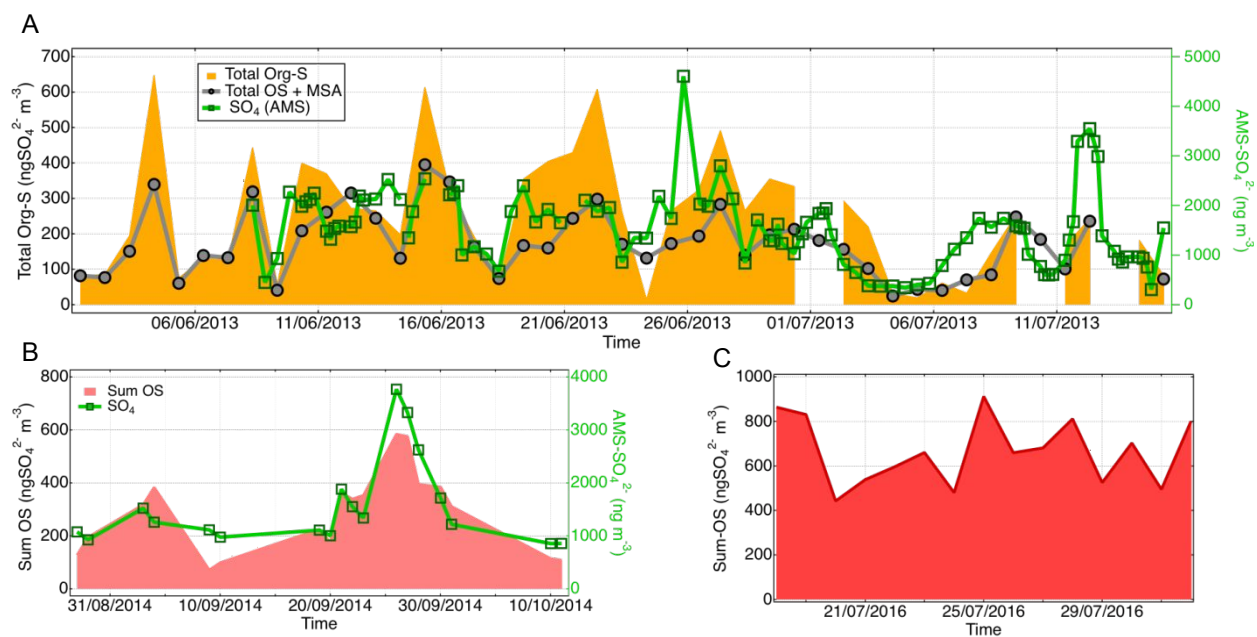
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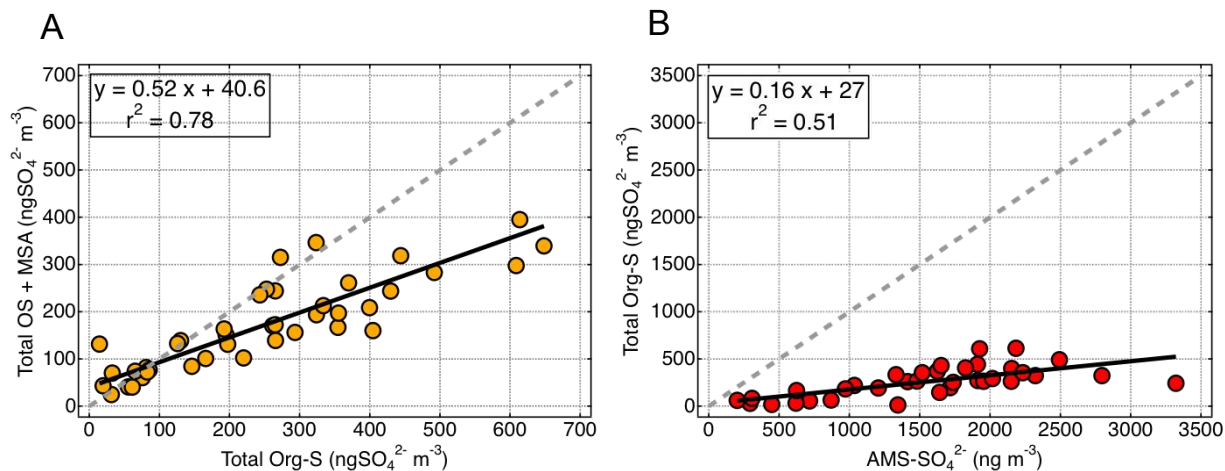
400 **Figure S3.** Comparison between aerosol sulfate measured by ICP-MS and AMS during the campaign in
401 the SE-U.S.

402

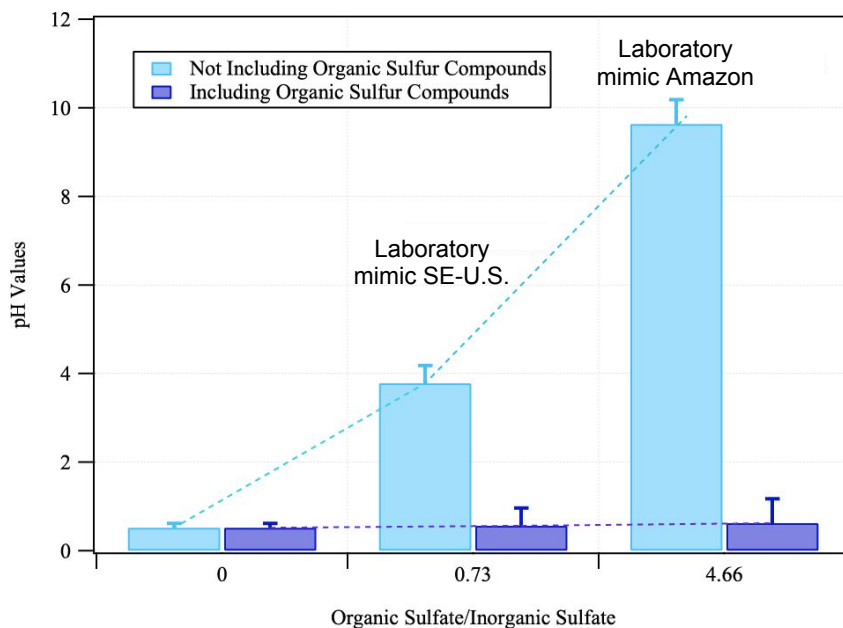


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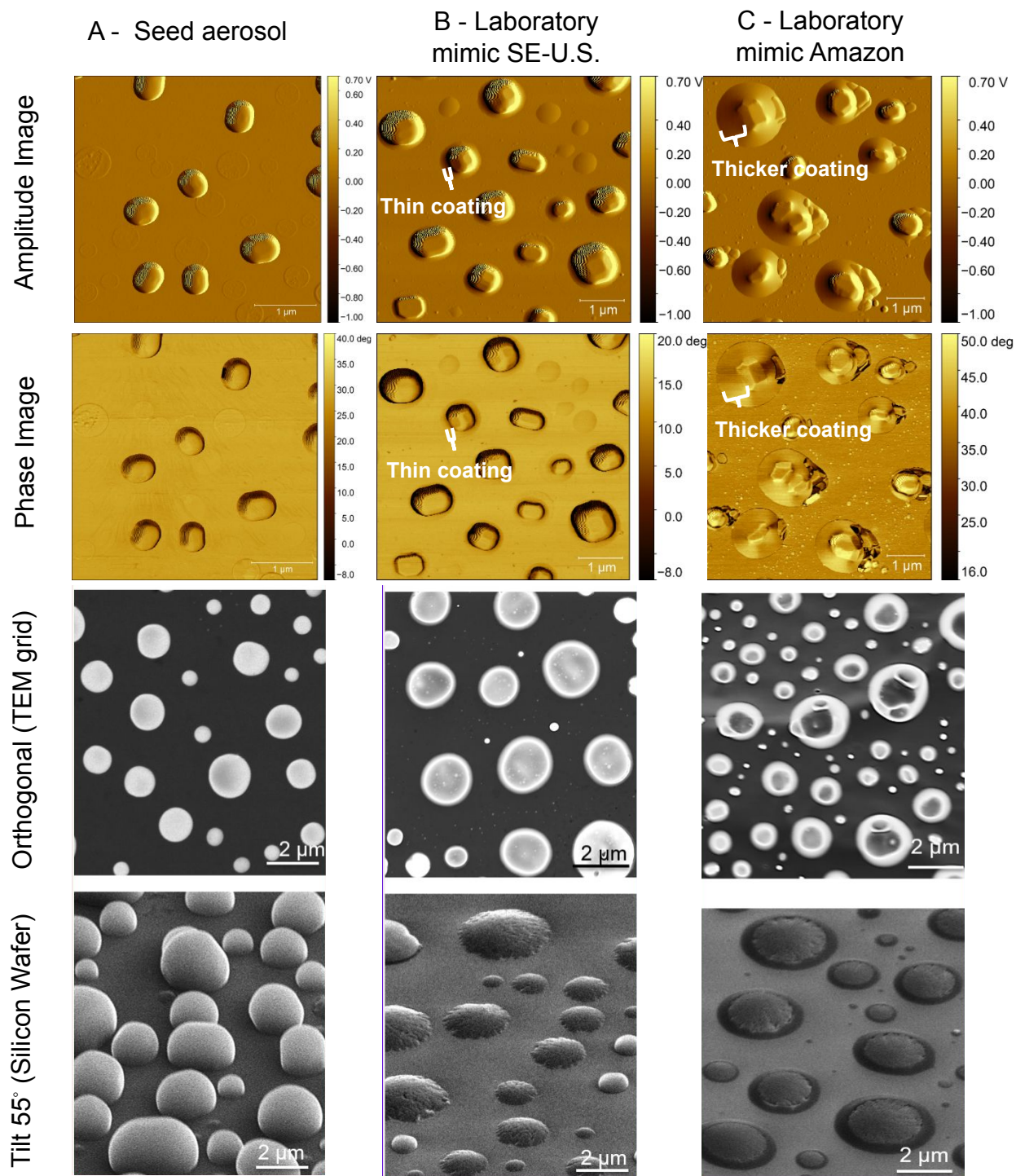
405 **Figure S4.** (A) Time-series of the sum of identified OS + MSA and total organosulfur compounds (Org-S)
 406 quantified by UPLC/ESI-HR-Q-TOFMS and IR-ICP-MS, respectively, in the $\text{PM}_{2.5}$ samples collected
 407 during the 2013 SOAS campaign. (B) and (C) present the concentrations of the sum of OS + MSA
 408 quantified in downwind Manaus and Manaus, respectively.



409
 410 **Figure S5.** (A) Correlation between the sum of OS + MSA (y-axis) and Org-S (x-axis). This correlation
 411 suggests significant unidentified sources of organosulfur compounds in SOA sampled in the SE-U.S. (B)
 412 Correlation between Org-S (y-axis) and the mass of sulfate measured by the AMS in SE-US. Dashed grey
 413 lines correspond to 1:1 line. OS + MSA were quantified by UPLC/ESI-HR-Q-TOFMS, and Org-S were
 414 quantified by IR-ICP-MS.
 415

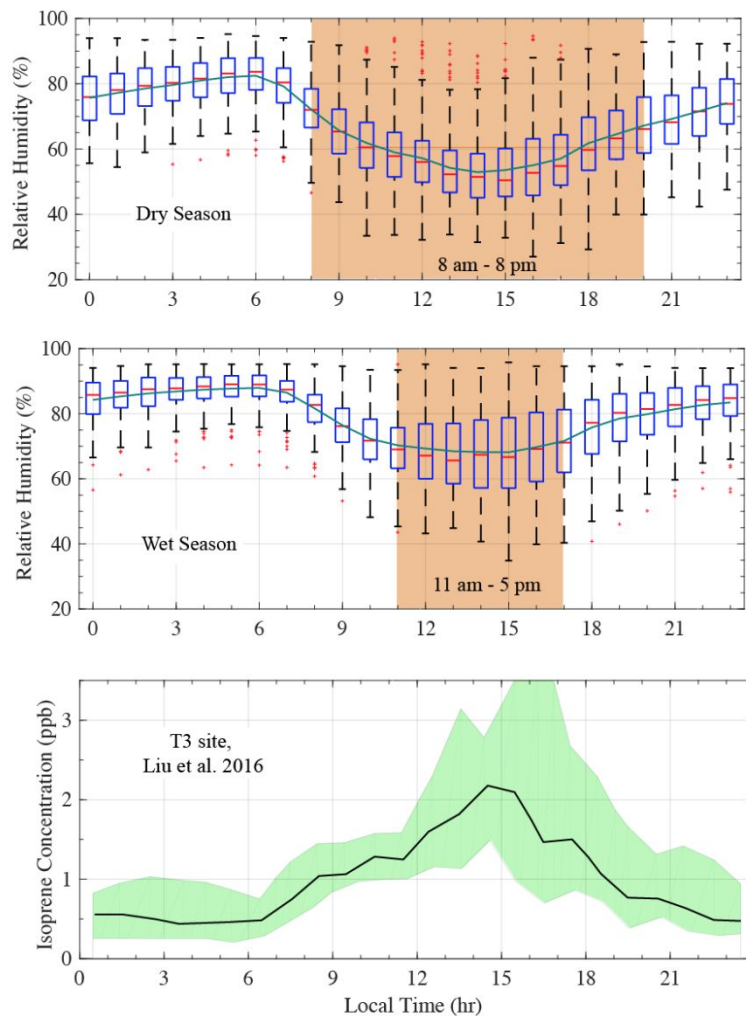


416
 417 **Figure S6.** Comparison of modeled aerosol pH when considering the acidity of IEPOX-OS vs. not
 418 considering IEPOX-OS. The x axis is the ratio of the total organic sulfur compounds vs. total inorganic
 419 sulfur compounds in the aerosol. The light blue bars represent the modelling results of the chamber data
 420 without considering IEPOX-OS. The purple bars represent is the modeling results after taking IEPOX-OS
 421 into consideration. The center and the right groups show the modelled pH based on aerosol composition
 422 measured at the end of the experiment from the SE-U.S. ratio and the Amazon ratio chamber experiments,
 423 respectively. The error bar range is calculated based on the pK_a values of IEPOX-OS to be 0-4. The dashed
 424 lines are trendlines. A growth factor of 1.3 and Henry's law constant for 0.0161 atm M⁻¹ are applied in the
 425 model. By varying the pK_a values of IEPOX-OS, the aerosol acidity changes up to 6 times (0.8 pH units),
 426 suggesting the importance of further research on the role of these SOA constituents in governing the aerosol
 427 acidity.

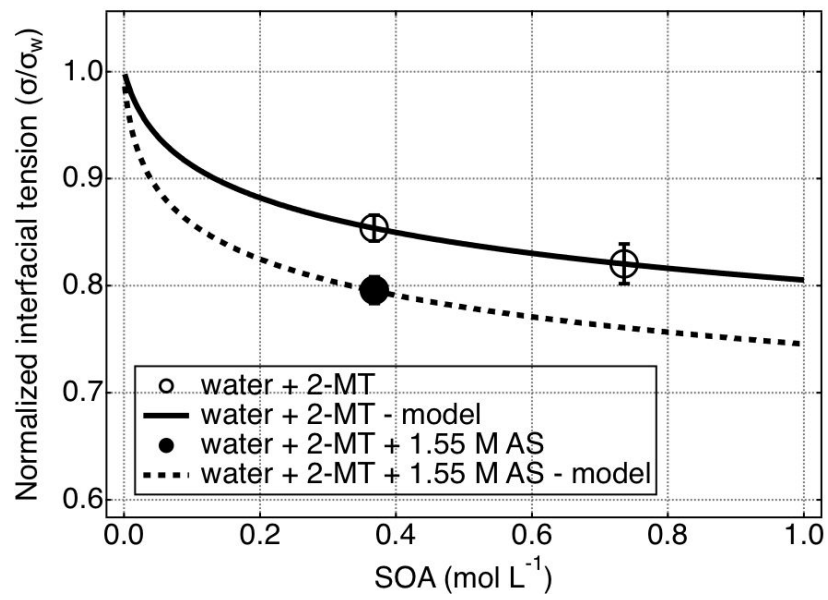


428
 429 **Figure S7.** AFM phase and amplitude images for the corresponding height images shown in Figure 3.
 430 Scanning electron microscopy images on TEM grids orthogonal to the electron beam (top) and at a 55
 431 degree tilt from the beam for (A) seed aerosol, (B) laboratory mimic SE-U.S. conditions and (C) laboratory
 432 mimic Amazon conditions (green). SE-U.S. and Amazon are after exposure to IEPOX.

433

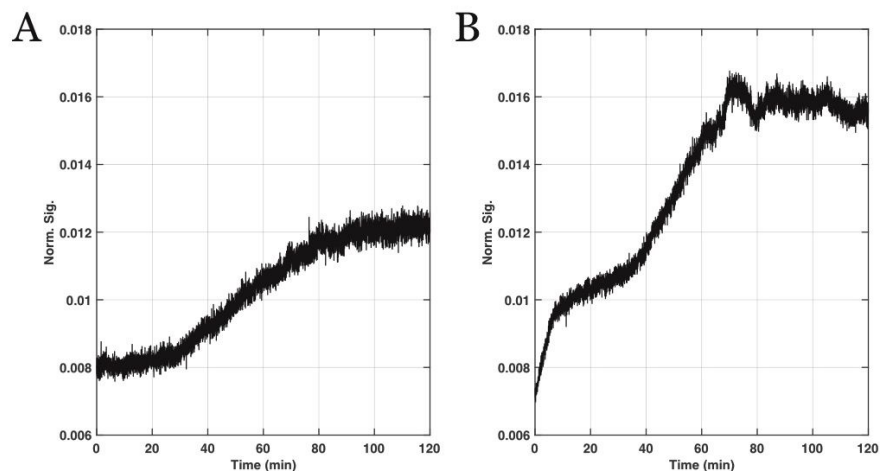


434
 435 **Figure S8.** The top two panels show the box plots of the hourly relative humidity during dry season (June-
 436 September) and wet season (December-March) from 2016-2018 at the Amazon basin (Manaus site). On
 437 each box, the central mark indicates the median, and the bottom and top edges of the box indicate the 25th
 438 and 75th percentiles, respectively. The whiskers extend to the most extreme data points not considered
 439 outliers, and the outliers are shown individually with red plus symbols. Boxplot assigns points as outliers
 440 if they are greater than $q_3 + 1.5 \times (q_3 - q_1)$ or less than $q_1 - 1.5 \times (q_3 - q_1)$, where q_1 and q_3 are the 25th and
 441 75th percentiles of the sample data, respectively. The black line in the bottom panel shows the measured
 442 hourly isoprene concentration downwind Manaus in the Amazon basin²⁰. For the stations, hourly RH values
 443 (calculated from measured temperature and dew point) were retrieved from NOAA's National Climatic
 444 Data Center (<http://www.ncdc.noaa.gov/>). The green lines in the top two panels represent the mean hourly
 445 RH. The orange-shaded areas represent the periods when the average RH < 70%.



446

447 **Figure S9.** Measured and modeled interfacial tensions for 2-methyltetrols (2-MT).



448

449 **Figure S10.** Time series of IEPOX signal in aerosol particles normalized by iodide signal measured by a
 450 high-resolution time-of-flight chemical ionization mass spectrometer equipped with iodide (I^-) reagent ion
 451 chemistry (I-HR-TOF-CIMS) during the reactive uptake of IEPOX chamber experiments under
 452 atmospheric IEPOX: SO_4^{2-} relevant to (A) SE-U.S. and (B) Amazon, corresponding to experiments shown
 453 in Figures 1A & 1B, respectively. Increasing slope of IEPOX signal indicates the injection rate of IEPOX
 454 was faster than its reactive and wall losses, a decreasing slope indicates reactive uptake to the particle phase
 455 is gradually slowing down as the reaction continues.

456
457
458

Table S1. Summary of indoor smog chamber conditions used for IEPOX uptake experiments using ammonium bisulfate (ABS) seed aerosols.

No	Description	Injected IEPOX (ppb)	SO ₄ ²⁻ (µg/m ³)	IEPOX/SO ₄	RH (%)
1	IEPOX + ABS	48	430.2	0.5	49.8
2	IEPOX + ABS	104	658.4	0.8	47.0
3	IEPOX + ABS	122	664.0	0.9	46.3
4	IEPOX + ABS	172	669.6	1.3	53.2
5	IEPOX + ABS	140	437.4	1.5	52.2
6	IEPOX + ABS	166	179.2	4.5	48.9
7	IEPOX + ABS	218	175.8	6.0	47.7
8	IEPOX + ABS	240	111.1	10.5	48.9
9	IEPOX + ABS	174	76.4	11.1	46.4
10	IEPOX + ABS	176	70.7	12.1	48.0
11	IEPOX + ABS	230	65.6	17.1	50.4
12	IEPOX + ABS	124	31.0	19.5	46.5
13	IEPOX + ABS	174	41.1	20.7	48.8

459

460 **Table S2.** Average and maximum concentrations of the OS identified during the 2013 SOAS campaign.

Suggested Formula	[M-H]⁻ ion	Average concentration (ng m⁻³)	Max concentration (ng m⁻³)	Average OS/Sum OS (%)	Precursor group
CH ₃ O ₃ S ^{-a,b}	94.98	39.27	167.59	10.53	Dimethylsulfide
CH ₃ O ₄ S ^{-a,c}	110.97	3.03	20.95	0.82	-
C ₂ H ₃ O ₅ S ^{-a,c}	138.97	0.83	2.51	0.26	Isoprene / Anthropogenic
C ₃ H ₅ O ₅ S ^{-a,c}	152.98	4.95	18.22	1.34	Isoprene / Anthropogenic
C ₃ H ₅ O ₆ S ^{-a,c}	168.98	12.83	35.33	3.10	Isoprene / Anthropogenic
C ₅ H ₉ O ₅ S ^{-a,d}	181.01	0.14	2.51	0.04	Isoprene
C ₄ H ₇ O ₆ S ^{-a,c}	182.99	8.60	23.51	2.30	Isoprene
C ₅ H ₉ O ₆ S ^{-a,e}	197.01	4.79	16.87	0.74	Isoprene
C ₄ H ₇ O ₇ S ^{-a,d}	198.99	10.72	43.07	2.90	Isoprene
C ₅ H ₁₁ O ₆ S ^{-a,e}	199.02	2.78	9.04	0.73	Isoprene
C ₈ H ₉ O ₄ S ^{-a,e}	201.02	1.39	4.17	0.37	Isoprene
C ₅ H ₇ O ₇ S ^{-a,e}	210.99	7.31	27.45	1.96	Isoprene
C ₅ H ₉ O ₇ S ^{-a,e}	213.00	7.84	23.99	2.10	Isoprene
C ₅ H ₁₁ O ₇ S ^{-a,e}	215.02	217.14	755.3	58.39	Isoprene
C ₅ H ₇ O ₈ S ^{-a,e}	226.98	4.25	14.12	1.14	Isoprene
C ₆ H ₁₁ O ₇ S ^{-a,e}	227.02	0.43	2.94	0.11	Isoprene
C ₅ H ₉ O ₈ S ^{-a,e}	229.00	6.91	21.33	1.82	Isoprene
C ₅ H ₁₁ O ₈ S ^{-a,e}	231.01	0.38	3.72	0.12	Isoprene
C ₇ H ₁₁ O ₇ S ^{-f,g}	239.02	4.20	14.22	1.29	Limonene
C ₁₀ H ₁₇ O ₅ S ^{-f,h}	249.08	0.93	2.56	0.26	Monoterpenes

Suggested Formula	[M-H] ⁻ ion	Average concentration (ng m ⁻³)	Max concentration (ng m ⁻³)	Average OS/Sum OS (%)	Precursor group
C ₅ H ₁₀ NO ₉ S ^{-f,g}	260.00	16.41	124.14	4.40	Isoprene
C ₉ H ₁₅ O ₇ S ^{-f,h}	267.05	1.39	3.36	0.37	Isoprene / Limonene
C ₁₀ H ₁₅ O ₇ S ^{-f,h}	279.05	1.02	3.41	0.30	Monoterpenes
C ₁₀ H ₁₇ O ₇ S ^{-f,h}	281.07	2.15	15.10	0.60	α-terpinene
C ₁₀ H ₁₆ NO ₇ S ^{-f,h}	294.06	4.24	88.69	1.17	Monoterpenes
C ₉ H ₁₄ NO ₈ S ^{-f,h}	296.04	0.53	6.76	0.15	Limonene
C ₁₀ H ₁₆ NO ₈ S ^{-f,h}	310.06	12.88	96.11	3.47	Monoterpenes
C ₁₀ H ₁₉ O ₉ S ^{-f,h}	315.07	0.36	1.44	0.19	Isoprene
C ₁₀ H ₁₄ NO ₉ S ^{-f,h}	324.04	1.12	13.56	0.31	Monoterpenes
C ₁₀ H ₁₆ NO ₉ S ^{-f,h}	326.05	1.47	15.89	0.41	Monoterpenes
C ₁₀ H ₂₁ O ₁₀ S ^{-f,g}	333.08	0.69	3.23	0.18	Isoprene
C ₁₀ H ₁₆ NO ₁₀ S ^{-f,h}	342.05	0.31	2.12	0.09	Monoterpenes
C ₁₀ H ₁₅ N ₂ O ₁₀ S ^{-f,h}	355.04	0.81	18.66	0.22	α-pinene
C ₁₀ H ₁₇ N ₂ O ₁₁ S ^{-f,h}	373.06	0.34	7.17	0.10	Limonene

461
462 ^a OS are eluted within the first 2 minutes; ^b quantified using methylsulfonic acid; ^c OS quantified
463 using 2-oxopropyl sulfate; ^d OS quantified using 2-methylglyceric acid sulfates; ^e OS quantified
464 using 2-methyltetrol sulfates; ^f OS are eluted during the elution gradient; ^g OS quantified using
465 propyl sulfate; ^h OS quantified using 3-pinanol-2-hydrogen sulfate; ⁱ OS quantified using octyl
466 sulfate.

467

468 **Table S3.** Average and maximum concentrations of the OS identified during the Go-Amazon
 469 campaign.

Suggested Formula	[M-H]⁻ ion	Average concentration (ng m⁻³)	Max concentration (ng m⁻³)	Average OS/Sum OS (%)	Precursor group
C ₃ H ₅ O ₅ S ⁻ ^{a,b}	152.98	29.9	69.6	4.8	Isoprene / Anthropogenic
C ₃ H ₅ O ₅ S ⁻ ^{a,b}	154.98	25.3	47.9	4.1	Isoprene / Anthropogenic
C ₃ H ₅ O ₆ S ⁻ ^{a,b}	168.98	19	45.6	1.4	Isoprene / Anthropogenic
C ₄ H ₇ O ₆ S ⁻ ^{a,b}	182.99	29.5	71.4	3.1	Isoprene
C ₅ H ₁₁ O ₆ S ⁻ ^{a,b}	199.02	101.0	283.4	16.3	Isoprene
C ₅ H ₁₁ O ₇ S ⁻ ^{a,b}	215.02	398.5	1450.2	64.2	Isoprene

470 ^a OS are eluted within the first 2 minutes; ^b OS quantified using 2-methyltetrol sulfates.

471

Table S4. Average and maximum concentrations of the OS identified in Manaus.

Suggested Formula	[M-H]⁻ ion	Average concentration (ng m⁻³)	Max concentration (ng m⁻³)	Average OS/Sum OS (%)	Precursor group
C ₂ H ₃ O ₅ S ^{-a,c}	138.97	19.8	25.2	1.3	Isoprene / Anthropogenic
C ₃ H ₅ O ₅ S ^{-a,c}	152.98	87.4	301.1	5.8	Isoprene / Anthropogenic
C ₃ H ₅ O ₅ S ^{-a,c}	154.98	87.0	122.8	5.8	Isoprene / Anthropogenic
C ₅ H ₇ O ₅ S ^{-f,g}	179.0	4.0	11.9	0.3	Cyclohexene
C ₄ H ₇ O ₆ S ^{-a,c}	182.99	47.1	120.7	3.1	Isoprene
C ₆ H ₁₁ O ₅ S ^{-f,g}	195.03	1.5	7.0	0.1	Cyclohexene/ Pinonaldehyde
C ₅ H ₉ O ₆ S ^{-a,e}	197.01	15.6	24.5	1.0	Isoprene
C ₅ H ₁₁ O ₆ S ^{-a,e}	199.02	58.9	84.2	3.9	Isoprene
C ₅ H ₇ O ₇ S ^{-a,e}	210.99	9.7	13.8	0.6	Isoprene
C ₅ H ₉ O ₇ S ^{-a,e}	213.00	100.5	137.8	6.7	Isoprene
C ₅ H ₁₁ O ₇ S ^{-a,e}	215.02	823.3	1159.7	55.0	Isoprene
C ₇ H ₁₁ O ₅ S ^{-f,g}	223.02	6.5	11.2	0.4	α-pinene
C ₅ H ₉ O ₈ S ^{-a,e}	229.00	20.1	27.1	1.3	Isoprene
C ₅ H ₁₁ O ₈ S ^{-a,e}	231.01	7.0	9.7	0.5	Isoprene
C ₇ H ₁₁ O ₇ S ^{-f,g}	237.04	6.1	8.3	0.4	α-pinene
C ₇ H ₁₁ O ₇ S ^{-f,g}	239.02	14.5	19.2	1.0	Limonene
C ₁₀ H ₁₇ O ₅ S ^{-f,h}	249.04	24.1	69.4	1.6	Monoterpenes
C ₁₀ H ₁₉ O ₅ S ^{-f,h}	251.09	7.6	19.7	0.5	Cyclodecane
C ₈ H ₁₃ O ₇ S ^{-f,g}	253.03	3.4	4.7	0.2	Isoprene

Suggested Formula	[M-H]⁻ ion	Average concentration (ng m⁻³)	Max concentration (ng m⁻³)	Average OS/Sum OS (%)	Precursor group
C ₅ H ₁₀ NO ₇ S ^{-f,g}	260.00	48.7	19.2	0.6	Isoprene
C ₁₀ H ₁₅ O ₇ S ^{-f,h}	279.05	9.6	13.9	0.6	Monoterpene
C ₁₂ H ₂₃ O ₅ S ^{-f,i}	279.12	6.3	10.9	0.4	Dodecane
C ₁₀ H ₁₇ O ₇ S ^{-f,h}	281.06	13.8	22.7	0.9	Monoterpene
C ₁₀ H ₁₆ NO ₇ S ^{-f,h}	295.06	39.0	56.0	2.6	Monoterpene
C ₉ H ₁₄ NO ₈ S ^{-f,h}	296.04	2.7	5.9	0.2	Limonene
C ₁₀ H ₁₉ O ₁₀ S ^{-f,h}	331.06	16.5	26.7	1.1	Isoprene
C ₁₀ H ₂₁ O ₁₀ S ^{-f,h}	333.08	50.1	90.6	3.3	Isoprene
C ₂₀ H ₃₉ O ₁₅ S ^{-f,g}	451.14	1.8	5.5	0.1	Isoprene

473 ^a OS are eluted within the first 2 minutes; ^b quantified using methylsulfonic acid; ^c OS quantified
474 using 2-oxopropyl sulfate; ^d OS quantified using 2-methylglyceric acid sulfates; ^e OS quantified
475 using 2-methyltetrol sulfates; ^f OS are eluted during the elution gradient; ^g OS quantified using
476 propyl sulfate; ^h OS quantified using 3-pinanol-2-hydrogen sulfate; ⁱ OS quantified using octyl
477 sulfate.

478

479 **Table S5.** Calibration factors and retention times of the OS standards.

Formula	Name	[M-H]⁻	Calibration factor	r²	Retention time(s) (min)
C ₄ H ₇ O ₇ S ⁻	2-methylglyceric acid sulfate	198.99	1.01 x 10 ⁶	0.99	1.8
C ₅ H ₁₁ O ₇ S ⁻	2-methyltetrol sulfate	215.02	1.02 x 10 ⁶	0.99	1.1/1.8
C ₂ H ₃ O ₆ S ⁻	Glycolic acid sulfate	154.96	1.09 x 10 ⁶	0.98	1.8
C ₃ H ₅ O ₅ S ⁻	2-oxopropyl sulfate	152.98	1.16 x 10 ⁶	0.99	1.8
C ₃ H ₇ O ₄ S ⁻	Propyl sulfate	139.00	3.68 x 10 ⁶	0.99	3.5
C ₉ H ₁₃ O ₆ S ⁻	3-pinanol-2-hydrogen sulfate	249.04	3.01 x 10 ⁶	0.90	8.6
C ₈ H ₁₇ O ₄ S ⁻	Octyl sulfate	209.08	1.10 x 10 ⁸	0.99	11.5

480

481

482 **Table S6.** The fitting parameters for the glass transition calculation at dry condition.

n_C^0	b_C	b_H	b_{CH}	$b_O (b_S)$	b_{CO}
12.13	10.95	-41.82	21.61	118.96	-24.38

483
 484 **Table S7.** Parameters Used for the Viscosity Calculation of IEPOX-derived OS. Values in the parathesis
 485 represent the upper and lower bound values.

Compound	Glass Transition Temperature of Dry SOA ($T_{g,org}$) (K)	Hygroscopicity (κ)	Fragility (D)	The Gordon– Taylor Constant (k_{GT}):
α -Pinene SOA	278.5 (268.5-300)	0.1 (0.1)	10 (10-20)	2.5 (2.5-3.0)
2-Methyltetrol Sulfate	298 (288-320)	0.12 (0.1-0.15)	13 (10-20)	2.5 (2.5-3.0)
IEPOX-derived OS Mixture	313 (303-330)	0.12 (0.1-0.15)	13 (10-20)	2.5 (2.5-3.0)

486
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