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2 **Supplementary Information for**

3 **Insect biomass decline scaled to species diversity: general patterns derived from a hoverfly** 4 **community**

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8 **This PDF file includes:**

9 Supplementary text

10 Figs. S1 to S7

11 Tables S1 to S2

12 References for SI reference citations

13 Supporting Information Text

14 Supplementary Methods.

15 **Determining the relative biomass contribution of Hoverflies.** In order to assess how much hoverflies contribute to the total biomass
16 (of all flying insects) in the malaise traps, we utilized two independent datasets. We used measurements of body length made
17 by the Krefeld Entomological Society (Axel Ssymank, all species in our data), and fresh-weight measurements kindly provided
18 by Nick Hofland (Radboud University). The fresh-weight data were collected in 2016 and 2017 and included in total 97
19 measurements, over 13 hoverfly genera). We then paired the two datasets by species (if unknown, by genus) and regressed the
20 log of the body weight to body length (intercept = -6.33, slope coefficient= 0.24). The resulting model coefficients were used to
21 allometrically predict the weight per individual in our data, which were subsequently summed over all individuals per year.
22 Based on these calculations, we predicted total hoverfly mass of 321.6 and 52.2 gram for 1989 and 2014, respectively. This
23 implied a relative contribution of 4.4% and 3.0% to the total flying insect biomass collected in the Wahnachtal malaise traps
24 in 1989 and 2014, respectively.

25 **Steps in deriving hypothetical scenarios of variation species decline rate.** In the main text, we describe three alternative hypothetical
26 scenarios of species decline rates. Here we describe the steps and assumptions that were made while designing these scenarios.
27 We started off with a rank-abundance curve that was similar to the observed rank-abundance curve of the hoverflies (Fig. 2B
28 in the main text), i.e. by using a zipf-mandelbrot distribution with arbitrary parameters of $\beta_0 = -1.5$ and $\beta_1 = 2$ (see
29 equation 4 in main text), for a pool of 200 species. The total hoverfly community was scaled so that the most abundant species
30 arbitrarily consisted of 1500 individuals. Next we defined the rate of decline in each of the three scenarios. For equal rates of
31 decline between species (scenario I) we set $\lambda_i = 0.2$, i.e. at 80% decline for each species i . In scenarios II and III we allowed
32 decline rates to scale linearly to species rank, where the relationship was negative in scenario II and positive in scenario III,
33 at arbitrary slopes of -0.015 and 0.020, respectively. Finally we scaled the resulting species decline rate vector in order to
34 achieve a total abundance loss of 80% (see Fig. 1A in main text). Using these three scenarios of hypothetical decline rates, we
35 proceeded in calculating persistence (equation 1), rank abundance distributions, and fraction of species lost (Fig. 1B,C,D),
36 under perfect and imperfect (at 40%) detection efficiency. We also examined how other diversity measures, i.e. Hill numbers
37 (1, 2) of orders 0, 1, 2, and 3, behave in the presence of the three scenarios of decline rates, and under imperfect detection. To
38 this end, we simulated 1000 hypothetical hoverfly communities (based on parameters as above) and for each community we
39 calculated the percentage change in Hill numbers of orders $q=0-3$ (see Fig. S1).

40 **Approximating average seasonal species availability.** If detectability of individual species is invariant during the season, i.e. they
41 are equally likely to be trapped on each of the sampling days, then the distribution of number of species in each pot could
42 be approximated in a straightforward manner by a sampling-without-replacement process, conditional on the accumulated
43 community data. However, hoverfly species are not likely to be active during the entire season, leading to non-uniform
44 detectability during the season. Formally, the number of species expected to be trapped in a single pot (\hat{s}) will depend on the
45 relative abundance of each species (N_i), the total abundance in the pot (N_j) and total species richness S , according to

$$\hat{s}(N_j, S) = \sum_i^S \left(1 - \frac{\binom{N-N_i}{N_j}}{\binom{N}{N_j}} \right)$$

46 (3) where $N = \sum N_j = \sum N_i$.

47 In equation 11, main text, we introduced a correction factor c , that measures the average availability of species during the
48 season. We used the following approach to obtain an estimate of c .

49 First we produced average daily total abundance per pot j (abundance per pot divided by exposure length) which we denote
50 as \hat{n}_j . We then calculated the expected number of species given total richness (S), total abundance (N) and relative species
51 abundance (N_i). Additionally, and for each pot, we calculated the expected number of species per day conditional on the
52 number of species seen in each pot (S_j).

$$\hat{s}_j^{(2)}(\hat{n}_j, S_j) = \sum_i^{S_j} \left(1 - \frac{\binom{N-N_i}{\hat{n}_j}}{\binom{N}{\hat{n}_j}} \right)$$

53 The relationship between $\hat{s}_j^{(1)}(\hat{n}_j, S)$ and $\hat{s}_j^{(2)}(\hat{n}_j, S_j)$ is linear, with zero intercept and slope $0 < c \leq 1$, because typically
54 $S_j \leq S$. The coefficient c is hence obtained as:

$$c = \frac{\hat{s}_j^{(2)}(\hat{n}_j, S_j)}{\hat{s}_j^{(2)}(\hat{n}_j, S)}$$

55 .

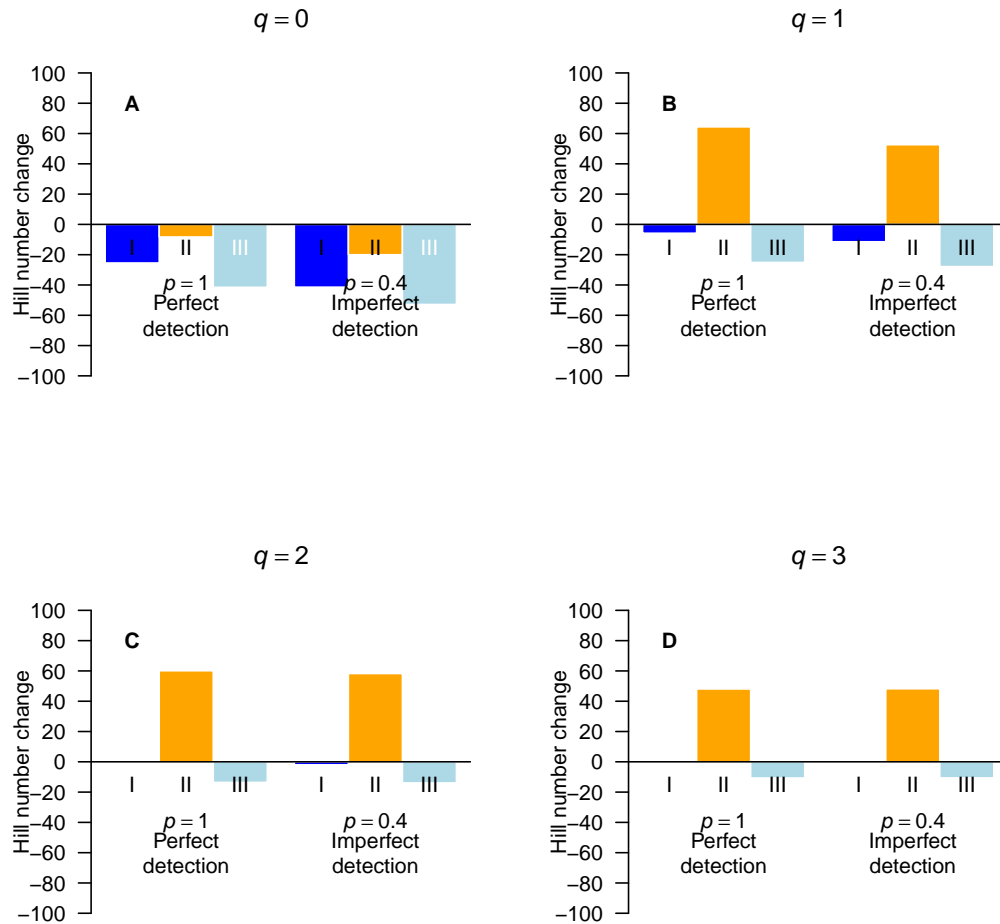


Fig. S1. Change in diversity measures (Hill numbers for orders $q \in \{0, 1, 2, 3\}$) in each theoretical scenario under perfect ($p=100\%$) and imperfect ($p=40\%$) sampling efficiency. A: Change in species richness ($q=0$), B: Change in Shannon diversity ($q=1$), C: Change in inverse Simpson index ($q=2$), D: Change in a higher-order diversity measure ($q=3$).

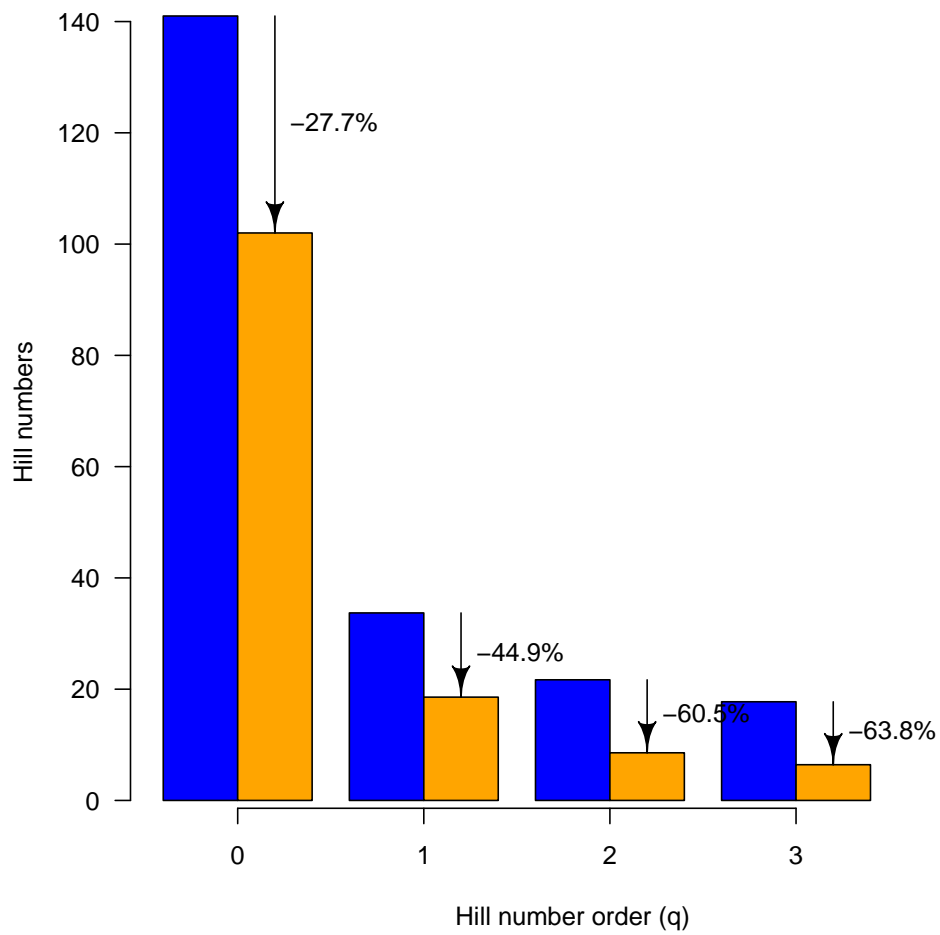


Fig. S2. Hill numbers of order 0-3 for 1989 (blue bars) and 2014 (orange bars), with accompanying amount of decline between the years. Orders of 0-2 denote species richness, exponent of Shannon entropy, and Simpson diversity, respectively, while for $q=3$ emphasis is placed predominantly on the more common species in the assemblages.

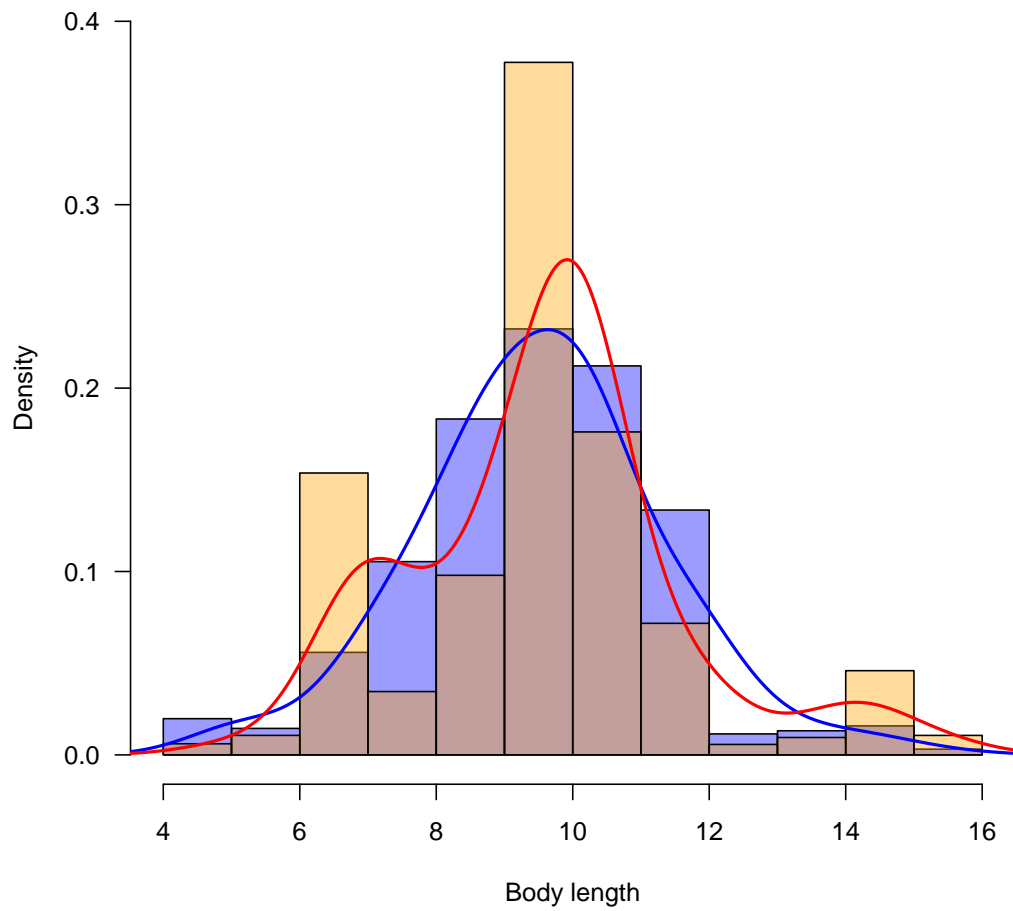


Fig. S3. Distribution of body length weighted by specie's abundances for 1989 (blue) and 2014 (red)

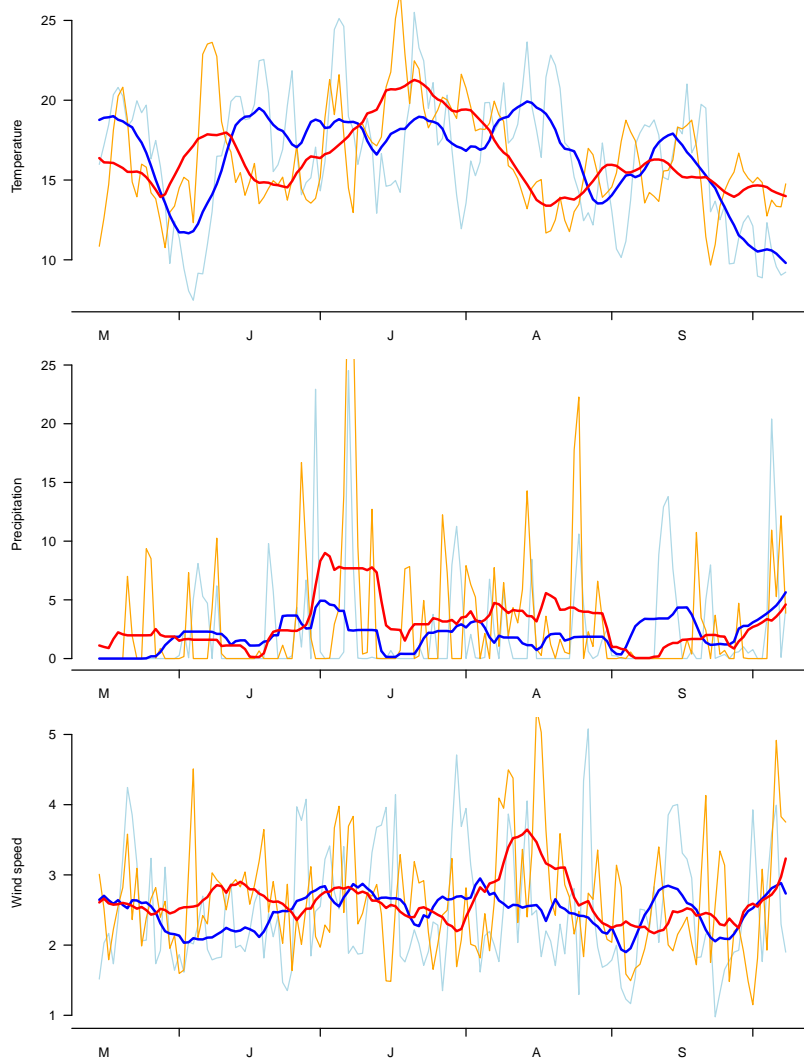


Fig. S4. Climatic variables in 1989 (light blue) and 2014 (orange) for temperature (in C°), precipitation (mm/day) and wind speed (m/s). Thick red and blue lines represent the 2-week moving average.

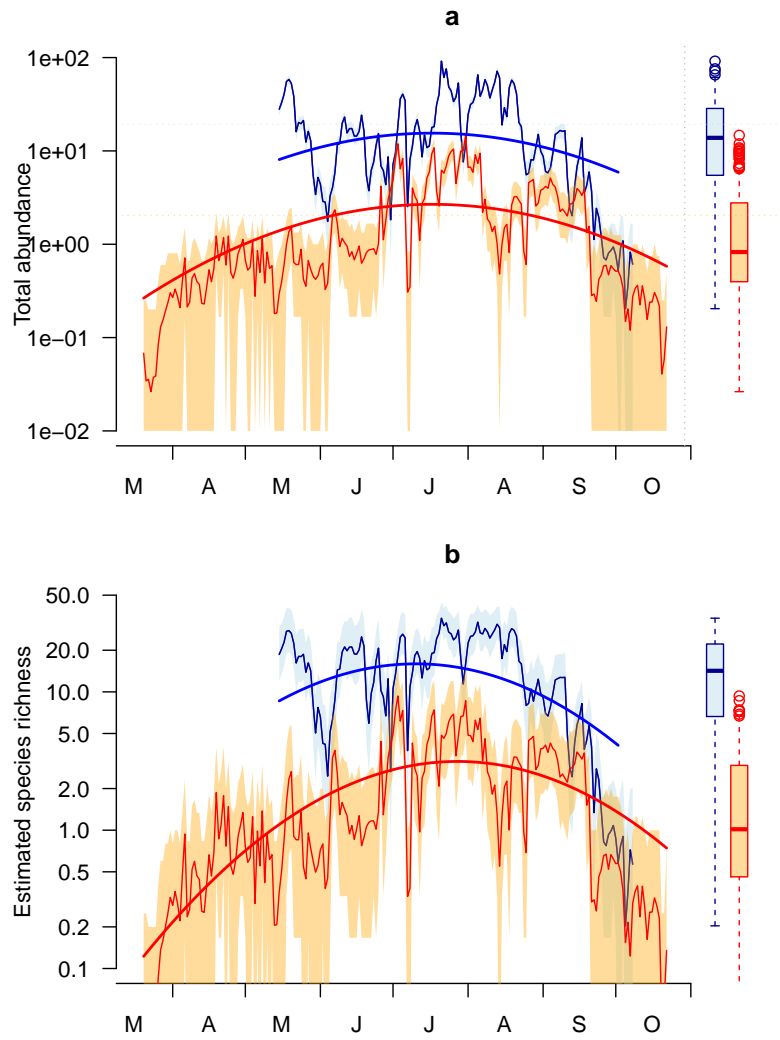


Fig. S5. Seasonal trajectory of estimated number of hoverfly individuals (A) and species (B) in 1989 (blue) and 2014 (red) along with 95% credible intervals. Boxplots provide the distribution of the mean daily values over the two seasons.

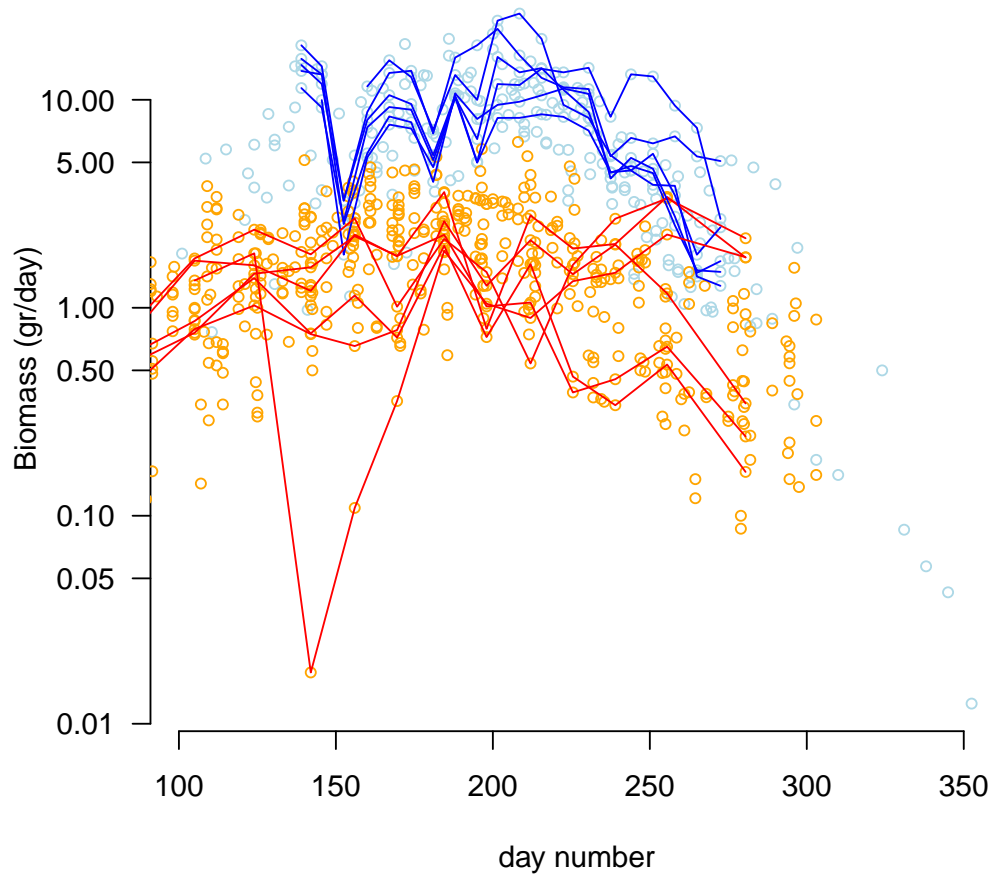


Fig. S6. Temporal distribution of biomass (in gram per day) of total flying insects for all pots in the period 1989-1992 (light blue dots) and period 2012-2015 (orange dots). Blue and red lines depict the seasonal biomass distribution for the six Wahnbachtal traps examined in 1989 and 2014

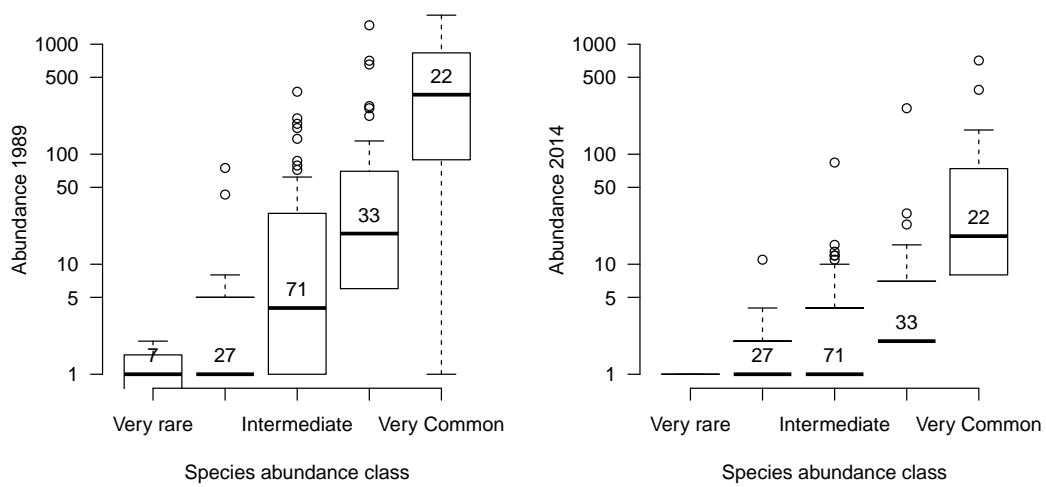


Fig. S7. Observed abundance (sum of 1998 and 2014 by species) versus abundance-class of species in Germany as classified in (4). Numbers inside boxplots represent the number of species in that class

Table S1. Parameter estimates from posterior distribution for daily total hoverfly abundance. *d*: climatic parameters. *c*: seasonal (quadratic effect) parameters, *b*: trap effects, and $\log(\lambda)$: the log-rate of decline from 1989 to 2014.

	mean	sd	2.5%	97.5%	\hat{R}
Intercept	2.477	0.027	2.424	2.529	1.002
$\log(\lambda)$	-1.756	0.028	-1.808	-1.697	1.001
<i>c</i> ₁	0.090	0.014	0.063	0.116	1.001
<i>c</i> ₂	-0.480	0.019	-0.516	-0.443	1.002
<i>c</i> ₃	0.476	0.033	0.412	0.541	1.001
<i>c</i> ₄	-0.614	0.035	-0.683	-0.548	1.001
<i>d</i> ₁	0.590	0.013	0.564	0.615	1.001
<i>d</i> ₂	-0.367	0.032	-0.432	-0.310	1.001
<i>d</i> ₃	-0.048	0.023	-0.094	-0.003	1.001
<i>b</i> ₂	0.318	0.027	0.264	0.371	1.001
<i>b</i> ₃	0.024	0.028	-0.031	0.082	1.002
<i>b</i> ₄	0.631	0.025	0.583	0.678	1.001
<i>b</i> ₅	0.629	0.025	0.581	0.678	1.001
<i>b</i> ₆	-0.050	0.029	-0.107	0.007	1.001

Table S2. Parameter estimates from posterior distribution for daily hoverfly species richness. *d*: climatic parameters. *c*: seasonal (quadratic effect) parameters, *b*: trap effects, and $\log(\lambda)$: the log-rate of decline from 1989 to 2014.

	mean	sd	2.5%	97.5%	\hat{R}
Intercept	2.748	0.048	2.656	2.846	1.002
$\log(\lambda)$	-1.671	0.040	-1.750	-1.592	1.001
<i>c</i> ₁	-0.036	0.033	-0.101	0.029	1.002
<i>c</i> ₂	-0.571	0.041	-0.652	-0.491	1.002
<i>c</i> ₃	0.325	0.024	0.277	0.373	1.001
<i>c</i> ₄	-0.568	0.029	-0.627	-0.514	1.001
<i>d</i> ₁	0.349	0.019	0.311	0.385	1.003
<i>d</i> ₂	-0.271	0.030	-0.331	-0.212	1.002
<i>d</i> ₃	-0.010	0.023	-0.054	0.035	1.001
<i>b</i> ₂	0.031	0.054	-0.076	0.139	1.002
<i>b</i> ₃	-0.112	0.053	-0.217	-0.009	1.004
<i>b</i> ₄	0.185	0.051	0.081	0.281	1.001
<i>b</i> ₅	0.117	0.051	0.019	0.215	1.003
<i>b</i> ₆	-0.100	0.059	-0.215	0.016	1.003

58 References

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