Supplementary Information for: *Ectomycorrhizal access to organic nitrogen mediates CO² fertilization response in a dominant temperate tree*

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Supplementary Methods:

Five soil cores, 5-cm diameter and 10-cm deep, were collected both May and August 2018, were taken radially around the dripline of each focal *Q. rubra* individual. Cores were immediately transferred to the laboratory on ice, sieved at 2 mm, and homogenized for each tree. Net N mineralization rates were calculated by averaging technical replicates, and summing the concentrations of NO_3^- and NH_4^+ prior to and following incubation; net N mineralization was the difference between these values¹. All analyses presented here utilize August 2018 rates of net N mineralization due to several May incubations being disturbed. Total free primary amines (TFPA) in soil (primarily amino acids and amino sugars) was measured using fresh sieved soil extracted with 2M KCl. TFPA is expressed as umol leucine equivalents, because leucine was used as the analytical standard estimates of TFPA availability may be considered relative indices of labile organic N availability in soil solution, which is distinct from N-SOM. Extracts used for TFPA quantification were identical to those used to measure extractable $NO₃$ and $NH₄$ ⁺ prior to aerobic incubation. Contents of dried sieved bulk soil (% of dry mass) were determined using combustion analysis on a LECO TruMac CN analyzer (LECO Corporation, St. Joseph, MI, USA) for soil collected in May 2018. Soil pH was determined for 2:1 deionized water-soil slurries with an Accumet 15 pH meter for soil collected in August 2018 (Fisher Scientific, Waltham, MA, USA). See $²$ for all values.</sup>

Fungal cores were obtained immediately adjacent to the soil cores described above. ECM cores were pooled for each individual focal tree and within 12 days of sampling, definitive ECM root-tips with high turgor were manually excised using a dissecting microscope after visually eliminating non-*Quercus* roots. Sampling was standardized by visually assessing the tips of \sim 90% (wet weight) of all *Quercus* roots in each of the root cores.

Molecular characterization of ECM communities is described in 2 . CTAB from each ECM root-tip sample was removed prior to lyophilization at -50°C. Lyophilized root-tips were homogenized and the totality of each root-tip sample was split into 20-25mg components and each component was placed in a lysis-tube with three sterilized 2.38mm metal beads. Each lysis tube also contained 800µl of Buffer AP1 and 4µl of RNase A from a Qiagen DNeasy Plant Mini Kit. Tubes were vortexed and placed in a 65°C waterbath for 20 minutes. DNA was then extracted using the Qiagen DNeasy Plant Mini Kit following manufacturers recommended protocol. DNA extraction replicates were combined and DNA yield was assessed using gel electrophoresis. Assessment of DNA quality was conducted using a Nanodrop Spectrophotometer (Thermo Fisher). The Quant-iT PicoGreen dsDNA Assay Kit (LifeTechnologies) and a BioTek SynergyHT Multi-Detection Microplate Reader (BioTek Instruments) were used to quantify DNA concentrations prior to PCR.

The ITS2 region was amplified using Illumina dual-indexed primers 5.8S Fun and ITS4 Fun³. The forward and reverse primer each contained the appropriate Illumina Nextera adaptor, linker sequence and error correcting Golay barcode for use with the Illumina MiSeq platform. All PCRs were performed in triplicate following Taylor et al. (2016), using Phusion High Fidelity DNA Polymerase and master mix (New England BioLabs). Each PCR contained 6 μl High Fidelity Phusion $5 \times$ buffer, 0.75 μl each primer (10 μm initial concentration), 0.42 μl dNTPs (20 mmol^{-1}) initial concentration of each dNTP), 1.5 μl of template DNA (mean concentration 3.76 ng/μl, SD=2.82) and 0.23 μl of Taq (2 U/μl) brought to a final volume of 20 μl with molecular-grade water. PCR conditions consisted of an initial denaturation step at 94°C for 3 min, followed by 27 cycles of the following: 30 s at 94°C, 45 s at 57°C and 90 s at 72°C followed by a final extension step of 72°C for 10 min. PhiX oligonucleotides were spiked for

base diversity.

Taxonomic identification of fungi encountered in this study are described in full elsewhere 2 . The DADA2 pipeline was implemented in QIIME 2 in order to denoise sequences, detect and remove chimeras and remove PhiX contaminants and infer representative sequences named absolute sequence variants $(ASV)^4$. A maximum of 2 expected errors $(MaxEE = 2)$ was allowed. ASV were inferred using a total 6,869,462 of filtered forward sequences (mean= 5.64 x $10⁵$, *SD*= 1.39 x 10⁵ sequences per sample). ASV were assigned taxonomy using the dynamic (97-99% sequence similarity) UNITE database $(v.8)^5$ and the scikit-learn naive Bayes machinelearning algorithm⁶. This dynamic classification system captures known variation among fungal clades in delimited species sequence similarity $⁷$. ASV were collapsed at the finest possible</sup> taxonomic level using the *taxa collapse* command in QIIME 2. Taxa that could not be assigned to Fungi, and appeared less than twice across all samples were removed.

After filtering and processing, sequence-based rarefaction curves were highly asymptotic implying that sequencing depth was adequate to capture the diversity of fungi encountered in our samples². The ectomy corrhizal status of fungal genera was assigned using literature searches⁸; fungal genera with mixed, unidentified, or non-ECM status were removed from subsequent analyses (~20% of overall sequences). Individual root-systems hosted a mean of 27 ECM OTU $(SE = 0.95)$, and an average of 80% sequences per sample consisted of ECM taxa (SE = 0.020%). This is almost certainly a slight under-estimation of per-sample ECM sequence abundance, because fungal taxa with questionable or uncertain biotrophic associations, although detected from ECM root-tips samples, were not scored as ECM⁹.

We used the DEEMY (characterization and DEtermination of EctoMYcorrhizae) database (http://www.deemy.de/) to gather trait information on the exploration type (hyphal foraging distance) and rhizomorph formation of ECM fungal species present in our dataset. When fungal species in our study were not represented in DEEMY, congeners were surveyed and, if 90% of the entries agreed, consensus trait values were assigned to that taxon. This classification system is supported by the fact that foraging-related functional traits for fungal hyphae are typically conserved at the genus level ¹⁰. This also allowed incorporation of ECM taxa that could only be identified to genus 11 . Long-distance foraging types were rare in our study system, composing less than 7% of ECM-derived sequences in each sample ($SE = 1.38$) and were removed from subsequent analyses. We were able to assign morphological hyphal trait data for 28 ECM genera comprising more than 93% all identified ECM sequences. The proportion of ECM sequences assigned morphological attributes using DEEMY did not vary across the soil gradient ($P = 0.50$).

Threshold Indicator Analysis was used to detect ECM community level threshold responses – community change points- along the continuous soil N mineralization gradient, using the TITAN2 package in $R¹²$. Only validated ECM genera were used in this analysis, and taxa that occurred in less than five samples, less than five times per sample were removed following ¹². Community-level changes are strongest where either sum (z−) or sum (z+) reaches a maximum. Evidence for a community-level threshold is obtained when (a) many species exhibit similar change points, (b) a large maximum z-score occurs relative to sums elsewhere on the gradient and (c) z-score maxima across bootstrap replicates occupy a relatively narrow range of environmental values. We conducted this analysis after Hellinger transformation using 2000 bootstrap and permutation replicates. We used relative cutoff and threshold scores of 0.85, Indvals were calculated using the relative abundance obtained by the ratio of summed abundance in each partition to the total, to address skew.

Prior to metagenomic sequencing library preparation, DNA extracts were quantified (Agilent 4200 TapeStation; Santa Clara, CA). 40ng of input DNA was used for library construction, however six of the 60 samples had lower total DNA yield. For these samples, the totality of all DNA was used. Libraries were then custom sheared using a Covaris S2 Focused-Ultrasonicator Woburn, MA), to a target of 200 bp (duty =10%, intensity = 5, cycles/burst = 200, time =120 seconds); previous trials confirmed the efficacy of these settings. Libraries were prepared using the NEB Next Ultra 2 DNA Library Prep kit (New England Biolabs) with seven cycles of PCR. 59 out of the 60 samples successfully yielded libraries suitable for sequencing. Sequencing was conducted using an S4 flow cell of the Illumina NovaSeq 6000 instrument.

In total, 23,203,326,006 sequences were generated. Reads were then dereplicated, adapters trimmed, sequence Q >20 retained, and reads shorter than 40 bp were removed using BBDuk (jgi.doe.gov). 23,177,098,622 paired-end reads passed initial quality filtering. We then used an additional filtering step to remove non-fungal sequences using Kraken2 paired-end mode with default parameters 13 . Sequences were mapped against the standard Kraken2 database containing bacterial, archaeal and UniVec reads (containing sequencing adapters, linkers, and primer sequences), and further supplemented with sequences obtained from published *Quercus rubra* ¹⁴ and *Qurcus lobata* genomes ¹⁵ in order to remove plant sequences (contaminants). All mapped reads were removed. On average, 21.7% of sequences per sample were removed during this Kraken2 filtering step, and the mean number of sequences remaining in each sample after Kraken mapping was 307,041,274

Filtered reads were mapped to functional reference gene databases CAZy (accessed March 2019)¹⁶ and Peroxibase (accessed February 2019)¹⁷. Translated reads were mapped to CAZy using 'sensitive' mode in DIAMOND v. 0.9.29, with an -e value: $1e^{-4}$, following best practices for unmerged reads ¹⁸. BWA-MEM was used to map sequences to representatives downloaded from Peroxibase, using default settings ¹⁹. The number of mapped reads was averaged for unmerged forward and reverse reads for each reference gene to avoid double counting, the geometric mean of all mapped reads were then averaged across all reference sequences for a given gene family. Note that gene counts were not normalized to gene size (for example, average length of each gene in CAZy) because this was unnecessary: comparisons of CAZy relative abundances were primarily to environmental parameters and not to each other.

We tabulated the number of near-single copy genes, as a proxy for the number of Dikaryotic fungal genomes present in each sample, using the OrthoDB v.9 orthologous ancestral gene database, which comprised 1312 near-single copy gene variants 20 Filtered forward and reverse reads for each sample were mapped to the Dikaryotic OrthoDB database of 1312 orthologs using 'sensitive' DIAMOND as above. Mapped reads to each ortholog were averaged to prevent double-counting. Dikaryotic near single-copy genes were chosen because the majority (>95% sequences) are Dikaryotic. Instead of relying on a single arbitrarily chosen house-keeping gene that may not be at true single-copy in complex environmental samples, we calculated the geometric mean number of 'single-copy' genes present across all orthologs and the standard error of orthologous gene counts for each sample.

The following analyses were run in JAGS 3.4 21 using the rjags package in R (R Development Core Team 2013). Three chains were run until convergence of the parameters, ~50,000 iterations, and run again for another 50,000 to estimate posterior parameter means, variances and covariances, after thinning every 100th iteration.

BAI analysis: rjags analysis code

```
model{ 
#missing temp 
for(i in 1:38){MayMinTemp[i]~dnorm(0,1)} 
  for(i in 2:55){ #individuals 
    J[i]<-step(miner[i]-cp) 
   for(y in 72:109){ #years 1980-2017 
   bai[y,i]~dlnorm(D[y,i],tau1[y,i]) 
   bai.h[y,i]~dlnorm(D[y,i],tau1[y,i])#predicted 
   D[y,i]<- (alpha[1]+alpha[2]*J[i])+(alpha[3]+alpha[4]*J[i])*(miner[i])+alpha[5]*log(dbh[y,i])+ 
alpha[6]**baiS[y-1,i] alpha[7]**MayMinTemp[y-71] +SpatRE[IDSpatial[i]] 
  tau[y,i] < -1/(a+b^*log(dbh[y,i])) } 
  } 
  #priors 
  cp~dunif(0,1.25) 
  a~dlnorm(1,0.001) 
  b~dnorm(0,0.001) 
  for(i in 1:6){ 
   alpha[i]~dnorm(0,0.001) 
   } 
   ccomb<-c[1]+c[2] 
   acomb<-a[1]+a[2] 
    #spatial effects 
    SpatRE[1:55]~spatial.exp(mu[1:55],lon[1:55],lat[1:55],tauS,phi,1) 
   tauS~dgamma(0.001,0.001) 
    phi~dunif(0.001,10) 
   for(i in 1:55){ mu[i] <- 0 }
}
```
Analysis of GNES:

model{

```
 for(i in 1:2051){
```

```
 #missign values 
 MayMinTemp[i]~dnorm(0,1)
```

```
 GNEbaiS[i]~dnorm(B[i],tau1[Indv[i]]) 
 GNEbaiS.h[i]~dnorm(B[i],tau1[Indv[i]]) #predictions
```
B[i]<- beta[Indv[i]]+lambda[Indv[i]]*CO2[i]+gamma[Indv[i]]*MayMinTemp[i]+SpatRE[IDSpatial[i]]#

}

```
 #priors 
  for(i in 2:55){ 
   beta[i]~dnorm(0,0.001) 
   lambda[i]~dnorm(0,0.001) 
   gamma[i]~dnorm(0,0.001) 
   tau1[i]~dgamma(0.0001,0.0001) 
   } 
    SpatRE[1:55]~spatial.exp(mu[1:55],lon[1:55],lat[1:55],tauS,phi,1) 
    tauS~dgamma(0.001,0.001) 
    phi~dunif(0.001,10) 
    for(i in 1:55){ mu[i]<-0 } 
  } 
Analysis of the slopes: 
model{ 
  for(i in 2:NN){ 
  J[i]<-step(minerS[i]-cp) 
  Ltau[i]<-1/(Lsd[i]*Lsd[i]) 
  Lambdamean[i]~dnorm(L [i],Ltau[i]) 
  Lambdamean.h[i]~dnorm(L[i],Ltau[i]) 
  L[i]<-(theta[1]+(theta[2]*J[i])+(theta[3]+theta[3]*J[i])*(minerS[i]) 
  } 
  #priors 
 for(i in 1:4)\{ theta[i]~dnorm(0,0.001) 
   } 
   Thetacomb[1]<-theta[1]+theta[2] 
   Thetacomb[2]<-theta[3]+theta[4] 
   cp~dunif(0,1.25)
```
}

Supplementary Figures

Supplementary Figure 1. Map of the twelve forest sites in Wexford and Manistee Counties, Manistee National Forest, Michigan, USA. All trees lay between 70-130m elevation. Pins are colored by rates of net N mineralization (μ g g⁻¹ d⁻¹). Satellite imagery May 2018. Insets: Michigan USA with sites in red; continental United States and location of Michigan, blue box. Study area is extensively described in^{1,2}. Satellite imagery is derived from Google Earth.

Supplementary Figure 2: Bayesian model fits, goodness of fit (predicted versus observed [in our case calculated data]) for our three dendrochronological analyses of plant growth. Solid line indicates the 1:1 relationship between the two variables. BAI = Basal Area Increment. GNES= standardized Growth Nitrogen Efficiency. See methods for calculation and estimations of parameters.

Supplementary Figure 3: Top panel, community change points where compositional change in ECM taxonomic membership is greatest. ECM taxa are grouped at the genus level. Blue indicates genera with $(z+)$ scores, red indicates genera with $(z-)$ scores. Community change point = 0.47μ g g⁻¹ d⁻¹ (z+) and 0.45 μ g g⁻¹ d⁻¹ (z-) scores. The observed (z+) and (z-) maxima are plotted as circles with the 95th percentile of their distributions as horizontal lines. The bottom panel shows the estimated probability densities across all booststrap replicates.

Supplementary Figure 4. NMDS plot of ECM fungal communities (points) colored by soil mineralization rates (μ g N per gram dry soil per day; legend bar). The vectors indicate direction and degree of correlation between the two first NMDS axes and ECM fungal morphotypes. Medium and Short indicate hyphal exploration types and Contact and Rhizomorphic indicate rhizomorph presence. Location of plotted genus names are scaled centroid coordinates for abundant ECM fungal genera.

Supplementary Figure 5: Number of colonized ectomycorrhizal (ECM) fungal root-tips on northern red oak individuals across the studied soil gradient. Linear regression model: R^2_{adj} = 0.25, P = 0.0001 . Method = 'qr'. Coloured band indicates 95% confidence interval.

Supplementary Figure 6. Freeze-dried weight (mg) of ectomycorrhizal (ECM) fungal root-tips collected from northern red oak individuals across the studied inorganic N gradient: Linear regression model. Method = 'qr'. $R^2_{\text{adj}} = 0.10$, P= 0.009. Coloured band indicates 95% confidence interval.

Supplementary Figure 7. Principal Component Analysis (PCA) for the abundances of the 100 decay genes studied here using shotgun metagenomic sequencing. Gene abundances were standardized using genes present at near single genome copies, and then weighted by the number of root-tips present in each sample and log-transformed. Points represent individual ECM communities colored by rates of net N mineralization (μ g inorganic N g⁻¹ d⁻¹).

Supplementary Figure 8: Fold change values for all 100 gene families above and below the statistical BAI change point $(0.53 \mu g$ inorganic N $g^{-1} d^{-1}$). Each gene family was evaluated at a greater relative abundance below the statistical change point than above. Dashed line indicates equal gene counts above and below the change point.

Supplementary Figure 10. A). Enrichment of (log₂ fold change) CAZy gene families below the statistical change point, relative to those occurring above the changepoint (basal area increment (BAI) change point). Gene families with greatest fold changes presented. Gene family names presented in Supplementary Table 2 with statistical information presented in Supplementary Table 3 **B).** Results from generalized dissimilarity model (GDM) documenting threshold response of ECM community aggregated decay potential for all 100 gene families studied here, measured as the partial ecological distance (y-axis) in response to empirical supply rates of inorganic N availability (Net N Mineralization). The slope of the blue line shows the rate of compositional metagenomic change along the soil gradient. Vertical dashed line denotes independently derived dendrochronological BAI changepoint (Figure 2; 0.53 μ g inorganic N • g soil⁻¹ • day ⁻¹). The near-zero slope to the left of the vertical dashed line denotes negligible compositional change in the decay attributes of ECM communities along this portion of the soil gradient.

Supplementary Figure 11: Relationship between Spring and Fall mineralization rates for soils collected 2018. Paired samples are derived from the base of the same individual trees (Linear regression model: R^2_{adj} : 0.58. P < 0.00001). Method = 'qr'. Black line is 1:1 plot. Several samples were disturbed during the Spring 2018 incubation, accounting for the fewer total points plotted. Colored band indicates 95% confidence interval.

Supplementary Figure 12. Neighborhood overstory plant communities. Relative abundance of plant stems greater than 10cm within 10m radius of each focal *Quercus rubra* individual along the soil Net N mineralization gradient. Only stems at greater than 1% relative abundance shown.

Supplementary Figure 13. Overstory stand density for the 'neighborhood' surrounding each focal *Quercus rubra* individual (i.e. within 10m radius). No significant differences along the soil mineralization gradient ($P = 0.64$). Linear regression model, method = 'qr'. Shaded band denotes 95% confidence interval.

Supplementary Figure 14. Metagenomic sequencing yield for each sample. QC (red) represent quality filtered metagenomic reads, see main text. Kraken Unmapped (blue), represent reads that remain after Kraken filtering against plant and contaminant databases (putative fungal reads). No significant relationships across the soil gradient: linear regression: QC: *P =* 0.36. Kraken Unmapped: $P = 0.73$. Linear regression model, method = 'qr'.

Supplementary Tables:

Supplementary Table 1: PCR primers employed in this study.

Supplementary Table 2: Gene families, enzymes they encode and putative substrates. Sourced from (CAZy: [http://www.cazy.org;](http://www.cazy.org/) [http://peroxibase.toulouse.inra.fr/\)](http://peroxibase.toulouse.inra.fr/) and (CAZypedia.org)

Supplementary Table 3. Results of one-way ANOVA (1-sided) for the relative abundance of CAZy gene families above and below the basal area increment (BAI) change point. P values are Bonferroni corrected.

Gene ID	SS	MS	statistic	p.value	p.adjust
AA1	1358.285	1358.285	9.837	0.003	0.206
AA10	26.058	26.058	13.862	0.000	0.040
AA11	1221.870	1221.870	13.025	0.001	0.056
AA12	1370.698	1370.698	16.511	0.000	0.015
AA13	62.767	62.767	6.160	0.016	0.557
AA2	245.721	245.721	4.213	0.045	0.720
AA3	33027.744	33027.744	8.549	0.005	0.323
$AA3_1$	591.208	591.208	9.559	0.003	0.217
AA3 2	10085.599	10085.599	11.747	0.001	0.095
AA3 3	61977.388	61977.388	6.435	0.014	0.557
AA3 4	2891.849	2891.849	2.390	0.128	0.987
AA5	1872.023	1872.023	2.996	0.089	0.981
$AA5_1$	1181103.298	1181103.298	9.954	0.003	0.199
$AA5_2$	395.009	395.009	3.917	0.053	0.741
AA6	1375.186	1375.186	6.895	0.011	0.516
AA8	2.039	2.039	2.825	0.099	0.987
AA9	223.837	223.837	13.122	0.001	0.054
CBM1	635.321	635.321	18.337	0.000	0.007
CBM12	134.078	134.078	5.389	0.024	0.602
CBM13	185.945	185.945	5.860	0.019	0.568
CBM18	634.917	634.917	14.707	0.000	0.029
CBM19	551.007	551.007	6.510	0.014	0.557
CBM21	308.888	308.888	2.763	0.102	0.987
CBM43	427.682	427.682	9.829	0.003	0.206
CBM48	111.285	111.285	7.661	0.008	0.440
CBM52	0.931	0.931	0.188	0.666	0.987

Supplementary Table 4: Generalized Dissimilarity Model (GDM) output: Variables contribute to differentiation of fungal community decay (metagenomic) potential for the gene families studied here. Backwards model selection initially procedure initially included all environmental

Supplementary Table 5: Analysis of BAI, parameter posterior means, SD, and 95%CI. In bold: coefficients that are statistically significant (95%CI does not include zero).

Supplementary Table 6: Analysis of GNES results, parameter posterior means, SD, and 95%CI.

Supplementary Table 7: Analysis of GNES slopes parameter posterior means, SD, and 95%CI.

Supplementary Table 8: Site locations in degrees for each of the forest stands depicted in Figure S1.

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