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Identifying monitoring information needs that support the management of fish in large rivers --Manuscript Draft--

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Keywords:	large river; monitoring; long-term; information needs; fish; conceptual model, ecosystem
Abstract:	Management actions intended to benefit fish in large rivers can directly or indirectly affect multiple ecosystem components. Without consideration of the effects of management on non-target ecosystem components, unintended consequences may limit management efficacy. Monitoring can help clarify the effects of management actions, including on non-target ecosystem components, but only if data are collected to characterize key ecosystem processes that could affect the outcome. Scientists from across the U.S. convened to develop a conceptual model that would help identify monitoring information needed to better understand how natural and anthropogenic factors affect large river fishes. We applied the conceptual model to case studies in four large U.S. rivers. The application of the conceptual model indicates the model is flexible and relevant to large rivers in different geographic settings and with different management challenges. By visualizing how natural and anthropogenic drivers directly or indirectly affect cascading ecosystem tiers, our model identified critical information gaps and uncertainties that, if resolved, could inform how to best meet management objectives. Despite large differences in the physical and ecological contexts of the river systems, the case studies also demonstrated substantial commonalities in the data needed to better understand how stressors affect fish in these systems. For example, in most systems information on river discharge and water temperature were needed and available. Conversely, information regarding trophic relationships and the habitat requirements of larval fishes were generally lacking. This result suggests that there may be a common need for a better understanding of certain factors across large-river systems.
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PONE-D-21-12444 Identifying monitoring information needs that support the management of fish in large rivers

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Reviewer 1 wonders about the utility of such conceptual models in the more realistic case of multiple species contexts. How could the model be expanded in this regard, besides the case of simple richness measures? In addition, how could the model help in actually prioritising or ranking the variables or interactions identified? Finally, reviewer 1 also noticed that an important stressor related to fragmentation and connectivity is only marginally discussed and included in the model. This is critical for meta-population dynamics and should be given more emphasis.

Reviewer 2, similarly raises the critical issue of the multi-species context, and how the needs of different species could be simultaneously identified. Therefore, I advice to expand the Discussion in this regard, eventually acknowledging limitations and suggesting future research needs. Reviewer 2 also wonders how such models could effectively guide restoration and decision-making beyond monitoring needs; if the relative importance of data gaps and interactions is not guantified (e.g. via a costbenefit analysis), how could it help prioritise the focus of monitoring and action?

Besides carefully responding to each reviewers' comments and modify the manuscript accordingly, I also suggest to simplify the manuscript, which feels rather long. Perhaps some of the background information from each case study could be included as supplementary or shortened. Also, caption from Fig.5 (and sister-figures) is rather hard to digest for the reader. I wonder if this could be simplified as well.

Response: Thank you for the opportunity to revise our manuscript. We have tried to address comments of Reviewer 1 and 2 below and in the revised manuscript. To shorten and simplify the manuscript, we relocated some of the contextual information from the case studies and moved it to an Appendix in Supplemental Information. With respect to Fig 5 and sister figures, we have discussed trying to simplify the figure captions but have not come up with a good solution. There is a lot of information contained within the figures and feel that further generalizations would not be clarifying. We do, however, acknowledge that the figure caption format is awkwardly long. What we propose is that we retain Fig 5 in the main body of the text as an example, and then move subsequent sister figures to Supplemental Materials. Please let us know if this satisfies your and the reviewer's comments to reduce the length of the manuscript.

Comments to the Author

5. Review Comments to the Author

Reviewer #1: In this study a conceptual model is used to aid the development of best practices of large river monitoring programs. The model was developed based on former scientific works and during scientist's workshop negotiations. Case study applications prove that the application of this complex conceptual model can be useful to identify critical information gaps, which can then be used to develop management and monitoring objectives.

I like the approach of developing such conceptual models, which can reveal information gaps, and think that the model in general can be useful to adopt across large river systems with some refinements and local adaptations. Consequently, I believe showing such an approach can provide useful information for the readers. 1)What I lack is to show more convincingly how such complex models can be used for multispecies systems, where not only the requirements of a single species is evaluated, which in fact the more realistic situation. How can individual species level models be put together to provide meaningful information for management? It would be useful to discuss this in more detail in the Discussion section.

Response: Thank you for this comment. We have added text in the discussion that describes how the CM could contribute to our understanding of the need for a multispecies context in management activities. Please see: L645-691

2)Also a critical issue which should be briefly discussed is how the identified critical target variables should be prioritized, especially in a multispecies systems, where several variables will appear. Development of this section could convince the reader and could clearly show the applicability of such conceptual models by management.

Response: Thank you for this insight. We have provided language in the discussion that describes how the CM could be used to identify critical target variables in a multispecies context. Please see: L650-668

3)Although channel morphology/hydraulics may contain fragmentation/connectivity issues this should be made more clear in the material, because this is one of the most critical issue, which determine fish (meta)population or metacommunity dynamics. In fact fragmentation is often used as one of the most critical variable of anthropogenic drivers and as such is a critically important target to mitigate by management. However, it does not appear either in Fig 2 or Fig 3, but only on the case study figures belonging to morphology/hydraulics TIER1 components.

Response: Thank you. We agree with your assessment. We have further emphasized the importance of fragmentation by specifically mentioning it in the manuscript section describing the CM form. Please see: L223-234. Also, in Table 2, habitat fragmentation is emphasized as affecting multiple facets of the CM in multiple river systems and is mentioned in the text describing Table 2. This result indicates the need to better study the effects of habitat fragmentation on multiple biotic components.

Reviewer #2: Dear Editor,

This study demonstrates how a conceptual model can be used to identifying knowledge gaps in the mechanisms by which Essential Ecosystem Characteristics influence large-river fish species in the USA. These gaps should then be filled to improve the effectiveness of restoration and management.

I agree on the value of these conceptual models to identify knowledge gaps and inform decisions on what to monitor to fill them and allow a better understanding of the system and, therefore, enhance our capacity to manage them adequately. However, I disagree with some of the arguments:

1)The conceptual model represents potential interactions across different structural element of the river system but does not allow quantitative evaluations of strength of those interaction. As such, the value of is conceptual model is limited to identifying knowledge gaps and cannot be used to evaluate the relative importance of each interaction. Therefore, this conceptual model should only be used for identifying knowledge gaps and not for decision-making, as argued (see L771-773), beyond monitoring.

Response: This is an excellent point and we have removed the statement in L771-773 and elaborated on the considerations that need to be accounted for, and the difficulties with, assessing benefit:cost ratios. Please see: L692-726. Also, you are correct that the CM does not provide quantitative evaluations of the strength of the relations. We do acknowledge this and suggest that the CM could provide a basis for developing quantitative assessments in L749-754 and have added language that describes how the CMs could be the basis for Structural Decision Making and Adaptive Management processes (see L727-741).

2)The conceptual model lacks a cost analysis to evaluate the most efficient way of filling knowledge gaps. Some of the gaps might be more difficult/ costly or even feasible to fill. Without such analysis we can only identify the gaps but cannot prioritise where to focus monitoring on a cost-effective way and just confirm where gaps exist.

Response: Thank you for this comment. In addition to addressing the benefit:cost issue above, we have provided language in the discussion that describes how the CM could be used to identify critical target variables in a multispecies context but that there are critical uncertainties that need to be considered. See: L650-668. We agree with you about prioritizing based on cost effectiveness but respectfully suggest that the CM could provide information that would suggest where to focus monitoring effort.

	 3)Three of the case studies present conceptual models for individual species. While I see the value of developing these conceptual models for charismatic endangered species, I wonder how feasible/ useful it would be this method when facing management needs for many species simultaneously. One of the case studies does present a conceptual model for the full fish community, but focused on diversity, rather than individual species, so no information of particular species issues are addressed. Would it be feasible to elaborate a conceptual model that addressed all individual species needs/ issues simultaneously? This would allow identifying knowledge gaps common to multiple species simultaneously. Response: Thank you for this comment. We have added text in the discussion that describes how the CM could contribute to our understanding of the need for a multispecies context in management activities. Please see: Please see: L645-691 4)Minor comments: L304 & 369. What does ATV stand for? Response: Thank you for bringing this to our attention. ATV stands for all-terrain vehicle. We have removed the acronym from the revision. The manuscript is quite long, especially because of the description of each case study. It would be good to present the information of these case studies in a more synthetic way (maybe on a table?). Response: Thank you for the comment. Per your and the Associate Editor's recommendation we have pulled out some of the contextual information from the case studies and moved the information to an Appendix in Supplemental Information. We have also moved three figures and associated captions to the supplemental information.
Additional Information:	
Question	Response
Financial Disclosure Enter a financial disclosure statement that describes the sources of funding for the work included in this submission. Review the <u>submission guidelines</u> for detailed requirements. View published research articles from <u>PLOS ONE</u> for specific examples. This statement is required for submission and will appear in the published article if the submission is accepted. Please make sure it is accurate.	This work was funded in part by U.S. Geological Survey's Core Science Systems Mission Area. This research also was conducted using in-kind contributions of the Ball State University, Illinois Natural History Survey, Oregon Department of Fish and Wildlife, Pacific Northwest Aquatic Monitoring Partnership, the Oklahoma and Alabama Cooperative Fish and Wildlife Research Units, and the U.S. Geological Survey Columbia Environmental Research Center, Grand Canyon Monitoring and Research Center, Oregon Water Science Center, Upper Midwest Environmental Sciences Center, and Western Fisheries Research Center.

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- Include the approval number and/or a statement indicating approval of this research
- Indicate the form of consent obtained (written/oral) or the reason that consent was not obtained (e.g. the data were analyzed anonymously)

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- Provide the name of the Institutional Animal Care and Use Committee (IACUC) or other relevant ethics board that reviewed the study protocol, and indicate whether they approved this research or granted a formal waiver of ethical approval
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et
Additional data availability information:

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30 **Abstract** – Management actions intended to benefit fish in large rivers can directly or indirectly affect multiple ecosystem components. Without consideration of the effects of management on 31 non-target ecosystem components, unintended consequences may limit management efficacy. 32 33 Monitoring can help clarify the effects of management actions, including on non-target 34 ecosystem components, but only if data are collected to characterize key ecosystem processes 35 that could affect the outcome. Scientists from across the U.S. convened to develop a conceptual model that would help identify monitoring information needed to better understand how natural 36 and anthropogenic factors affect large river fishes. We applied the conceptual model to case 37 38 studies in four large U.S. rivers. The application of the conceptual model indicates the model is flexible and relevant to large rivers in different geographic settings and with different 39 management challenges. By visualizing how natural and anthropogenic drivers directly or 40 indirectly affect cascading ecosystem tiers, our model identified critical information gaps and 41 uncertainties that, if resolved, could inform how to best meet management objectives. Despite 42 43 large differences in the physical and ecological contexts of the river systems, the case studies also demonstrated substantial commonalities in the data needed to better understand how 44 stressors affect fish in these systems. For example, in most systems information on river 45 46 discharge and water temperature were needed and available. Conversely, information regarding trophic relationships and the habitat requirements of larval fishes were generally lacking. This 47 48 result suggests that there may be a common need for a better understanding of certain factors 49 across large-river systems.

50

51 Introduction

52 Long-term monitoring has benefited a variety of marine and freshwater ecosystems, including large rivers like the Ohio [1, 2] and Illinois rivers [3]. Large-river systems are 53 complex, making the development of effective monitoring programs especially difficult. Large 54 rivers are dynamic systems with high variability in spatio-temporal physicochemical 55 56 characteristics and biotic assemblages [4]. The inherent complexity of large rivers makes biotic 57 assemblages logistically difficult to sample [5] and the mechanisms of change difficult to understand. Large rivers represent the culmination of vast stream networks and, thus, integrate 58 and accumulate the effects of multiple stressors at varying spatial scales [6]. The spatial and 59 60 temporal complexity associated with large rivers has hindered the identification of mechanisms driving declining populations of aquatic species [7-10]. To exacerbate the complexity, large 61 62 rivers commonly have within-channel structural alterations (e.g., dams, river training structures, 63 [11]) and often exhibit legacy effects from historical land uses [12]. To deal with the complexity, some areas of aquatic science recommend monitoring be used to test the linkages developed first 64 through conceptual models (e.g., environmental flows, [13-15]). 65

Conceptual models are useful tools to help guide the design of monitoring programs [16].
The identification of questions relevant to conservation and management efforts requires some
foresight and knowledge of the complexities of the system being monitored. For example, it is
generally well accepted that the native range of the federally-listed Arkansas River Shiner
(*Notropis girardi*) is truncated [17], though there is uncertainty surrounding the multiple threats
affecting the species [18]. Reducing the uncertainty associated with the decline of the Arkansas
River Shiner through the implementation of a hypothesis-driven monitoring program would

facilitate confidence in moving forward with a recovery plan. This is where conceptual models
are quite useful; they can serve as the foundation to guide hypothesis-driven monitoring
programs [14, 16] and identify key ecosystem processes and factors that may directly or
indirectly affect management outcomes [19-22].

Understanding factors affecting the status and trends of fishes is of interest to multiple 77 78 stakeholder groups across multiple jurisdictions. Fishes provide economic benefits to businesses 79 that serve recreational interests, commercial and recreational fishers, tribal members for whom fish are an integral part of their cultural identity [23], and to local and state governments who 80 81 derive revenue from these activities. Fish populations are affected by the integration of physical habitat, water quality, environmental contamination, habitat fragmentation, and overall 82 ecosystem productivity [24-27]. Consequently, fish are often the focus of management and 83 monitoring programs (e.g., [28]). However, because fish integrate the effects of so many 84 components of the ecosystem, the success of efforts to manage fishes can be affected by 85 unintended consequences of mitigation on factors not directly targeted by the actions. Without 86 consideration of the effects of management on non-target ecosystem components, unintended 87 consequences may limit management efficacy. 88

Our goal is to demonstrate how a structured, yet flexible, conceptual model (CM) can be used to identify the types of monitoring information needed to understand the range of factors affecting large-river fishes. Our CM includes a hierarchically structured conceptualization of ecosystem characteristics based on CMs originally developed by Harwell, Myers (29) and elaborated by Jacobson and Berkley (30). We chose to incorporate the tiered conceptualization of ecosystem characteristics proposed by Jacobson and Berkley (30) in part because it allows users to define their own biotic or abiotic interests. In this paper, we discuss the structure and

development of the CM. We apply the CM to case studies to illustrate the flexibility and
applicability of this approach and use it to identify monitoring information needs specific to
disparate management goals. More specifically, for each case study, we use the CM to
hypothesize how human activities affect fish populations and then identify information needs
required to evaluate the hypothesized relationships. We then posit the spatial and temporal scales
of the management goal addressed in the conceptual model, inferences needed to inform the
management goal, and data collection requirements needed to make the inferences.

103 Conceptual Model

104 **Overview of Approach**

Since 2012, scientists working on large rivers across the United States have participated 105 in a forum intended to improve our understanding of large-river ecosystems. The collaborative 106 forum has worked to identify best practices of long-term monitoring programs [31] and evaluate 107 108 trends in fish assemblages across rivers [32]. As this group of scientists moved toward linking 109 changes in fish populations and assemblages to human activities, there was a need to develop a process to help identify and prioritize the information needed to assess trends in large river 110 fishes. To that end, a workshop was convened in Hood River, Oregon in May 2017, to jointly 111 112 adapt, apply, and qualitatively evaluate a conceptual model for developing hypotheses that detail stressors affecting fishes arising from natural and anthropogenic sources [33]. 113

Our general approach was to first identify human activities that affect large-river fishes and then hypothesize how the activities related to physical and chemical factors and biological communities. Prior to the workshop, we elicited opinion from experts that comprise a U.S. Geological Survey (USGS) led forum on large river monitoring about important anthropogenic activities that could affect fish populations or communities in the river basins they work in. (Fig 1). We summarized the information from this exercise and grouped the anthropogenic activities into driver categories (Fig 2) and proposed a general form of the CM. We then disseminated the information to the experts prior to the workshop.

Fig 1. Map of rivers and watersheds represented by scientists that convened to develop a conceptual model that depicts how natural and anthropogenic drivers interact with habitats, biological systems, and fish in large rivers. River segments where we conducted case studies that applied the conceptual model to identify monitoring information needs associated with management goals are highlighted in red.

Fig 2. The results of a query to scientists from the Canadian, Colorado, Columbia, Hudson,
Illinois, Ohio, Missouri, Red, Tallapoosa, Upper Mississippi, and Wabash Rivers, U.S to identify
anthropogenic activities that affect large-river fishes in the river systems they represent.
Anthropogenic activities were classified into five driver categories.

During the workshop we discussed and refined the CM form (Fig 3). We then had representatives from each river system represented at the workshop choose a management goal to address. Then, through a facilitated discussion led primarily by the representative of the river system being addressed, we 1) elaborated tiered conceptualizations of ecosystem characteristics to reflect the large-river systems and management goals being examined, 2) used knowledge of the fish species' life history and population bottlenecks to relate biological ecosystem characteristics to habitat requirements, 3) hypothesized pathways describing how anthropogenic and natural drivers affect large-river fish populations either indirectly (e.g., effects on flow
regime, habitat, trophic resources, etc.) or directly (e.g., competition with invasive species), and
4) hypothesized interactions within ecosystem characteristic tiers that could affect the
management goal. Based on this exercise, we chose four case studies to refine for use in this
manuscript (Fig 1).

143 Fig 3. Tiered hierarchical conceptualization of how anthropogenic and natural drivers relate to 144 physical and biological components of large-river ecosystems. Essential ecosystem characteristics (EECs) are groupings of ecosystem components. Tier 1 EECs represent physical 145 and chemical effects; fundamental measures of process that are directly affected by 146 147 anthropogenic and natural drivers. Tier 2 EECs represent a broad habitat category that is intended to encompass the physical, chemical, and biological components of the riverine habitats 148 149 that influence reproduction, growth, and survival of biotic communities. The Tier 3 EEC 150 represent components of the biological systems that respond to changes in the hierarchical 151 components of the conceptual model.

After the workshop, we held a series of conference calls with workshop participants with expertise in the selected case studies to refine all aspects of the CMs and associated information. During the calls, we started with the CM from the workshop and discussed and clarified CM components, pathways, and inter-tier interactions. We then characterized whether, based on the expert knowledge of workshop participants, there was a strong, moderate, or weak understanding of the pathways and interactions. A list of the information needed to understand the relationships described in the CM was developed. We then had the case study leader classify whether the data required to understand the information needs were available, insufficient, or not available. For information needs that were classified as insufficient or not available, we characterized the spatial and temporal scales at which data should be collected to make inferences that support the evaluation of management goals.

We then encouraged representatives from each basin to share the CM and the case study narrative with other experts familiar with the river system and management goal. This outreach took several forms including sharing the CM with working groups tasked with implementing the management goal, discussions with peers familiar with the management goal, and presenting the CM at regional conferences. The intent was to garner opinions from outside the workshop participants. If needed, the CMs incorporated the feedback received,

169 Hierarchical structure of Conceptual Model

Our CM is a hierarchical conceptualization of how anthropogenic and natural drivers 170 171 relate to multiple tiers representing the physical and biological components of large-river ecosystems (Fig 3). Natural drivers included in the CM were physiographic, climatic, and 172 biogeographic factors that control fluxes of water, mass, energy, and genetic information in a 173 174 watershed [30]. The physiographic factors, such as lithology, soils, and watershed topography, 175 exert control on water, sediment, and geochemical fluxes (e.g., nutrients) into the river corridor. 176 Physiography is generally static over time frames of decades to centuries. Climate controls fluxes 177 of atmospheric energy and moisture into the watershed. Unlike physiography, climate is more 178 likely to vary over relatively shorter temporal scales. Biogeography describes the native 179 organism assemblage in the watershed (e.g., [34]) and the natural flux of genetic information due

to immigrations, emigrations, mutations, and extinctions. Changes arising from the biogeography
driver includes altered spatial distribution of organisms within the watershed, which, in turn, may
alter the effects of natural system regimes on the river corridor. For example, natural variation of
the type and distribution of vegetation can affect the time series and magnitude of runoff events
[30].

We created five categories of anthropogenic drivers to characterize a range of human 185 activities that affect large rivers: land use, commercial use, biological community alteration, 186 water use, and recreation (Fig 2). The land-use category is intended to reflect different ways 187 188 humans use landscape resources that affect large rivers. We defined commercial use as the use of 189 river resources for marketable enterprises that did not involve water removal or transfer. We included biological community alteration to represent the intentional or non-intentional human 190 191 alteration or manipulation of the river's biological community (e.g., introductions of non-native fish). Recreational use was defined as the use of river resources for leisure activities (e.g., 192 fishing, boating). We considered water use a direct commercial or non-commercial use of river 193 194 water that involved the removal or transfer of water.

Our CM includes hierarchically structured essential ecosystem characteristics (EEC), 195 196 originally developed by Harwell, Myers (29) and described in detail by Jacobson and Berkley (30). Briefly, EECs are characteristics that can be classified into similar groups based on the way 197 they link to biological endpoints [30]. Tier 1 EECs are measurable characteristics that describe 198 199 processes that can significantly alter the morphological or chemical characteristics within a river channel. The Tier 1 categories we considered were 1) hydrology, 2) channel 200 morphology/hydraulics, 3) sediment transport, and 4) biogeochemistry/thermodynamics. Tier 2 201 202 EECs are broadly described as physicochemical or biological components of "habitat" that are

203 hypothesized to affect (e.g., growth, survival, reproduction, [35]) fish populations or 204 assemblages. Lastly, Tier 3 EECs represent components of the hypothesized biological system that are affected by the cascading (e.g., degradation of egg quality caused by increases in 205 206 sediment deposition) or direct (e.g., predation by invasive species) effects of anthropogenic and natural drivers. Tier 2 characteristics are particularly important because these are the factors that 207 208 can be examined at scales most often sampled by fisheries managers [36]. The specific components that comprise Tier 2 and 3 EECs are flexible and can be adapted and elaborated 209 depending on the river system and specific management goal being addressed. 210 211 We retained aspects of the approach taken by Jacobson and Berkley (30) with respect to how our model represents interactions between drivers and EECs, but with key differences. 212 Since we were interested in representing how human activities affect large river ecosystems, our 213 approach acknowledges that anthropogenic and natural drivers interact and alter the expected 214 characteristics of Tier 1 EECs. Similar to Jacobson and Berkley (30), our model depicts a stress 215 associated with a natural or anthropogenic driver to Tier 1 EECs as fluxes in natural system 216 217 regimes that alter the frequency, magnitude, duration, timing, or rate of change in natural systems or by the imposition of a hard-structural constraint on channel form. The natural system 218 219 regimes considered in our CM were hydraulic, hydrologic, sediment, temperature, light, and biogeochemistry. Graphically, the natural system fluxes were represented by arrows connecting 220 anthropogenic and natural drivers to Tier 1 EECs. Similarly, hypothesized pathways between 221 222 EECs, that depict the expression of the cascading effects of anthropogenic and natural drivers, and interaction within EECs were depicted as arrows. For example, fragmentation of river 223 224 systems resulting from altered hydrologic and/or hydraulic regimes caused by dams, weirs, 225 levees, and other factors are frequently cited sources of stress to large-river fishes [37].

226 Fragmentation can prevent fish from migrating and/or dispersing through their natural 227 reproductive ranges and from accessing critical habitats [38]. To depict a scenario where the presence of a dam is altering hydrologic and/or hydraulic regimes resulting in habitat 228 229 fragmentation, the CM would show an arrow from an anthropogenic stress (i.e., dam as a commercial activity) to a Tier 1 EEC (e.g., channel morphology/hydraulics) that would depict a 230 natural system flux (e.g., altered hydrologic and/or hydraulic regime) that would then manifest as 231 a stress caused by habitat fragmentation depicted by an arrow between the Tier 1 EEC and a Tier 232 2 EEC (e.g. habitat) that would then manifest as an effect on a Tier 3 component, shown by an 233 234 arrow between Tier 2 and Tier 3. All stress pathways and interactions were classified with respect to the strength of understanding of the relationships based on expert opinion. Arrows 235 with solid blue lines depict a strong understanding of the relationship, dotted-dashed blue lines 236 represent a moderate understanding of the relationship, and with a black dashed line represent a 237 weak understanding of the relationship. 238

239

Spatial and temporal context

The successful characterization of how human activities influence large-river fishes is dependent upon integrated concepts of scale. Fish distributions in rivers can vary spatially within river basins in relation to naturally occurring and human induced landscape characteristics [32]. Fish distributions can also vary seasonally, annually, and over longer times in response to changing environmental conditions [39]. Consequently, the spatial and temporal scope of fish management goals often varies within and between large-river systems and agencies. For data collected by monitoring programs to have the highest relevance, the spatial and temporal scales 247 appropriate for scientific investigation and management must also be time and geographiccontext specific. For example, the management of White Sturgeon (Acipenser transmontanus) in 248 the Columbia River varies by reservoir or river segment and season [40, 41]. The effects of 249 250 hydropower development on White Sturgeon vary spatially and temporally as well, so the spatial and temporal context of the data needed to understand the effects needs to be considered. For 251 252 instance, hydropower peaking operations, that can vary by dam and season, affect river discharge in a river reach on a diel and even hourly basis [42], whereas water storage and other 253 management actions can affect seasonal discharges over a broader geographic scale [40]. 254 255 Understanding the spatial and temporal context needed to inform management will help ensure relevant information is collected. 256

We considered spatial and temporal resolution in our CMs. We defined spatial extent as : 257 local network – synonymous with Hydrologic Unit Code (HUC) 2 basins [43, 44]; segment – the 258 portion of a river between two major tributary confluences [45] or other hydrogeomorphic 259 features [46]; reach - the length of river occurring between breaks in channel slope caused by 260 261 man-made dams or other hydrogeomorphic features [45]; patch – an area used by an organism (e.g., for reproduction or resource attainment) that can vary both spatially and temporally 262 263 depending on the species of interest [47, 48]. For our purposes, the spatial scales considered are nested such that segments occur within local networks, reaches occur within segments, and 264 patches occur at the sub-reach scale. Temporal units considered were daily, seasonal, annual, and 265 266 decadal. The temporal units were used to denote both the scale of inferences needed to support the management goal and the scale at which data should be collected to inform the inferences. 267

268 Case Studies

269 We applied the CM to four case studies. For each, we followed the pathways of stress 270 from Tier 1 EECs to the biological endpoint associated with the management goal to identify information needs. We then characterized the spatial and temporal scales of the management 271 272 goal, the scientific inferences needed to inform the management goal, and that data collection needs to occur to support the inferences for monitoring information needs identified as requiring 273 274 additional data for each of the case studies. To summarize similarities across case studies, we generalized the stressors and inter- tier interactions identified in the case studies and then 275 summarize the similarities by EEC tier. More context for the river systems characterized in the 276 277 case studies can be found in Appendix S1.

278 South Canadian River

Native populations of the federally-threatened Arkansas River Shiner are believed to be 279 restricted to two fragmented portions of the South Canadian River [49]. The Arkansas River 280 281 Shiner is hypothesized to be affected by several anthropogenic activities that primarily affect water quality and quantity (Fig 4). Three reservoirs on the South Canadian River have altered 282 discharge patterns (Fig S1), and fragmented river habitats. Two known native populations of 283 284 Arkansas River Shiner occupy the two remaining river segments of sufficient length and complexity to allow eggs to drift the time required to successfully complete their early life 285 286 history. Small impoundments for agriculture use, road crossings, groundwater pumping and other 287 local water extractions (e.g., oil and gas) threaten to further fragment existing habitat. 288 Fragmentation could also be problematic for upstream fish migrations; there is some evidence 289 that Arkansas River Shiners migrate upstream to spawn to achieve adequate drift distances for

290 their offspring [50]. It has also been speculated that this species might benefit from access to 291 floodplain habitats [51], but we are unaware of efforts to examine that hypothesis. Changes in the flow patterns may also relate to the expansion of salt cedar *Tamarix* spp. and other non-292 293 native riparian species that constrain the channel and inhibit channel habitat complexity [52, 53]. Changes to the riparian corridor can also alter the availability of drifting invertebrates for 294 Arkansas River Shiner feeding (i.e., Coleoptera, Hymenoptera; [54]). Channel complexity acts to 295 slow the transport of eggs [49] and may prevent eggs from being washed into downstream 296 reservoirs where survival is hypothesized to be extremely low. Climate change is expected to 297 298 increase the intensity and frequency of drought events within this region [55, 56], which may exacerbate habitat fragmentation, promote all-terrain vehicle traffic within the river channel 299 causing direct mortality on stranded fish (Gene Wilde, Texas Tech University, Personal Comm.), 300 and concentrate contaminants and salinity [57]. The tolerances of Arkansas River Shiner to 301 salinity concentrations and many other contaminants are unknown (see Table S1; [18]). Lastly, 302 introductions of non-native fishes have occurred within the basin. The primary concern is the 303 304 presence of Red River Shiner (Notropis bairdi) because it is suspected to reproduce in a similar manner and be a possible competitor to the Arkansas River Shiner [57]. 305

Fig 4. Conceptual model describing the relationship of natural and anthropogenic drivers to
essential ecosystem characteristics (EECs) affecting the recruitment of the Arkansas River
Shiner in the South Canadian River in New Mexico, Texas, and Oklahoma. Essential ecosystem
characteristics are groupings of ecosystem components. Tier 1 EECs represent physical and
chemical effects; fundamental measures of process that are directly affected by anthropogenic
and natural drivers. Tier 2 EECs represent a broad habitat category that is intended to encompass

312 the physical, chemical, and biological components of the riverine habitats that influence reproduction, growth, and survival of biotic communities. The Tier 3 EECs represent 313 components of the biological systems that respond to changes in the hierarchical components of 314 the conceptual model. The strength of our understanding of how natural and anthropogenic 315 drivers interact with habitats, biological systems, and fish in large rivers is represented by the 316 317 different types of lines in the figure. Solid blue lines depict a strong understanding of the relationship, the dotted-dashed blue line represents a moderate understanding of the relationship, 318 and the black dashed line represents a weak understanding of the relationship. The different 319 320 types of lines also represent the strength of our understanding of within EEC-tier relationships.

321 The results of the CM exercise that characterized factors affecting the Arkansas River 322 Shiner in the South Canadian River suggested the critical life-history bottlenecks for the 323 Arkansas River Shiner are successful spawning and recruitment to the first year. Impediments 324 that limit our understanding of factors that lead to successful spawning and recruitment included the effects of channel morphology and hydraulics on the quality and quantity of larval rearing 325 habitat, and subsequent effects on larval production (Table S1). Water use and other drivers 326 occurring at relatively coarse spatial and temporal scales are the hypothesized drivers related to 327 degradation of reproductive habitat for the Arkansas River Shiner (Fig 5). A temporal lag in 328 329 responses at finer scales (i.e., improved habitat) would be anticipated with management actions at these coarser spatial scales (e.g., water releases from dams); though, providing connectivity 330 via minimal water releases would occur relatively quickly. Although there are gages on the 331 332 South Canadian River, the spacing of the gages is not sufficient to have a full understanding of flow patterns between the gages given the semi-arid nature of the basin and potential for reaches 333

to be affected by water withdrawals such as groundwater pumping. Our understanding of the
species life history is well established; however, the effects of human pressures on the species
and associated habitat has not been well studied (i.e., production, survival). As stressors
propagate through Tier 1 to Tier 2 and Tier 3, the level of uncertainty increased such that it is not
possible to define a preferred hypothesis for Arkansas River Shiner recruitment failure. The
status of information needed to understand the hypothesized stress pathways and interactions was
mostly characterized as insufficient or not available (Table S1).

Fig 5. The spatial and temporal scales of the management goal, the scientific inferences needed 341 342 to inform the management goal, and that data collection needs to occur to support the inferences 343 for monitoring information needs identified as requiring additional data in the case study addressing the recruitment of the Arkansas River Shiner in the Canadian River, Oklahoma (see 344 345 Table S1 for additional detail). A:Tier 1 EEC=channel morphology/hydraulics; Stressor=altered 346 hydraulic regime; B:Tier 1 EEC=biogeochemistry/thermodynamics; Stressor=altered water temperature regime; C:Tier 1 EEC=biogeochemistry/thermodynamics; Stressor=altered 347 biogeochemical regime; D:Tier 2 EEC=Arkansas River Shiner spawning habitat; 348 Stressors=contaminants, water temperature, habitat fragmentation; E:Tier 2 EEC=larval 349 Arkansas River Shiner habitat, Stressors=water temperature, habitat fragmentation and Tier 2 350 351 EEC=invertebrate habitat, Stressors=altered riparian plant community, discharge, sediment deposition; F:Tier 3 EEC=primary production, Stressor = nutrient flux; G:Tier 3 352 EEC=invertebrate production; Stressor=invertebrate habitat quantity and quality and Tier 3 353 354 EEC=Arkansas River Shiner larvae production, Stressor=predation by invasive species; H:Tier 1 EEC= biogeochemistry/thermodynamics; Inter-tier interaction= Sediment adsorption of 355

356 contaminants and nutrients and Tier 3 EEC=Arkansas River Shiner larvae production, Stressors=Arkansas River Shiner larvae habitat quantity and quality and Tier 3 EEC=Arkansas 357 River Shiner age-0 recruitment; Inter-tier interaction=Arkansas River Shiner larvae mortality and 358 Tier 3 EEC=Arkansas River Shiner age-1+ recruitment, Inter-tier interaction= Arkansas River 359 Shiner age-0 mortality and Tier 3 EEC=all, Inter-tier interaction=trophic level interactions; 360 361 I:Arkansas River Shiner larvae production; Stressor= direct mortality from recreational use (i.e., all-terrain vehicle and in-river traffic); J:Tier 3 EEC=Arkansas River Shiner egg quality and 362 production; Stressor=Arkansas River Shiner spawning habitat quantity and quality; K:Tier 3 363 364 EEC=Arkansas River Shiner larvae production; Inter-tier interaction=Arkansas River Shiner egg mortality. 365

366 Colorado River

The Humpback Chub (Gila cypha), a fish native to the Colorado River, was listed as 367 368 endangered by the U.S. Fish and Wildlife Service in 1967 and given full protection under the Endangered Species Act of 1973 (ESA). To mitigate the effects of anthropogenic changes to the 369 river on Humpback Chub, an understanding of the mechanisms by which Glen Canyon Dam and 370 371 non-native species affect Humpback Chub is needed. A critical life-history bottleneck for Humpback Chub is recruitment into the first year class (Fig 6). Temperature, light, and seasonal 372 373 high river discharge from snowmelt are thought to cue spawning behavior [58]. Hydropower 374 development has dampened the range of river discharges of the Lower Colorado River within Grand Canyon. Historically, river discharge varied between 15 and 3400 m³/s, however 375 376 discharge was greater in large flood events; current dam operations limit flows to a range of 140

377 to 1000 m^3/s (Fig S2). Resulting changes in turbidity and water temperature create risks to endangered Humpback Chub, and other endemic fish. For example, the quantity and quality of 378 habitat is reduced through changes in turbidity, biogeochemistry, and the temperature regime. 379 Hypolimnetic water releases from Lake Powell maintain cold temperatures in the Colorado River 380 downstream from Glen Canyon Dam; currently, spawning is limited to a single tributary, the 381 382 Little Colorado River. As embryos survive into the larval stage, nursery habitats to support growth and foraging are essential [59-61]. A secondary risk to juvenile survival post-larval stage 383 is predation by non-native species including Rainbow Trout (Oncorhynchus mykiss) and Channel 384 385 Catfish (Ictalurus punctatus) [62]. Temperatures for well over 100 km downstream of Lake Powell are excellent for non-native, cold water species, including a closely managed world-class 386 Rainbow Trout fishery at Lees Ferry. Rainbow and Brown Trout (Salma trutta) are currently 387 managed as an invasive species downstream of the confluence of the Colorado and Little 388 Colorado River, approximately 97 km downstream of Glen Canyon Dam, to mitigate predation 389 390 upon native fishes including the endangered Humpback Chub. Fig 6. Conceptual model describing the relationship of anthropogenic drivers to essential 391

ecosystem characteristics (EECs) affecting the recruitment of Humpback Chub in the Colorado River between Glen Canyon Dam and Lake Mead, Arizona. Essential ecosystem characteristics (EECs) are groupings of ecosystem components. Tier 1 EECs represent physical and chemical effects; fundamental measures of process that are directly affected by anthropogenic and natural drivers. Tier 2 EECs represent a broad habitat category that is intended to encompass the physical, chemical, and biological components of the riverine habitats that influence reproduction, growth, and survival of biotic communities. The Tier 3 EECs represent components of the biological systems that respond to changes in the hierarchical components of the conceptual model. The strength of our understanding of how natural and anthropogenic drivers interact with habitats, biological systems, and fish in large rivers is represented by the different types of lines in the figure. Solid blue lines depict a strong understanding of the relationship, the dotted-dashed blue line represents a moderate understanding of the relationship, and the black dashed line represents a weak understanding of the relationship. The different types of lines also represent the strength of our understanding of within EEC-tier relationships.

The CM exercise documented a high understanding of the relationships of anthropogenic 406 drivers to Tier 1 EECs, cascading to multiple, less-understood hypotheses about how these 407 408 factors would combine to affect habitats at Tier 2 (Fig 6). High confidence in the linkages from 409 Tier 2 invertebrate habitat to Tier 3 insect production is followed by a lesser understanding of 410 how insect production is linked to larval production. The uncertainty of the relations between 411 food resources stands in contrast to high certainty that was ascribed to the linkages from larval chub habitat to larval chub production, and from spawning habitat to larval chub production and 412 then to recruitment. Since 1997, the Glen Canyon Dam Adaptive Management Program has 413 supported extensive monitoring and research across the spatial and temporal landscape of the 414 Colorado River. As a result, the information needed to characterize some of the stressors is 415 readily available (Table S2). However, the status of some existing information was characterized 416 as insufficient or not available. For the information needs characterized as being insufficient or 417 not available, we identified the spatial and temporal scales at which data collection would 418 419 facilitate the inferences needed to inform the management goal (Fig S3). Understanding how

these knowledge gaps affect Humpback Chub recruitment could identify strategies that will helpachieve the management goal of improving Humpback Chub recruitment in the Colorado River.

422 Columbia River

White Sturgeon is the largest freshwater fish in North America [63]. Like other sturgeon 423 424 species, anthropogenic stressors have negatively affected White Sturgeon productivity. Our knowledge of factors affecting White Sturgeon productivity remain poorly understood [40]. 425 426 Therefore, we used the CM to identify knowledge gaps associated with the hypotheses that dam 427 construction and operation, land-use practices, and invasive species, in some combination, affect the recruitment of age-0 White Sturgeon (Fig 7). Within the basin, development of hydroelectric 428 429 and water-storage dams have changed the magnitude and seasonality of the natural river discharge (Fig S4) and thermal regimes [64], reduced the quantity and quality of spawning 430 habitats [65, 66], and disrupted historical migration patterns [67]. Prior to hydropower 431 432 development, White Sturgeon experienced a hydrograph that peaked during June-July due to snowmelt [64]. However, from 1949 to 1993 the average discharge in June decreased from 433 14,000 m³/s to 6,000 m³/s and the maximum water temperature has increased by 1.8°C [64]. 434 435 White Sturgeon likely used the natural hydrograph and thermal regime as cues to seek out optimal spawning habitats and initiate spawning [65, 68]. 436

Fig 7. Conceptual model describing the relationship of anthropogenic drivers to essential
ecosystem characteristics (EECs) affecting the recruitment of White Sturgeon in the Columbia
River, U.S. Essential ecosystem characteristics (EECs) are groupings of ecosystem components.
Tier 1 EECs represent physical and chemical effects; fundamental measures of process that are

441 directly affected by anthropogenic and natural drivers. Tier 2 EECs represent a broad habitat category that is intended to encompass the physical, chemical, and biological components of the 442 riverine habitats that influence reproduction, growth, and survival of biotic communities. The 443 444 Tier 3 EECs represent components of the biological systems that respond to changes in the hierarchical components of the conceptual model. The strength of our understanding of how 445 446 natural and anthropogenic drivers interact with habitats, biological systems, and fish in large rivers is represented by the different types of lines in the figure. Solid blue lines depict a strong 447 understanding of the relationship, the dotted-dashed blue line represents a moderate 448 449 understanding of the relationship, and the black dashed line represents a weak understanding of the relationship. The different types of lines also represent the strength of our understanding of 450 451 within EEC-tier relationships.

452 Factors other than river discharge and water temperature may also be affecting age-0 453 White Sturgeon recruitment [40]. In areas of the Columbia and Snake Rivers with hydropower development, White Sturgeon populations are functionally isolated by dams. Consequently, 454 455 White Sturgeon depend on conditions within restricted reaches to sustain production. In some 456 reaches, suitable rearing habitat exists, and individual growth rates are high, but spawning habitat is limited and recruitment of fish is poor [65]. In other reaches, favorable spawning conditions 457 exist but growth of young fish may be density limited [69]. How the availability of food 458 resources for larval and juvenile White Sturgeon varies among reservoirs may affect age-0 White 459 Sturgeon recruitment. Research has also suggested that contaminants may affect White Sturgeon 460 461 reproductive biology [70]. The introduction of non-native fishes has clearly affected the native fish assemblage in the Columbia River [32, 71]. Channel Catfish, Smallmouth Bass *Micropterus* 462

dolomieu, and Walleye (*Sander vitreus*) that have been introduced into the Columbia River have
all been shown to prey upon or compete with native fish species [71, 72] and may also affect
White Sturgeon.

The CM (Fig 7) provides structure to the multiple competing hypotheses and indicates 466 how anthropogenic drivers may be affecting Tier 1, 2, and 3 EEC's. Not surprisingly, the 467 468 strength of our understanding of the effects of stressors on White Sturgeon was greater for relationships between life stages that are more easily sampled (e.g., adults) and that rely on data 469 that are readily accessible as part of monitoring associated with hydropower development (e.g., 470 471 discharge and water temperature) or metrics that are from combinations of these variables (e.g., 472 estimates of White Sturgeon spawning habitat; Table S3). However, for relationships between harder to sample White Sturgeon life stages (e.g., larvae), biota that require expertise and 473 equipment atypical of traditional fisheries assessments in large rivers (e.g., benthic 474 macroinvertebrates), or stressors that are described by metrics that require specialized modeling 475 expertise (e.g., sediment transport dynamics), the existing information was insufficient or not 476 477 available. For example, we identified the need to better understand the effects of channel morphology and hydraulics on benthic macroinvertebrate habitat, invertebrate production, and 478 479 subsequent larval White Sturgeon production. The most certain pathways connected changes in hydrology, hydraulics, and temperature regimes to reduced spawning habitat in Tier 2, then to 480 decreased egg quality and production at Tier 3. 481 482 Our results suggest there are stressors that can affect the management goal of increasing

age-0 White Sturgeon recruitment that are poorly understood and that could confound efforts to
manage White Sturgeon in the Columbia River. Our characterization of the spatial and temporal
scales that data should be collected at could help guide future efforts to fill data gaps to support

the inferences needed to address the goal of improving recruitment of age-0 White Sturgeon (FigS5).

488

Upper Mississippi and Illinois Rivers

In the agricultural Midwest, basin-wide land uses affect the delivery of sediments, 489 490 nutrients, and runoff to the Upper Mississippi and Illinois rivers [73-76]. Within the floodplain of these two large rivers, agriculture and residential land uses often rely upon the use of levees to 491 492 isolate productive or developed lands during seasonal high-flow events. Within the channel, 493 these rivers support commercial navigation with locks and dams and river-training structures, which have dramatically altered channel morphology and hydraulics throughout the system. 494 Together, the cumulative effects of these modifications to the basin, floodplain, and river have 495 implications for habitat diversity and native fish biodiversity [77]. Additionally, recent invasion 496 and expansion of non-native species, namely Silver Carp (Hypophthalmichthys molitrix) and 497 498 Bighead Carp (H. nobilis), have direct and indirect effects on native fishes that likely compound or confound stress pathways on native fish biodiversity [78, 79]. Therefore, we used the CM to 499 explore how these primary anthropogenic drivers have likely influenced fish habitats and 500 501 associated life stages (Fig 8).

Fig 8. Conceptual model of how anthropogenic drivers in the upper Mississippi and Illinois
Rivers influence native fish habitats and recruitment. Essential ecosystem characteristics (EECs)
are groupings of ecosystem components. Tier 1 EECs represent physical and chemical effects;
fundamental measures of process that are directly affected by anthropogenic and natural drivers.
Tier 2 EECs represent a broad habitat category that is intended to encompass the physical,

507 chemical, and biological components of the riverine habitats that influence reproduction, growth, 508 and survival of biotic communities. The Tier 3 EECs represent components of the biological systems that respond to changes in the hierarchical components of the conceptual model. The 509 510 strength of our understanding of the relationships of how natural and anthropogenic drivers interact with habitats, biological systems, and fish in large rivers is represented by the different 511 512 types of lines in the figure. Solid blue lines depict a strong understanding of the relationship, the dotted-dashed blue line represents a moderate understanding of the relationship, and the black 513 dashed line represents a weak understanding of the relationship. The different types of lines also 514 515 represent the strength of our understanding of within EEC-tier relationships.

516 Increased sediment loads in combination with altered hydraulics and morphology have resulted in high rates of sedimentation, homogeneity of depth, and loss of low-velocity, off-517 channel areas [80, 81]. A diversity of off-channel habitat conditions (i.e., increased residence 518 519 time, low velocity, warm temperatures, availability of food resources) support growth and development of larval and juvenile fishes [82, 83] and often provide important food resources for 520 adult fishes [84-88]. Further, deep, low-velocity off-channel habitats are recognized as important 521 refugia for a wide range of fishes during high-flow events and seasonal periods of low 522 temperatures [89-92]. Loss of floodplain connectivity has eliminated the seasonal exchange of 523 524 nutrients, organisms and organic matter between river and floodplain environments that support biological diversity and productivity [93, 94]. Reduced availability of spawning, nursery, 525 foraging, or overwintering habitat conditions can serve as bottlenecks to fish populations through 526 527 limited larval production, reduced growth, and increased overwinter mortality. For example, high sedimentation rates have been filling backwaters in the Illinois River for decades, thus limiting 528

the availability of overwintering conditions for fishes that bioenergetically need a deep refuge with slow water velocities. Missing year-classes in this reach, represented by truncated size structure in the Largemouth Bass (*Micropterus salmoides*) population are hypothesized to be a result of periodic winter mortality (Fig S6).

The application of our CM makes clear that while the general effects of anthropogenic 533 534 drivers on hydrology, sediment transport, biogeochemistry, and hydraulics and morphology are well understood, there is much less known about how those effects influence the quality and 535 availability of required habitat conditions (Tier 2, Fig 8). Although there is likely overlap of 536 537 habitat requirements among species with similar life histories, the diversity of habitat conditions necessary to support a native and diverse fish community has not been explored. Consequently, 538 the existing information needed to assess the relationship between habitat quality and quantity, 539 and egg production, juvenile recruitment, and adult survival of fish populations within the Upper 540 Mississippi and Illinois rivers was categorized as insufficient to not available (Tier 3 Inter-tier 541 interaction, Fig 8; Table S4). Addressing these knowledge gaps could improve the effectiveness 542 543 of habitat restoration efforts focused on maintaining a diverse native fish community. The spatial and temporal scales of data collection that would support needed inferences to address restoring 544 545 and maintaining native fish biodiversity and habitat quantity and quality are characterized in Fig S7. 546

547

47 Similarities across case studies

548 We observed similarities in the stressors and interactions within EEC tiers across the four 549 case studies. For Tier 1, an altered hydrologic regime was identified as a stressor to the

hydrology EEC in all four rivers (Table 1). Presumably this is due to the ubiquitous effects of 550 dams on the systems examined. However, in some rivers the altered hydrologic regime 551 originated from other anthropogenic (e.g., water use, land use, biological community alteration) 552 553 and natural (e.g., climate) drivers. Similarly, all four case studies listed an altered water temperature regime as a stressor to the biogeochemistry/thermodynamics EEC with linkages to 554 several anthropogenic drivers (Fig 4, 6-8; Tables S1-S4). There were also similarities across case 555 556 studies with respect to the identification of interactions between Tier 1 EECs with all four case studies noting interactions between Tier 1 EEC components. 557

Table 1. Stressors or inter-tier interactions affecting Tier 1 Essential Ecosystem Characteristics (EEC) identified as an information

need in the application of the conceptual model to case studies in the Canadian River (1), Colorado River (2), Columbia River (3), and

560 Upper Mississippi and Illinois rivers (4). Tier 1 EECs are measurable characteristics that describe processes that can significantly alter

the morphological or chemical characteristics within a river channel.

	Tier 1 EEC						
Stressor or inter-tier interaction	Hydrology Channel Morphology/Hydraulics Sediment Transport			Biogeochemistry/ Thermodynamics			
Altered Hydrologic Regime	1, 2, 3, 4	1	1, 2				
Altered Hydraulic Regime		1, 3, 4	3				
Altered Sediment Regime			3, 4				
Altered Water Temperature Regime				1, 2, 3, 4			
Altered Biogeochemical Regime				1, 3, 4			
Channel forming processes		1, 2, 3, 4					
Sediment transport dynamics		1, 2, 3, 4	1, 2, 3, 4				
Sediment adsorption of contaminants and nutrients				1, 2, 3, 4			

562

564	For Tier 2, there were similarities across case studies; however, the adaptation and
565	elaboration of the components to the management goal in the case studies was apparent (Table
566	2). The management goal associated with the case study for the Upper Mississippi and Illinois
567	Rivers resulted in Tier 2 EEC components (e.g., overwintering habitat) and stressors (e.g.,
568	dissolved oxygen) that were unique. Spawning habitat was identified as a Tier 2 EEC component
569	in all the case studies and multiple stressors were identified as affecting this component in two or
570	more of the case studies. Larval fish and invertebrate habitat were noted as Tier 2 EEC
571	components with some similarities in stressors across case studies. Habitat fragmentation,
572	sediment deposition, and water temperature were listed as stressors to Tier 2 EECs in all four
573	case studies. No interactions between Tier 2 EEC components were listed for the case studies.

Table 2. Stressors affecting Tier 2 Essential Ecosystem Characteristics (EEC) identified in the application of the conceptual model to

575 case studies in the Canadian River (1), Colorado River (2), Columbia River (3), and Upper Mississippi and Illinois rivers (4). Tier 2

576 EECs are broadly described as physical, chemical, or biological components of "habitat" that are hypothesized to have overall fitness

577 consequences.

Stressor	Tier 2 EEC (habitat)					
	Spawning	Overwintering	Juvenile fish	Larval fish	Invertebrate	
Altered riparian community					1	
Channel stability					2, 3	
Contaminants	1, 3, 4		4			
Discharge					1	
Dissolved oxygen	4	4	4			
Habitat fragmentation	1, 3, 4			1, 2, 3	2, 3	
Sediment deposition	3, 4	4	4	3	1, 2, 3	
Turbidity				2		
Water Depth	4		4			
Water temperature	1, 2, 3, 4	4	4	1, 3		
Water velocity	3, 4	4	4			

The adaptation of Tier 3 EECs and elaboration of the biological system related to the 579 management goal addressed by the case studies resulted in EECs that were comprised of fish life 580 stages ranging from eggs to adult fish, primary and invertebrate production, and biodiversity 581 582 (Table 3). All Tier 3 EEC components, except biodiversity, were present in the four case studies. Not surprisingly, habitat quantity and quality were listed as stressors to all the EECs related to 583 fish and invertebrates. Six of eight stressors or inter-tier interactions were listed as affecting fish 584 585 larvae and five of eight were noted as affecting egg quantity and quality. In contrast to Tier 2, interactions were extensively noted between Tier 3 EECs and trophic level interspecific 586 interactions were listed in all four case studies. 587

Table 3. Stressors or inter-tier interactions affecting Tier 3 Essential Ecosystem Characteristics (EEC) identified in the application of

the conceptual model to case studies in the Canadian River (1), Colorado River (2), Columbia River (3), and Upper Mississippi and

590 Illinois rivers (4). Tier 3 EECs represent components of the hypothesized biological system upon which the cascading effects of

591 anthropogenic and natural drivers act, and interactions occur.

	Tier 3 EEC						
Stressor or inter-tier interaction	Adult fish recruitment	Juvenile fish recruitment	Larval fish production	Fish egg quality/ production	Invertebrate production	Primary production	Biodiversity
Direct mortality			1	-	-	-	
Predation/competition by invasive species			1, 2, 3	3			4
Habitat quantity/quality	4	4	1, 2, 3	1, 2, 3, 4	1, 2, 3		
Nutrient flux						1, 2, 3	
Trophic level interspecific interactions	1, 2, 3, 4	1, 2, 3, 4	1, 2, 3, 4	1, 2, 3, 4	1, 2, 3, 4	1, 2, 3, 4	4
Predation			3	3			
Mortality	1, 3, 4	1, 2, 3, 4	1, 2, 3				
Fish condition				1, 3, 4			

593 Discussion

The main objective of our exploration of CMs was to impose some structure on the complex ecosystems found in large rivers and from that structure, identify gaps in monitoring information that could inform the management of fish. Comparison across our four case studies provides some insights into large rivers and the utility of the CM to identify gaps in our understanding of factors affecting fish in large rivers.

599 Despite large differences in the physical and ecological contexts of the river systems, the case studies demonstrated substantial commonalities in the data needed to better understand how 600 601 human activities affect these systems and in the application of the CM. The general tiered 602 structure of drivers and cascading responses through EECs worked well with the four examples. 603 Each of the four rivers could be placed in the tiered CM to illustrate current perceptions about 604 drivers and responses. The hierarchical CM generally increased in complexity from top to 605 bottom. Among all rivers, there tended to be greater understanding of links from drivers to Tier 1 and Tier 2 EECs, and less understanding about linkages to Tier 3. 606

The strength of understanding of interactions between anthropogenic and natural drivers 607 and EECs, and between and within EECs, varied considerably among river systems, however, 608 resulting from both variable complexity and existing knowledge. For example, linkages from 609 610 drivers to Tier 1 EECs were considered strong in the case of the Humpback Chub in the Grand Canyon, but between Tier 1 and Tier 2 only moderate. This is probably indicative of the 611 612 substantial research investments in examination of physical processes in this river system [95]. 613 Although we did not prescribe a specific approach to the CM process, the case studies employed similar strategies. Our modelling exercises started with the definition of a management 614

615 goal. In all our case studies, the management goals pertained to a desired biological endpoint 616 represented in Tier 3. After the definition of the management goal, we conceptualized interactions between drivers and EECs and between EECs with a combination of top-down and 617 618 bottom-up approaches. A top-to-bottom approach to working with these models is generally 619 consistent with a management perspective wherein anthropogenic drivers that are most directly 620 managed in a large-river system (e.g., land and water use, etc.) cascade from top to bottom through fluxes to physical and chemical habitats, and then to biological responses. While this is 621 generally true for anthropogenic drivers, a notable exception to the top-to-bottom management 622 623 approach would be that in the U.S., there are few actions currently directed at reducing emissions affecting climate [96] which is a natural driver in our CM. Climate was hypothesized to be a 624 stressor in the case study application of the CM to Arkansas River Shiner management in the 625 South Canadian River and is hypothesized to be affecting hydrologic regimes elsewhere [97, 98], 626 but was not specifically mentioned in other case studies. The CMs can readily be modified to 627 incorporate other factors or pathways (e.g., climate effects) as new information or perspectives 628 629 become available. A bottom to top approach is equally or more valuable as it starts with the foundation of understanding about the species or community, and then seeks to identify which 630 631 stressors affect population or community responses. A bottom-up approach can readily identify information gaps in linkages from ecological processes to demographic parameters [99]. 632 633 The top-to-bottom and bottom-up approaches meet in the middle in Tier 2 in the concept

of habitat: the resources and conditions present in an area that produce occupancy [100]. Tier 2 is

635 critical as it has little value if it is not defined based on biological requirements or if managers

lack understanding on how habitat is formed. Among our examples, the Upper Mississippi River

637 is notable for asserting strong understanding of the linkages from land-use stressors to sediment

regime to diminished overwintering habitat for native adult fishes. After that, interactions with
other processes and life stages combine to increase uncertainty about whether overwintering
habitat is a limiting factor in biodiversity. In contrast, the high confidence in understanding how
White Sturgeon egg quality and production are linked to spawning habitat in the Columbia River
Basin, provides a strong linkage upward through Tier 1 EECs and potential management actions
(Fig 7). Although at times elusive, the concept of habitat is critical for linking management to
biotic endpoints [101].

Large rivers are typically managed for multiple objectives, including fisheries, multi-645 646 species, or ecosystem objectives. Management decisions typically require an understanding of how management actions propagate through a river ecosystem. Although the emphasis may be 647 on a biological endpoint (among other objectives), understanding the intermediate steps and the 648 processes linking them, and potential interactions between processes or EEC components, can 649 help formulate effective management strategies; especially as multiple objectives compete. In a 650 multi-species context, the conceptual models can help identify commonalities and differences in 651 652 in how stressors propagate to biota and therefore provide a basis for prioritizing monitoring efforts. In the case where species or guilds have similar habitat affinities and life histories, a 653 654 dominant anthropogenic stress pathway may be hypothesized and focus on a single or few monitoring components may be justified. An example may be multiple large-river species that 655 are known to be cued to spawn by spring flow pulses. In such a case, the characteristics of the 656 657 annual hydrograph would be a dominant physical monitoring variable and biological monitoring could focus on reproductive success of one or more of the species. In the case where multiple 658 species of concern have different reproductive strategies – for example, rheophilic species like 659 sturgeon that may require in-channel dispersion of young to flowing habitats compared to 660

invasive carp whose young thrive when they can disperse to lentic floodplain pools - pathways 661 662 and monitoring strategies will diverge. In the latter case, it would probably not be sufficient to monitor and assess the characteristics of the annual hydrograph; instead, hydrologic metrics 663 664 would need to be integrated with hydraulic and geomorphic metrics to assess where and when the different habitats would be available and could be targeted for young-of-the-year sampling. 665 666 Effectively addressing multiple species would rely on detailed knowledge of life histories and how they play out on the landscape – such information is missing for many species and may 667 need to be developed for effective design of monitoring and management actions. 668

669 Management actions intended to benefit fish in large rivers can directly or indirectly 670 affect multiple ecosystem components. Without consideration of the effects of management on non-target ecosystem components, unintended consequences may limit management efficacy. 671 672 Hypothesizing inter-tier interactions in the Tier 3 EEC (e.g., see Fig 7), can provide insight on the potential interactions among fish species and other biological components in the context of 673 the hierarchical CM. In all our case studies, the lumping of multiple biological interactions in 674 Tier 3 resulted in a simplification of complex trophic interactions. For example, as Tier 3 675 encompasses all biological responses, it includes multiple life stages of many interacting species 676 677 at varying trophic levels. Because of this, the four CMs diverged significantly at Tier 3 as components were expanded to accommodate existing understanding. Even as the Tier 3 678 components were expanded in complexity, they remained highly simplified views of the 679 680 ecosystem. Simplification was based, in part, on the importance of key species in management goals and the experts' existing knowledge. Even though the hypothesized Tier 3 interactions in 681 682 our case studies conveyed a simplification of the trophic interactions, the hypothesized 683 interactions do suggest the need for information that clarifies the trophic interactions and effects

684 of Tier 3 EEC components on the biological endpoint. If desired or warranted, the Tier 3 EEC could be elaborated to capture more complexity. For example, in Fig 7, the Tier 3 inter-tier 685 interaction between anadromous and resident fishes and white sturgeon larvae could expanded to 686 include interactions with specific fish species. Monitoring can help clarify the effects of 687 688 management actions, including on non-target ecosystem components, but only if data are 689 collected to characterize key ecosystem processes that could affect the outcome. The process of 690 considering and elucidating Tier 3 EEC interactions can help identify the non-target ecosystem components that could be affected when managing for a specific biological endpoint. 691 692 The CMs explored here also provide a framework for considering return on science investments. The knowledge needed for effective management of large rivers can be gained by 693 monitoring intermediate endpoints along the cascade, but the type of information and costs vary 694 widely. Costs for monitoring Tier 1 EECs can be high but some programs are already in place. 695 For example, large rivers are likely to have monitoring infrastructure installed for Tier 1 696 697 monitoring of discharge and temperature regimes, with varying potential for monitoring 698 sediment transport and water quality. Investment at Tier 2 may emphasize physical processes and habitats that can be measured at relatively low cost, assuming that habitats are adequately 699 700 defined based on biological criteria. In larger rivers, Tier 2 habitat assessments can be more cost effective compared to smaller rivers because they can rely on automated data collection through 701 hydroacoustics and remote sensing [101]. As discussed above, habitat assessments have value 702 703 only to the extent that they are based on well-defined biological requirements; it is notable that

some large-river management efforts have found that relatively simple habitat models are useful

705 to predict biological responses [102]. At Tier 3, costs can increase substantially because of

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706 structural uncertainties (i.e., which life stages, which species are most important to monitor) and

707 because of the inherent uncertainties of monitoring fish in large river systems where detection 708 probability can be low and highly variable [36, 103]. Generally, the cost of monitoring increases 709 from Tier 1 to Tier 3 in the CM hierarchy; at the same time, the relevance of information to 710 decision making is typically greater for biological responses depicted in Tier 3 [104]. Because both costs and information benefits increase from Tier 1 to Tier 3 in the CM 711 hierarchy, it is difficult to generalize about where the benefit:cost ratio would be optimized. 712 713 Indeed, as discussed by Jacobson and Berkley [30], the decision about where in the hierarchy monitoring resources would get the highest return on investment may depend more directly on 714 715 managers' and stakeholders' perceptions about risks of acting with incomplete information. For example, the details of how a fish's reproductive strategy depends on the nuances of a seasonal 716 hydrograph may not be known, but stakeholders may believe strongly that the natural 717 718 hydrograph was functional for the species and therefore monitoring of the flow regime will have the highest return on investment and, by extension, restoration of the flow regime is likely to 719 have the most positive effects. On the other hand, in systems where stakeholders opinions are 720 721 divided or socio-economic values would be compromised by a return to a natural flow regime, 722 managers may be required to demonstrate more precisely how elements of the flow regime propagate to species' benefits [105]. Thus, once information needs are identified and there is an 723 assessment of the availability of data identified as information needs, there needs to be a process 724 whereby the costs of collecting the information need to be placed in a socioeconomic context 725 726 (e.g., see [30]). The development of the CMs described in this manuscript can be a first step in 727 application of structured decision-making (SDM) and its iterative form-adaptive management 728

(AM) processes [106, 107]. Structured decision-making is a stakeholder driven process by which

730 a problem can be defined with conceptual models and decomposed into decision components 731 that include the problem context, stakeholder objectives, potential management actions, consequences of those actions on the objectives, and trade-offs related to different decisions 732 733 (actions) [107-109]. One primary focus of SDM is the identification of uncertainties such as those identified in the CMs for the case studies in this paper [110]. Quantification of the 734 735 influence of decision relevant uncertainties can be modeled using sensitivity analysis and other 736 techniques and ranked [107, 109, 111]. In addition, the quantitative techniques available to assist in solving complex ecological problems are robust and range in complexity from consequences 737 738 tables to Bayesian models to dynamic optimization models [107-109, 112, 113]. The SDM 739 process is often used as the set-up phase for adaptive management which includes monitoring over time to reduce uncertainty related to how management will influence important outcomes 740 (e.g. fish population status; [109, 112]. 741

The CM may also help to identify which processes or components are amenable to a field 742 monitoring effort and which are more aptly addressed through laboratory or mesocosm 743 744 experiments. For example, if it is hypothesized that the condition of age 1+ Arkansas Shiners is a critical determining factor in egg quality or production (Fig 4), it could be determined that the 745 746 best approach to developing a quantitative relation between condition and eggs is through a controlled laboratory experiment rather than field-based monitoring. The CM helps to visualize 747 where different types of information may be applied within a decision-making framework. 748 749 A large-river CM may also serve as a precursor to computational ecological or population models [30]. Similar questions about how monitoring and other science efforts should be 750 751 distributed among EECs and processes can be addressed iteratively by carrying out sensitivity analyses in a modeling framework. Indeed, given substantial uncertainties associated with 752

monitoring data, computation modeling can be considered a necessary component of large-river
monitoring and evaluation systems [99, 114].

755 Conclusions

We found the process of conceptualizing the relationships between and within EECs 756 757 fostered a critical assessment of what we know about factors affecting the management endpoint. 758 By visualizing how EEC drivers directly and indirectly affect management endpoints, our CM 759 identified critical information gaps and uncertainties that, if resolved, could improve our 760 understanding of how to best meet management objectives. The process of conceptualizing the EEC relationships affecting fish in large rivers could help to structure, or restructure, monitoring 761 programs around scientifically sound monitoring questions, promote the selection of relevant 762 ecological indicators that characterize resource condition or management outcomes, and 763 facilitate communication and information sharing within and between organizations managing or 764 765 researching management endpoints. Ultimately, understanding the mechanisms by which EECs influence large-river fishes will improve the effectiveness of restoration and management 766 actions. 767

As shown with our case studies, our CM is flexible and applicable to a wide range of river systems with different anthropogenic drivers and management objectives. We feel our CM provides a generic structure that scientists can adapt to their management goals and needs. By not being overly prescriptive, for example, with respect to the components of the Tier 2 and 3 EEC components, scientists can adapt the CM to different biological communities and

management endpoints. By doing so, we feel that users have the flexibility to place their
management questions in the context of EECs that are specific to their large-river system.

Although the case studies addressed management issues that were river or basin specific, there were similarities relative to information needs and data availability. For example, in most systems information on river discharge and water temperature were needed and available. Conversely, information regarding trophic relationships and the habitat requirements of larval fishes were generally lacking. This result suggests that there may be a common need for a better understanding of certain factors across large-river systems.

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- names is for descriptive purposes only and does not imply endorsement by the U.S. Government.
- An animal care and use protocol was not required for this research.

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1141 Supporting Information Captions

Fig S1. Discharge (m³/s) patterns over time in the Canadian River, Oklahoma near Canadian, TX
(USGS 2019; USGS Gage 07228000).

1144 Fig S2. Hydrograph showing pre- and post-Glen Canyon Dam closure in 1964 (dashed line)

1145 mean monthly discharge (m^3/s) , which transitions from seasonally stochastic to a more

1146 homogeneous regime focusing on anthropogenic interests.

1147 Fig S3. The spatial and temporal scales of the management goal, the scientific inferences needed to inform the management goal, and that data collection needs to occur to support the inferences 1148 1149 for monitoring information needs identified as requiring additional data in the case study 1150 addressing Humpback Chub recruitment in the Colorado River between Glen Canyon Dam and 1151 Lake Mead, Arizona (see Table S2 for additional detail). A:Tier 2 EEC= larval Humpback Chub 1152 habitat; Stressors=habitat fragmentation, turbidity and Tier 2 EEC= Humpback Chub spawning 1153 habitat; Stressor= water temperature and Tier 3 EEC=insect production; Stressor=benthic macroinvertebrate habitat quantity and quality and Tier 3 EECs=all; Inter-tier interaction=trophic 1154 1155 level interactions; B:Tier 3 EEC=larval Humpback Chub production; Stressor= larval Humpback chub habitat quantity and quality; C: Tier 3 EEC=Humpback Chub egg quality and production; 1156 1157 Stressors= Humpback Chub spawning habitat quantity and quality; D:Larval Humpback Chub production; Inter-tier interaction=mortality of Humpback Chub eggs; E:Tier 3 EEC= larval 1158 Humpback Chub production; Stressor=predation by invasive species and Tier 3 EEC=Primary 1159 1160 production; Stressor=nutrient flux; F:Tier 1 EEC= biogeochemistry/thermodynamics; Inter-tier interaction=sediment adsorption of contaminants and nutrients; G:Tier 3 EEC=Humpback Chub 1161 1162 age-0 recruitment; Inter-tier interaction=mortality of larval Humpback Chub.

Fig S4. Proportion of total annual Columbia River discharge at The Dalles, OR occurring in themonth of June from 1879 to 2015.

Fig S5. The spatial and temporal scales of the management goal, the scientific inferences needed 1165 1166 to inform the management goal, and that data collection needs to occur to support the inferences 1167 for monitoring information needs identified as requiring additional data in the case study 1168 addressing White Sturgeon recruitment in the Columbia River (see Table S3 for additional detail). A:Tier 1 EEC= biogeochemistry/thermodynamic; Stressor= altered biogeochemical 1169 regime and Tier 2 EEC= benthic macroinvertebrate habitat; Stressors=channel stability, sediment 1170 1171 deposition, fragmentation and Tier 2 EEC=Larval White Sturgeon habitat; Stressors=habitat 1172 fragmentation sediment deposition, water temperature and Tier 2 EEC=White Sturgeon spawning habitat; Stressors=contaminants, sediment deposition and Tier 3 EEC=White Sturgeon 1173 1174 egg quality and production; Stressor=predation by invasive species and Tier 3 EEC=White Sturgeon larvae production; Stressor=predation by invasive species and Tier 3 EEC=benthic 1175 macroinvertebrate production; Stressor=benthic macroinvertebrate habitat quantity and quality; 1176 1177 B:Tier 1 EEC=sediment transport; Stressors=altered sediment regime, altered hydraulic regime and Tier 1 EECs=channel morphology/hydraulics, sediment transport; Stressor=altered hydraulic 1178 1179 regime and Tier 3 EEC=primary production; Stressor=nutrient fluxes; C:Tier 1 EEC=channel morphology/hydraulics, sediment transport; Stressor=altered hydraulic regime and Tier 1 EEC = 1180 channel morphology/hydraulics, sediment transport; Inter-tier interaction=sediment transport 1181 1182 dynamics; D:Tier 1 EEC=channel morphology/hydraulics, sediment transport; Stressor=altered 1183 hydraulic regime and Tier 1 EEC=channel morphology/hydraulics, sediment transport; Inter-tier 1184 interaction=sediment transport dynamics; E:Tier 1 EEC=biogeochemistry/thermodynamics; 1185 Inter-tier interaction=sediment adsorption of contaminants and nutrients; F: Tier 3 EEC=White

1186 Sturgeon larvae production; Stressors=larval White Sturgeon habitat quantity and quality and

1187 Tier 3 EEC=White Sturgeon larvae production, age-0 White Sturgeon recruitment; Inter-tier

1188 interaction=mortality and White Sturgeon egg quality and production; Inter-tier

1189 interaction=predation of White Sturgeon eggs by native fish and Tier 3 EECs=White Sturgeon

1190 larvae production; Inter-tier interaction=predation of White Sturgeon larvae by native fish and

1191 Tier 3 EECs=all; Inter-tier interactions=trophic level interactions.

1192 Fig S6. Largemouth bass (*Micropterus salmoides*) data from the Pool 13 of the Upper

1193 Mississippi River (A, B) and the La Grange Pool of the Illinois River (C, D). The two river

reaches are roughly the same latitude, but the La Grange Pool is more limited in overwintering

1195 habitat. Population abundance is presented in panels A and C where each point is an individual

1196 fish cumulatively caught with standardized day time electrofishing (Ratcliff et al. 2014) in a

specific year. The dashed triangle highlights 'missing' >400 mm size classes since 2000 in the

La Grange Pool. Population size structure is indexed by proportional stock density (PSD) is

1199 presented in panels B and D with the dashed line showing trends in the largest size classes over

time. Data and methodology were downloaded from the publicly available databases via the

1201 Upper Mississippi River Restoration's Long Term Resource Monitoring Graphical Fish Browser

1202 [115].

Fig S7. The spatial and temporal scales of the management goal, the scientific inferences needed to inform the management goal, and that data collection needs to occur to support the inferences for monitoring information needs identified as requiring additional data in the case study addressing native fish biodiversity and habitat diversity in the Mississippi and Illinois rivers (see Table S4 for additional detail). A:Tier 1 EEC=sediment transport; Stressor=altered hydraulic

regime and Tier 1 EEC=biogeochemistry/thermodynamics; Stressor=altered biogeochemical

regime and Tier 1 EEC=biogeochemistry/thermodynamics; Inter-tier interaction=sediment 1209 1210 adsorption of contaminants and nutrients; B: Tier 2 EEC=adult native fish overwintering habitat; Stressors=water velocity, water temperature, dissolved oxygen, sediment deposition and Tier 2 1211 1212 EEC=juvenile native fish habitat; Stressors=water depth, water velocity, water temperature, dissolved oxygen, contaminants, sediment deposition and Tier 2 EEC=native fish spawning 1213 1214 habitat; Stressors=water depth, water velocity, habitat fragmentation, sediment deposition, water temperature, dissolved oxygen, contaminants and Tier 3 EEC=adult and juvenile native fish 1215 recruitment; Inter-tier interaction=mortality and Tier 3 EEC=all; Stressors=invasive species and 1216 1217 Tier 3 EEC=all; Inter-tier interaction=trophic level interactions; C:Tier 1 EEC=channel 1218 morphology/hydraulics; Inter-tier interaction=channel forming processes; D:Tier 1 EEC=channel morphology/hydraulics, sediment transport; Inter-tier interaction=sediment transport dynamics; 1219 1220 E:Tier 3 EEC=adult native fish recruitment; Stressor=adult native fish overwintering habitat quantity and quality and Tier 3 EEC=juvenile native fish recruitment; Stressor=juvenile native 1221 fish habitat quantity and quality and Tier 3 EEC=native fish egg quality and production; 1222 1223 Stressor=spawning habitat quantity and quality. 1224 Table S1. Summary of information needs identified in the Conceptual Model describing factors 1225 affecting the recruitment of the Arkansas River Shiner in the South Canadian River, OK (Fig 4; this publication) by Essential Ecosystem Characteristic (EEC) Tier, EEC, and stressor or within 1226 Tier interactions and an assessment of the status of existing information that could be used to 1227 1228 address the information needs. 1229 Table S2. Summary of information needs identified in the Conceptual Model describing factors 1230 affecting the recruitment of the Humpback Chub in the Colorado River, Arizona (Fig 6; this

1231 publication) by Essential Ecosystem Characteristic (EEC) Tier, EEC, and stressor or inter-tier

interactions and an assessment of the status of existing information that could be used to addressthe information needs.

1234 Table S3. Summary of information needs identified in the Conceptual Model describing factors

1235 affecting the recruitment of age-0 White Sturgeon in the Columbia River (Fig 7; this

1236 publication), by Essential Ecosystem Characteristic (EEC) Tier, EEC, and stressor or inter-tier

1237 interactions and an assessment of the status of existing information that could be used to address

the information needs.

1239 Table S4. Summary of information needs identified in the Conceptual Model describing factors

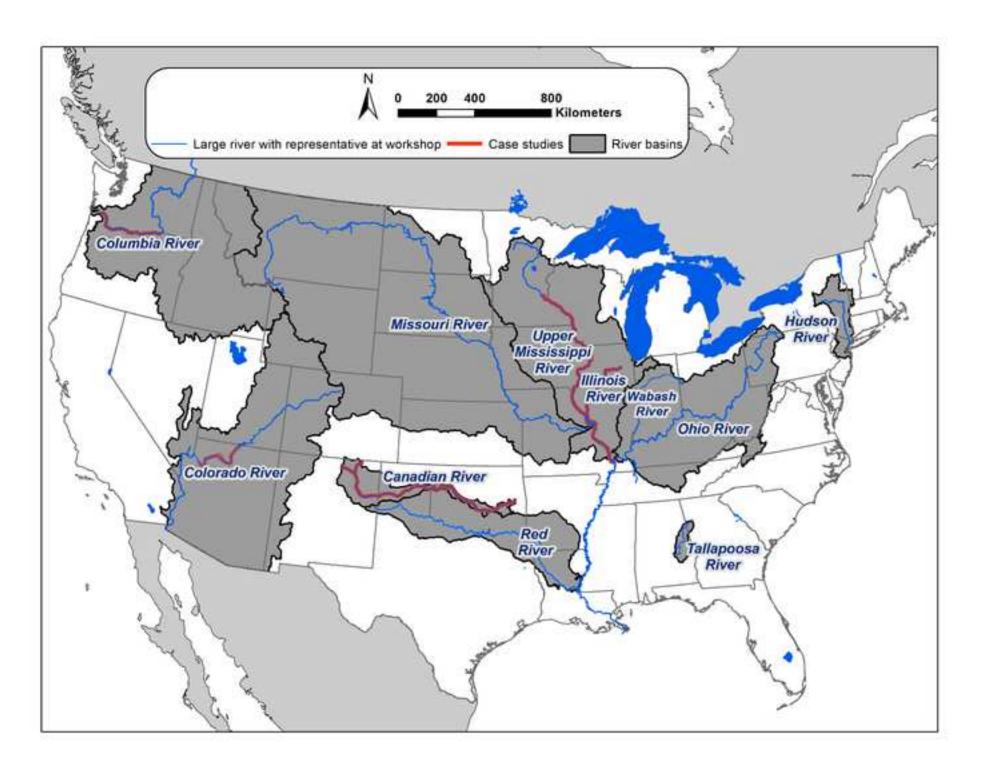
1240 affecting the restoration and maintenance of native fish biodiversity and habitat quantity and

1241 quality in the Upper Mississippi and Illinois rivers (Fig 8; this publication), by Essential

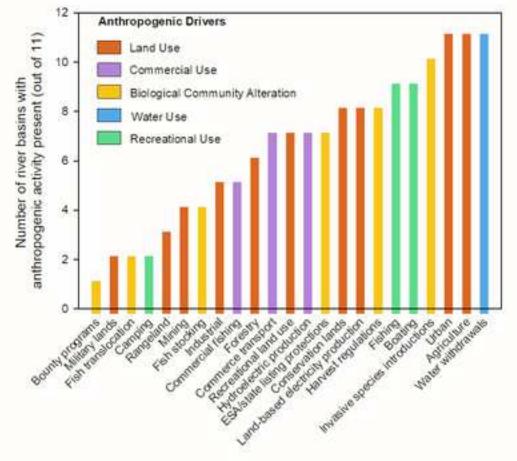
1242 Ecosystem Characteristic (EEC) Tier, EEC, and stressor or inter-tier interactions and an

assessment of the status of existing information that could be used to address the information

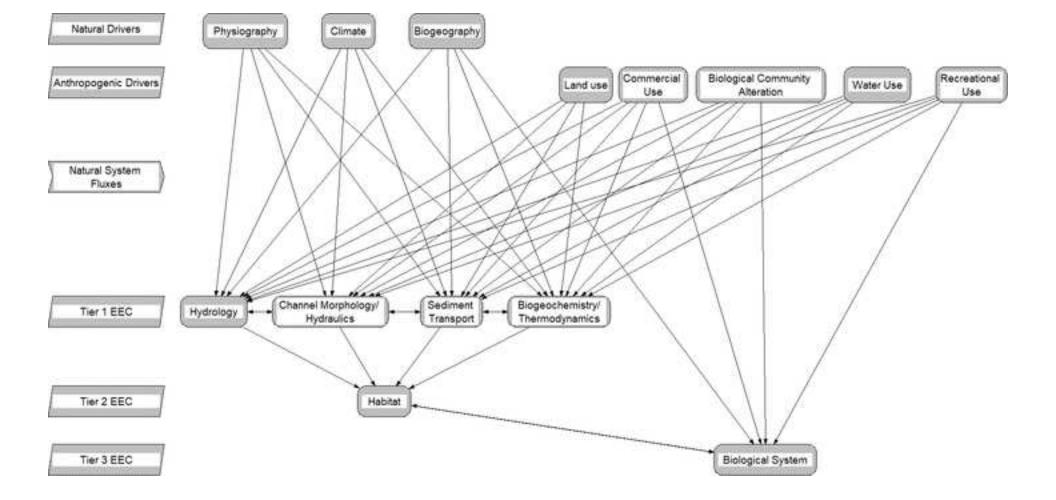
needs.

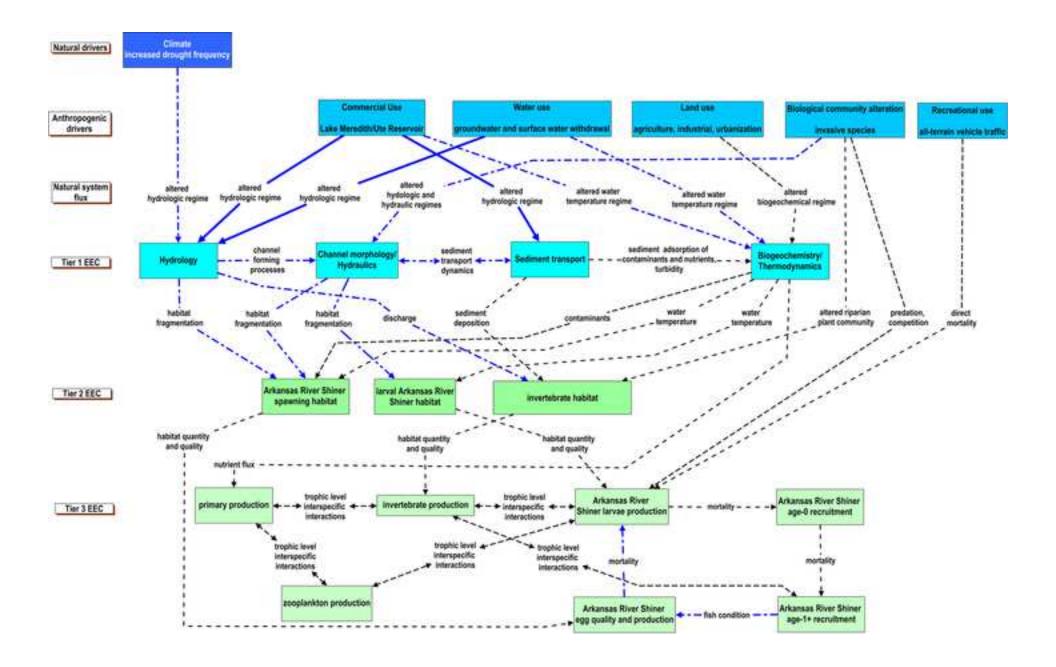


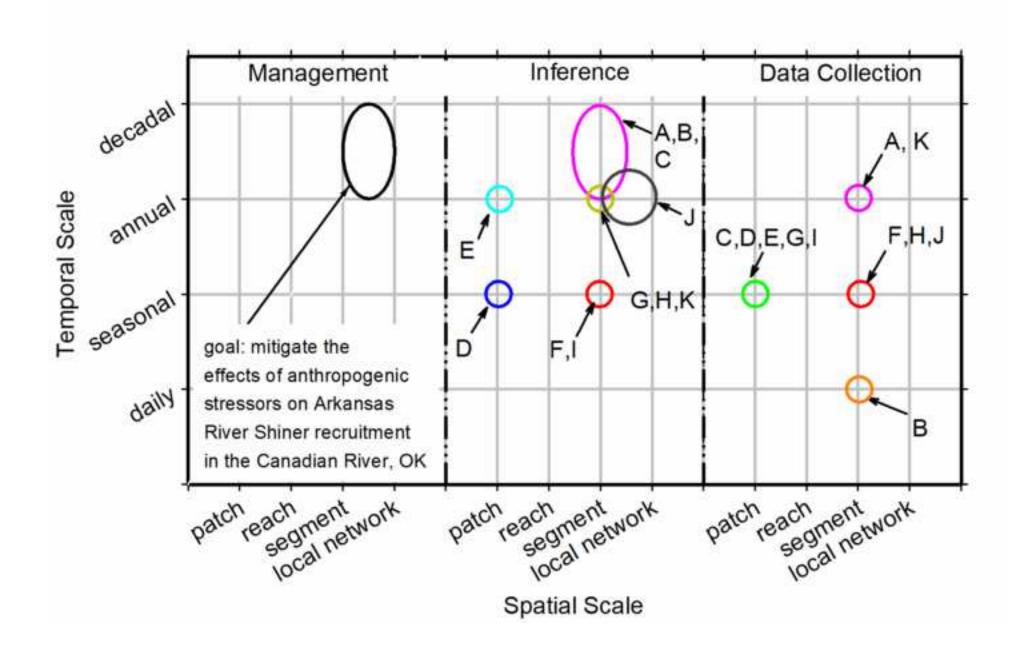


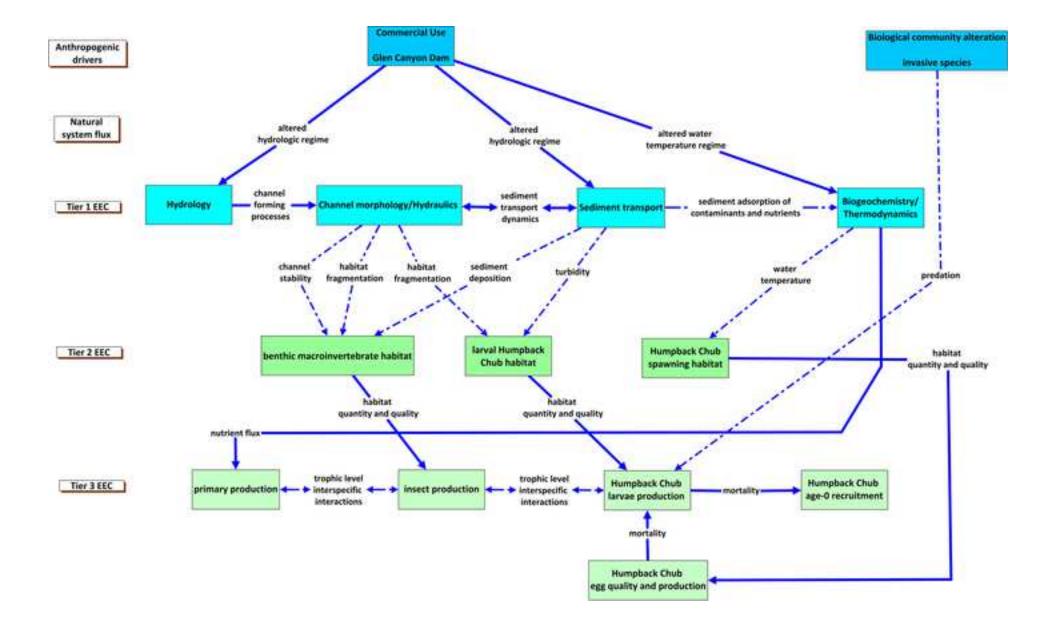


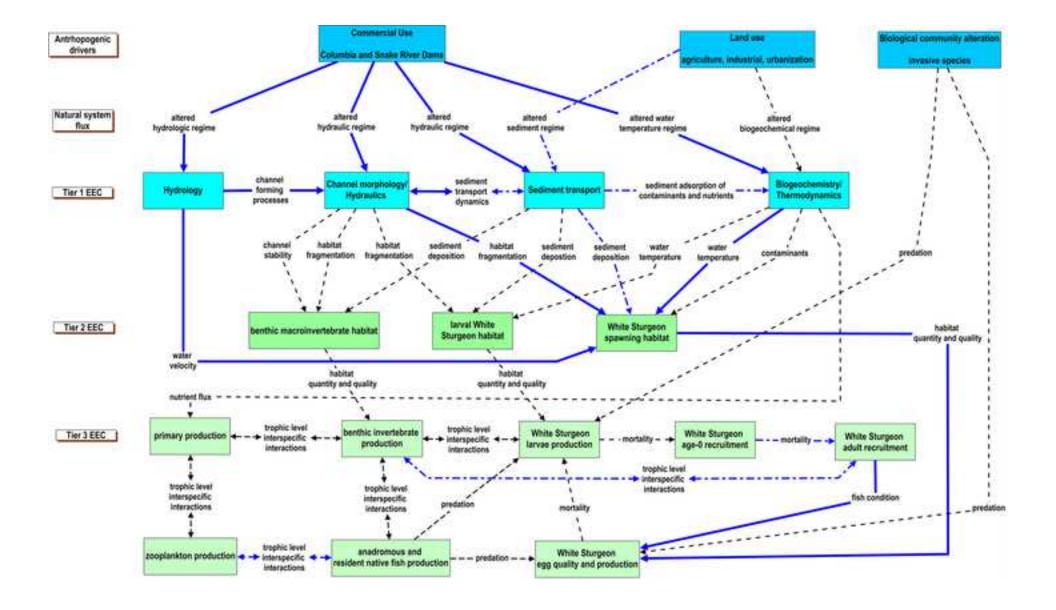
Anthropogenic activities

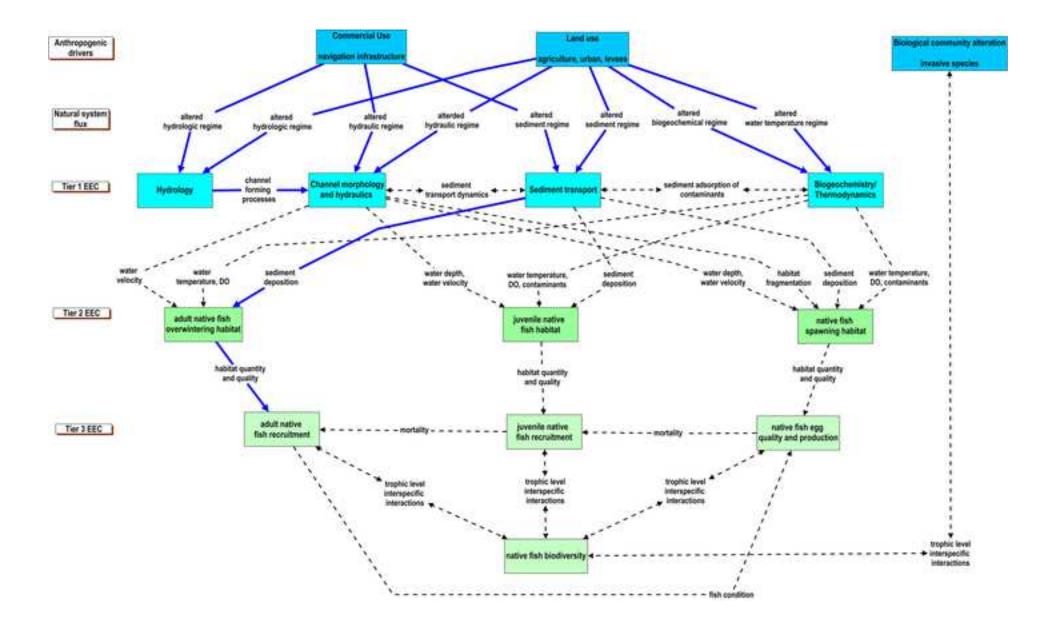












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Click here to access/download Supporting Information Table S4_Counihan_et al. LR CM.docx

- 1 Title: Identifying monitoring information needs that support the management of fish in large
- 2 rivers
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- 30

31	Abstract – Management actions intended to benefit fish in large rivers can directly or indirectly
32	affect multiple ecosystem components. Without consideration of the effects of management on
33	non-target ecosystem components, unintended consequences may limit management efficacy.
34	Monitoring can help clarify the effects of management actions, including on non-target
35	ecosystem components, but only if data are collected to characterize key ecosystem processes
36	that could affect the outcome. Scientists from across the U.S. convened to develop a conceptual
37	model that would help identify monitoring information needed to better understand how natural
38	and anthropogenic factors affect large river fishes. We applied the conceptual model to case
39	studies in four large U.S. rivers. The application of the conceptual model indicates the model is
40	flexible and relevant to large rivers in different geographic settings and with different
41	management challenges. By visualizing how natural and anthropogenic drivers directly or
42	indirectly affect cascading ecosystem tiers, our model identified critical information gaps and
43	uncertainties that, if resolved, could inform how to best meet management objectives. Despite
44	large differences in the physical and ecological contexts of the river systems, the case studies
45	also demonstrated substantial commonalities in the data needed to better understand how
46	stressors affect fish in these systems. For example, in most systems information on river
47	discharge and water temperature were needed and available. Conversely, information regarding
48	trophic relationships and the habitat requirements of larval fishes were generally lacking. This
49	result suggests that there may be a common need for a better understanding of certain factors
50	across large-river systems.

52 Introduction

53	Long-term monitoring has benefited a variety of marine and freshwater ecosystems,
54	including large rivers like the Ohio [1, 2] and Illinois rivers [3]. Large river systems are
55	complex, making the development of effective monitoring programs especially difficult. Large
56	rivers are dynamic systems with high variability in spatio-temporal physicochemical
57	characteristics and biotic assemblages [4]. The inherent complexity of large rivers makes biotic
58	assemblages logistically difficult to sample [5] and the mechanisms of change difficult to
59	understand. Large rivers represent the culmination of vast stream networks and, thus, integrate
60	and accumulate the effects of multiple stressors at varying spatial scales [6]. The spatial and
61	temporal complexity associated with large rivers has hindered the identification of mechanisms
62	driving declining populations of aquatic species-Long-term monitoring has benefited a variety of
63	marine and freshwater ecosystems, including large rivers like the Ohio [1, 2] and Illinois rivers
64	[3]. Large-river systems are complex, making the development of effective monitoring programs
65	especially difficult. Large rivers are dynamic systems with high variability in spatio-temporal
66	physicochemical characteristics and biotic assemblages [4]. The inherent complexity of large
67	rivers makes biotic assemblages logistically difficult to sample [5] and the mechanisms of
68	change difficult to understand. Large rivers represent the culmination of vast stream networks
69	and, thus, integrate and accumulate the effects of multiple stressors at varying spatial scales [6].
70	The spatial and temporal complexity associated with large rivers has hindered the identification
71	of mechanisms driving declining populations of aquatic species [7-10]. To exacerbate the
72	complexity, large rivers commonly have within-channel structural alterations (e.g., dams, river
73	training structures, [11]) and often exhibit legacy effects from historical land uses [12]. To deal

74	with the complexity, some areas of aquatic science recommend monitoring be used to test the
75	linkages developed first through conceptual models (e.g., environmental flows, . To exacerbate
76	the complexity, large rivers commonly have within-channel structural alterations (e.g., dams,
77	river training structures, [11]) and often exhibit legacy effects from historical land uses [12]. To
78	deal with the complexity, some areas of aquatic science recommend monitoring be used to test
79	the linkages developed first through conceptual models (e.g., environmental flows, [13-15]).
80	Conceptual models are useful tools to help guide the design of monitoring programs
81	[16][16]. The identification of questions relevant to conservation and management efforts
82	requires some foresight and knowledge of the complexities of the system being monitored. For
83	example, it is generally well accepted that the native range of the federally-listed Arkansas River
84	Shiner (<i>Notropis girardi</i>) is truncated [17][17], though there is uncertainty surrounding the
85	multiple threats affecting the species [18]. Reducing the uncertainty associated with the decline
86	of the Arkansas River Shiner through the implementation of a hypothesis-driven monitoring
87	program would facilitate confidence in moving forward with a recovery plan. This is where
88	conceptual models are quite useful; they can serve as the foundation to guide hypothesis-driven
89	monitoring programs [14, 16] and identify key ecosystem processes and factors that may directly
90	or indirectly affect management outcomes [19-22].
91	Understanding factors affecting the status and trends of fishes is of interest to multiple
92	stakeholder groups across multiple jurisdictions. Fishes provide economic benefits to businesses
93	that serve recreational interests, commercial and recreational fishers, tribal members for whom
94	fish are an integral part of their cultural identity [23][23], and to local and state governments who
95	derive revenue from these activities. Fish populations are affected by the integration of physical
96	habitat, water quality, environmental contamination, habitat fragmentation, and overall

97	ecosystem productivity [24-27]. Consequently, fish are often the focus of management and
98	monitoring programs (e.g., [28]). However, because fish integrate the effects of so many
99	components of the ecosystem, the success of efforts to manage fishes can be affected by
100	unintended consequences of mitigation on factors not directly targeted by the actions. Without
101	consideration of the effects of management on non-target ecosystem components, unintended
102	consequences may limit management efficacy.

103 Our goal is to demonstrate how a structured, yet flexible, conceptual model (CM) can be used to identify the types of monitoring information needed to understand the range of factors 104 105 affecting large-river fishes. Our CM includes a hierarchically structured conceptualization of 106 ecosystem characteristics based on CMs originally developed by Harwell, Myers (29) and 107 elaborated by Jacobson and Berkley (30). Jacobson and Berkley (30). We chose to incorporate 108 the tiered conceptualization of ecosystem characteristics proposed by Jacobson and Berkley 109 (30)Jacobson and Berkley (30) in part because it allows users to define their own biotic or abiotic 110 interests. In this paper, we discuss the structure and development of the CM. We apply the CM to case studies to illustrate the flexibility and applicability of this approach and use it to identify 111 monitoring information needs specific to disparate management goals. More specifically, for 112 113 each case study, we use the CM to hypothesize how human activities affect fish populations and then identify information needs required to evaluate the hypothesized relationships. We then 114 115 posit the spatial and temporal scales of the management goal addressed in the conceptual model, inferences needed to inform the management goal, and data collection requirements needed to 116 make the inferences. 117

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119 Conceptual Model

120 Overview of Approach

121 Since 2012, scientists working on large rivers across the United States have participated in a forum intended to improve our understanding of large-river ecosystems. The collaborative 122 forum has worked to identify best practices of long-term monitoring programs [31] and evaluate 123 trends in fish assemblages across rivers [32]. As this group of scientists moved toward linking 124 changes in fish populations and assemblages to human activities, there was a need to develop a 125 126 process to help identify and prioritize the information needed to assess trends in large river 127 fishes. To that end, a workshop was convened in Hood River, Oregon in May 2017, to jointly adapt, apply, and qualitatively evaluate a conceptual model for developing hypotheses that detail 128 stressors affecting fishes arising from natural and anthropogenic sources [33][33]. 129 Our general approach was to first identify human activities that affect large-river fishes 130 and then hypothesize how the activities related to physical and chemical factors and biological 131 132 communities. Prior to the workshop, we elicited opinion from experts that comprise a U.S. 133 Geological Survey (USGS) led forum on large river monitoring about important anthropogenic 134 activities that could affect fish populations or communities in the river basins they work in. (Fig 1). We summarized the information from this exercise and grouped the anthropogenic activities 135 into driver categories (Fig 2) and proposed a general form of the CM. We then disseminated the 136 information to the experts prior to the workshop. 137

Fig 1. Map of rivers and watersheds represented by scientists that convened to develop a
conceptual model that <u>would depictdepicts</u> how natural and anthropogenic drivers interact with

habitats, biological systems, and fish in large rivers. River segments where we conducted case 140 141 studies that applied the conceptual model to identify monitoring information needs associated with management goals are highlighted in red. 142 143 Fig 2. The types of anthropogenic activities affecting large river fishes across 11 river systems as + 144 identified from a query to expert participants of a workshop convened to develop a conceptual model that would depict how natural and anthropogenic drivers interact with habitats, biological 145 146 systems, and fish in large rivers. Meeting participants included research Fig 2. The results of a 147 query to scientists from the Canadian, Colorado, Columbia, Hudson, Illinois, Ohio, Missouri, 148 Red, Tallapoosa, Upper Mississippi, and Wabash Rivers, U.S. to identify anthropogenic 149 activities that affect large-river fishes in the river systems they represent. Anthropogenic activities were classified into five anthropogenic driver categories. 150 151 During the workshop we discussed and refined the CM form (Fig 3). We then had representatives from each river system represented at the workshop choose a management goal 152 153 to address. Then, through a facilitated discussion led primarily by the representative of the river 154 system being addressed, we 1) elaborated tiered conceptualizations of ecosystem characteristics 155 to reflect the large-river systems and management goals being examined, 2) used knowledge of the fish species' life history and population bottlenecks to relate biological ecosystem 156

the fish species' life history and population bottlenecks to relate biological ecosystem
characteristics to habitat requirements, 3) hypothesized pathways describing how anthropogenic
and natural drivers affect large-river fish populations either indirectly (i.e.g., effects on flow

regime, habitat, trophic resources, etc.) or directly (e.g., competition with invasive species), and

160 4) hypothesized interactions within ecosystem characteristic tiers that could affect the

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management goal. Based on this exercise, we chose four case studies to refine for use in thismanuscript (Fig 1).

- 163 Fig 3. Tiered hierarchical conceptualization of how anthropogenic and natural drivers relate to
- 164 physical and biological components of large-river ecosystems. Essential ecosystem
- 165 characteristics (EECs) are groupings of ecosystem components. Tier 1 EECs represent physical
- 166 and chemical effects; fundamental measures of process that are directly affected by
- 167 anthropogenic and natural drivers. Tier 2 EECs represent a broad habitat category that is
- 168 intended to encompass the physical, chemical, and biological components of the riverine habitats
- 169 that influence reproduction, growth, and survival of biotic communities. The Tier 3 EEC
- 170 represent components of the biological systems that respond to changes in the hierarchical
- 171 components of the conceptual model.

After the workshop, we held a series of conference calls with workshop participants with 172 expertise in the selected case studies to refine all aspects of the CMs and associated information. 173 174 During the calls, we started with the CM from the workshop and discussed and clarified CM 175 components, pathways, and inter-tier interactions. We then characterized whether, based on the 176 expert knowledge of workshop participants, there was a strong, moderate, or weak understanding of the pathways and interactions. We then developed aA list of the information needed to 177 178 understand the relationships described in the CM was developed. We then had the case study leader classify whether the data required to understand the information needs were available, 179 insufficient, or not available. For information needs that were classified as insufficient or not 180

available, we-then characterized the spatial and temporal scales at which data should be collected
to make inferences that support the evaluation of management goals.

We then encouraged representatives from each basin to share the CM and the case study narrative with other experts familiar with the river system and management goal. This outreach took several forms including sharing the CM with working groups tasked with implementing the management goal, discussions with peers familiar with the management goal, and presenting the CM at regional conferences. The intent was to garner opinions from outside the workshop participants. If needed, the CMs incorporated the feedback received.

189 Hierarchical structure of Conceptual Model

Our CM is a hierarchical conceptualization of how anthropogenic and natural drivers 190 191 relate to multiple tiers representing the physical and biological components of large-river ecosystems (Fig 3). Natural drivers included in the CM were physiographic, climatic, and 192 biogeographic factors that control fluxes of water, mass, energy, and genetic information in a 193 194 watershed [30][30]. The physiographic factors, such as lithology, soils, and watershed topography, exert control on water, sediment, and geochemical fluxes (e.g., nutrients) into the 195 196 river corridor. Physiography is generally static over time frames of decades to centuries. Climate controls fluxes of atmospheric energy and moisture into the watershed. Unlike physiography, 197 198 climate is more likely variable to vary over relatively shorter temporal scales. Biogeography 199 describes the native organism assemblage in the watershed (e.g., $\frac{[34]}{[34]}$) and the natural flux of genetic information due to immigrations, emigrations, mutations, and extinctions. Changes 200 201 arising from the biogeography driver includes altered spatial distribution of organisms within the

watershed, which, in turn, may alter the effects of natural system regimes on the river corridor.
For example, natural variation of the type and distribution of vegetation can affect the time series
and magnitude of runoff events [30][30].

205 We created five categories of anthropogenic drivers to characterize a range of human 206 activities that affect large rivers: land use, commercial use, biological community 207 manipulationalteration, water use, and recreation (Fig 2). The land-use category is intended to reflect different ways humans use landscape resources that affect large rivers. We defined 208 commercial use as the use of river resources for marketable enterprises that did not involve water 209 210 removal or transfer. We included biological community alteration to represent the intentional or 211 non-intentional human alteration or manipulation of the river's biological community (e.g., introductions of non-native fish). Recreational use was defined as the use of river resources for 212 213 leisure activities (e.g., fishing, boating). We considered water use a direct commercial or noncommercial use of river water that involved the removal or transfer of water. 214 215 Our CM includes hierarchically structured essential ecosystem characteristics (EEC), 216 originally developed by Harwell, Myers (29) and described in detail by Jacobson and Berkley 217 (30). Jacobson and Berkley (30). Briefly, EECs are characteristics that can be classified into 218 similar groups based on the way they link to biological endpoints [30] [30]. Tier 1 EECs are 219 measurable characteristics that describe processes that can significantly alter the morphological 220 or chemical characteristics within a river channel. The Tier 1 categories we considered were 1) 221 Hydrologyhydrology, 2) Channel Morphology/Hydraulicschannel morphology/hydraulics, 3) 222 Sedimentsediment transport, and 4) 223 Biogeochemistry/Thermodynamicsbiogeochemistry/thermodynamics. Tier 2 EECs are broadly

224 described as physicochemical or biological components of "habitat" that are hypothesized to 11 225 affect (e.g., growth, survival, reproduction, [35][35]) fish populations or assemblages. Lastly, 226 Tier 3 EECs represent components of the hypothesized biological system that are affected by the cascading (e.g., degradation of egg quality caused by increases in sediment deposition) or direct 227 228 (e.g., predation by invasive species) effects of anthropogenic and natural drivers. Tier 2 229 characteristics are particularly important because these are the factors that can be examined at 230 scales most often sampled by fisheries managers (e.g., Smallmouth Bass Micropterus dolomieu, 231 [36]).[36]. The specific components that comprise Tier 2 and 3 EECs are flexible and can be adapted and elaborated depending on the river system and specific management goal being 232 233 addressed.

234 We retained aspects of the approach taken by Jacobson and Berkley (30) Jacobson and Berkley (30) with respect to how our model represents interactions between drivers and EECs, 235 236 but with key differences. Since we were interested in representing how human activities affect 237 large river ecosystems, our approach acknowledges that anthropogenic and natural drivers 238 interact and alter the expected characteristics of Tier 1 EECs. Similar to Jacobson and Berkley 239 (30)Similar to Jacobson and Berkley (30), our model depicts a stress associated with a natural or anthropogenic driver to Tier 1 EECs as fluxes in natural system regimes that alter the frequency, 240 241 magnitude, duration, timing, or rate of change in natural systems or by the imposition of a hardstructural constraint on channel form. The natural system regimes considered in our CM were 242 hydraulic, hydrologic, sediment, temperature, light, and biogeochemistry. Graphically, the 243 natural system fluxes were represented by arrows connecting anthropogenic and natural drivers 244 to Tier 1 EECs. Similarly, hypothesized pathways between EECs, that depict the expression of 245 the cascading effects of anthropogenic and natural drivers, and interaction within EECs were 246 247 depicted as arrows. AllFor example, fragmentation of river systems resulting from altered

248	hydrologic and/or hydraulic regimes caused by dams, weirs, levees, and other factors are
249	frequently cited sources of stress to large-river fishes [37]. Fragmentation can prevent fish from
250	migrating and/or dispersing through their natural reproductive ranges and from accessing critical
251	habitats [38]. To depict a scenario where the presence of a dam is altering hydrologic and/or
252	hydraulic regimes resulting in habitat fragmentation, the CM would show an arrow from an
253	anthropogenic stress (i.e., dam as a commercial activity) to a Tier 1 EEC (e.g., channel
254	morphology/hydraulics) that would depict a natural system flux (e.g., altered hydrologic and/or
255	hydraulic regime) that would then manifest as a stress caused by habitat fragmentation depicted
256	by an arrow between the Tier 1 EEC and a Tier 2 EEC (e.g. habitat) that would then manifest as
257	an effect on a Tier 3 component, shown by an arrow between Tier 2 and Tier 3. All stress
258	pathways and interactions were classified with respect to the strength of understanding of the
259	relationships based on expert opinion. Arrows with solid blue lines depict a strong understanding
260	of the relationship, dotted-dashed blue lines represent a moderate understanding of the
261	relationship, and with a black dashed line represent a weak understanding of the relationship.
262	Spatial and temporal context
263	The successful characterization of how human activities influence large-river fishes is
264	dependent upon integrated concepts of scale. Fish distributions in rivers can vary spatially within
265	river basins in relation to naturally occurring and human induced landscape characteristics [32].
266	Fish distributions can also vary seasonally, annually, and over longer times in response to
267	changing environmental conditions [3739]. Consequently, the spatial and temporal scope of fish
268	management goals often varies within and between large-river systems and agencies. For data

13

269	collected by monitoring programs to have the highest relevance, the spatial and temporal scales
270	appropriate for scientific investigation and management must also be time and geographic-
271	context specific. For example, the management of White Sturgeon (Acipenser transmontanus) in
272	the Columbia River varies by reservoir or river segment and season [38, 3940, 41]. The effects of
273	hydropower development on White Sturgeon vary spatially and temporally as well, so the spatial
274	and temporal context of the data needed to understand the effects needs to be considered. For
275	instance, hydropower peaking operations, that can vary by dam and season, affect river discharge
276	in a river reach on a diel and even hourly basis [4042], whereas water storage and other
277	management actions can affect seasonal discharges over a broader geographic scale [38].[40].
278	Understanding the spatial and temporal context needed to inform management will help ensure
279	relevant information is collected.
280	We considered spatial and temporal resolution in our CMs. We defined spatial extent as :
280 281	We considered spatial and temporal resolution in our CMs. We defined spatial extent as : Local Networklocal network – synonymous with Hydrologic Unit Code (HUC) 2 basins [41,
281	Local Networklocal network – synonymous with Hydrologic Unit Code (HUC) 2 basins [41,
281 282	Local Network <u>local network</u> – synonymous with Hydrologic Unit Code (HUC) 2 basins [4 1, 4243, 44]; Segmentsegment – the portion of a river between two major tributary confluences
281 282 283	Local Networklocal network – synonymous with Hydrologic Unit Code (HUC) 2 basins [41, 4243, 44]; Segmentsegment – the portion of a river between two major tributary confluences [43][45] or other hydrogeomorphic features [44]; Reach – the length of river occurring between
281 282 283 284	Local Networklocal network – synonymous with Hydrologic Unit Code (HUC) 2 basins [41, 4243, 44]; Segmentsegment – the portion of a river between two major tributary confluences [43][45] or other hydrogeomorphic features [44]; Reach – the length of river occurring between breaks in channel slope caused by man-made dams or other hydrogeomorphic features [43];
281 282 283 284 285	Local Networklocal network – synonymous with Hydrologic Unit Code (HUC) 2 basins [41, 4243, 44]; Segmentsegment – the portion of a river between two major tributary confluences [43][45] or other hydrogeomorphic features [44]; Reach – the length of river occurring between breaks in channel slope caused by man made dams or other hydrogeomorphic features [43]; Patch[46]; reach - the length of river occurring between breaks in channel slope caused by man-
281 282 283 284 285 286	Local Networklocal network – synonymous with Hydrologic Unit Code (HUC) 2 basins [41, 4243, 44]; Segmentsegment – the portion of a river between two major tributary confluences [43][45] or other hydrogeomorphic features [44]; Reach – the length of river occurring between breaks in channel slope caused by man-made dams or other hydrogeomorphic features [43]; Patch[46]; reach - the length of river occurring between breaks in channel slope caused by man-made dams or other hydrogeomorphic features [43]; Patch[46]; reach - the length of river occurring between breaks in channel slope caused by man-made dams or other hydrogeomorphic features [45]; patch – an area used by an organism (e.g.,
281 282 283 284 285 286 287	Local Networklocal network – synonymous with Hydrologic Unit Code (HUC) 2 basins [41, 4243, 44]; Segmentsegment – the portion of a river between two major tributary confluences [43][45] or other hydrogeomorphic features [44]; Reach – the length of river occurring between breaks in channel slope caused by man-made dams or other hydrogeomorphic features [43]; Patch[46]; reach - the length of river occurring between breaks in channel slope caused by man- made dams or other hydrogeomorphic features [45]; patch – an area used by an organism (e.g., for reproduction or resource attainment) that can vary both spatially and temporally depending

291	decadal. The temporal units were used to denote both the scale of inferences needed to support
292	the management goal and the scale at which data should be collected to inform the inferences.

293 Case Studies

294	We applied the CM to four case studies. For each, we followed the pathways of stress
295	from Tier 1 EECs to the biological endpoint associated with the management goal to identify
296	information needs. We then characterized the spatial and temporal scales of the management
297	goal, the scientific inferences needed to inform the management goal, and that data collection
298	needs to occur to support the inferences for monitoring information needs identified as requiring
299	additional data for each of the case studies. To summarize similarities across case studies, we
300	generalized the stressors and inter- tier interactions identified in the case studies and then
301	summarize the similarities by EEC tier. More context for the river systems characterized in the
302	case studies can be found in Appendix S1.
303	South Canadian River
303 304	South Canadian River The South Canadian River is the mainstem river of the Canadian River basin (south-
304	The South Canadian River is the mainstem river of the Canadian River basin (south-
304 305	The South Canadian River is the mainstem river of the Canadian River basin (south- central U.S., Fig 1). The basin occupies a significant west-east climate gradient with
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304 305 306 307	The South Canadian River is the mainstem river of the Canadian River basin (south- eentral U.S., Fig 1). The basin occupies a significant west-east climate gradient with precipitation ranging from 40 to 145 cm/yr [47]. Land use in the basin is primarily cropland or pasture that transitions to the urban area near Oklahoma City. Three major reservoirs on the
304 305 306 307 308	The South Canadian River is the mainstem river of the Canadian River basin (south- central U.S., Fig 1). The basin occupies a significant west-east climate gradient with precipitation ranging from 40 to 145 cm/yr [47]. Land use in the basin is primarily cropland or pasture that transitions to the urban area near Oklahoma City. Three major reservoirs on the South Canadian River have significantly altered the river flow patterns, particularly the

311	particularly in the upper basin, has severed groundwater connections in many areas transforming
312	the fish assemblage [49]. Further, groundwater pumping in the alluvial aquifer has decreased
313	stream baseflows within this region [50, 51]. The rivers of the Great Plains are characterized
314	by extreme physicochemical conditions (i.e., water temperatures 4-40°C; salinities greater than
315	ocean water) and extensive flooding and extended periods of drought, and native fishes are well
316	adapted to those extremes. Pelagic broadcast spawning minnows (i.e., pelagophils) belong to a
317	reproductive guild of diminutive fishes that are emblematic of these stream systems. Of the
318	approximately 20 species in this reproductive guild, 13 are of conservation concern; the status of
319	the remaining seven species is poorly understood [18, 52]. Native populations of the federally-
320	threatened Arkansas River Shiner are believed to be restricted to two fragmented portions of the
321	South Canadian River [5349].

322 The Arkansas River Shiner is hypothesized to be affected by several anthropogenic 323 activities that primarily affect water quality and quantity (Fig 4). Three reservoirs on the South 324 Canadian River have altered discharge patterns (Fig S1), and fragmented river habitats. Two known native populations of Arkansas River Shiner occupy the two remaining river segments of 325 sufficient length and complexity to allow eggs to drift the time required to successfully complete 326 their early life history. Small impoundments for agriculture use, road crossings, groundwater 327 pumping and other local water extractions (e.g., oil and gas) threaten to further fragment existing 328 329 habitat. Fragmentation could also be problematic for upstream fish migrations; there is some evidence that Arkansas River Shiners migrate upstream to spawn to achieve adequate drift 330 distances for their offspring [54].[50]. It has also been speculated that this species might benefit 331 332 from access to floodplain habitats [55][51], but we are unaware of efforts to examine that hypothesis. Changes in the flow patterns may also relate to the expansion of salt cedar Tamarix 333

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334	spp. and other non-native riparian species that constrain the channel and inhibit channel habitat
335	complexity [56, 5752, 53]. Changes to the riparian corridor can also alter the availability of
336	drifting invertebrates available for Arkansas River Shiner feeding (i.e., Coleoptera,
337	Hymenoptera; [58]).[54]). Channel complexity acts to slow the transport of eggs [5349] and may
338	prevent eggs from being washed into downstream reservoirs where survival is hypothesized to be
339	extremely low. Climate change is expected to increase the intensity and frequency of drought
340	events within this region [59, 6055, 56], which may exacerbate habitat fragmentation, promote
341	ATVall-terrain vehicle traffic within the river channel causing direct mortality on stranded fish
342	(Gene Wilde, Texas Tech University, Personal Comm.), and concentrate contaminants and
343	salinity [6157]. The tolerances of Arkansas River Shiner to salinity concentrations and many
344	other contaminants are unknown (see Table 3 <u>S1;</u> [18]). Lastly, introductions of non-native fishes
345	via bait buckets have occurred within the basin. The primary concern has related tois the
346	presence of Red River Shiner (Notropis bairdi) because it is suspected to reproduce in a similar
347	manner and be a possible competitor to the Arkansas River Shiner [6157].
348	Fig 4. Conceptual model describing the relationship of natural and anthropogenic drivers to
349	essential ecosystem characteristics (EECs) affecting the recruitment of the Arkansas River
350	Shiner in the South Canadian River in New Mexico, Texas, and Oklahoma. Essential ecosystem
351	characteristics are groupings of ecosystem components. Tier 1 EECs represent physical and
352	chemical effects; fundamental measures of process that are directly affected by anthropogenic
353	and natural drivers. Tier 2 EECs represent a broad habitat category that is intended to encompass
354	the physical, chemical, and biological components of the riverine habitats that influence
355	reproduction, growth, and survival of biotic communities. The Tier 3 EECs represent

356	components of the biological systems that respond to changes in the hierarchical components of
357	the conceptual model. The strength of our understanding of how natural and anthropogenic
358	drivers interact with habitats, biological systems, and fish in large rivers is represented by the
359	different types of lines in the figure. Solid blue lines depict a strong understanding of the
360	relationship, the dotted-dashed blue line represents a moderate understanding of the relationship,
361	and the black dashed line represents a weak understanding of the relationship. The different
362	types of lines also represent the strength of our understanding of within EEC-tier relationships.

363 The results of the CM exercise that characterized factors affecting the Arkansas River Shiner in the South Canadian River suggested the critical life-history bottlenecks for the 364 Arkansas River Shiner are successful spawning and recruitment to the first year. Impediments 365 that limit our understanding of factors that lead to successful spawning and recruitment included 366 the effects of channel morphology and hydraulics on the quality and quantity of larval rearing 367 368 habitat, and subsequent effects on larval production (Table S1). Water use and other drivers 369 occurring at relatively coarse spatial and temporal scales are the hypothesized drivers related to 370 degradation of reproductive habitat for the Arkansas River Shiner (Fig 5). A temporal lag in responses at finer scales (i.e., improved habitat) would be anticipated with management actions 371 372 at these coarser spatial scales (i-e.g., water releases from dams); though, providing connectivity via minimal water releases would occur relatively quickly. Although there are gages on the 373 374 South Canadian River, the spacing of the gages is not sufficient to have a full understanding of flow patterns between the gages given the semi-arid nature of the basin and potential for reaches 375 to be affected by water withdrawals such as groundwater pumping. Our understanding of the 376 species life history is well established; however, the effects of human pressures on the species 377

378	and associated habitat has not been well studied (i.e., production, survival). As indicated in Fig 4,
379	the only stress pathway that was characterized as being well understood was from land use
380	stressors to Tier 1 processes. As those As stressors propagate through Tier 1 to Tier 2 and Tier 3,
381	the level of uncertainty increases increased such that it is not possible to define a preferred
382	hypothesis for Arkansas River Shiner recruitment failure. The status of much existing
383	information wasneeded to understand the hypothesized stress pathways and interactions was
384	mostly characterized as insufficient or not available (Table S1).
385	Fig 5. The spatial and temporal scales of the management goal, the scientific inferences needed
386	to inform the management goal, and that data collection needs to occur to support the inferences
387	for monitoring information needs identified as requiring additional data in the case study
388	addressing the recruitment of the Arkansas River Shiner in the Canadian River, Oklahoma (see
389	Table S1 for additional detail). A:Tier 1 EEC= Channel
390	morphology/Hydraulieshydraulics; Stressor=altered hydraulic regime; B:Tier 1
391	EEC=Biogeochemistry/Thermodynamicsbiogeochemistry/thermodynamics; Stressor=altered
392	water temperature regime; C:Tier 1
393	EEC=Biogeochemistry/Thermodynamicsbiogeochemistry/thermodynamics; Stressor=altered
394	biogeochemical regime; D:Tier 2 EEC=Arkansas River Shiner spawning habitat;
395	Stressors=contaminants, water temperature, habitat fragmentation; E:Tier 2 EEC=larval
396	Arkansas River Shiner habitat, Stressors=water temperature, habitat fragmentation and Tier 2
397	EEC=invertebrate habitat, Stressors=altered riparian plant community, discharge, sediment
398	deposition; F:Tier 3 EEC=primary production, Stressor = nutrient flux; G:Tier 3
399	EEC=invertebrate production; Stressor=invertebrate habitat quantity and quality and Tier 3

400	EEC=Arkansas River Shiner larvae production, Stressor=predation by invasive species; H:Tier 1
401	EEC= Biogeochemistry/Thermodynamicsbiogeochemistry/thermodynamics; Inter-tier
402	interaction= Sediment adsorption of contaminants and nutrients and Tier 3 EEC=Arkansas River
403	Shiner larvae production, Stressors=Arkansas River Shiner larvae habitat quantity and quality
404	and Tier 3 EEC=Arkansas River Shiner age-0 recruitment; Inter-tier interaction=Arkansas River
405	Shiner larvae mortality and Tier 3 EEC=Arkansas River Shiner age-1+ recruitment, Inter-tier
406	interaction= Arkansas River Shiner age-0 mortality and Tier 3 EEC=all, Inter-tier
407	interaction=trophic level interactions; I:Arkansas River Shiner larvae production; Stressor=
408	direct mortality from recreational use (i.e., ATVall-terrain vehicle and in-river traffic); J:Tier 3
409	EEC=Arkansas River Shiner egg quality and production; Stressor=Arkansas River Shiner
410	spawning habitat quantity and quality; K:Tier 3 EEC=Larval Arkansas River Shiner larvae
411	production; Inter-tier interaction=Arkansas River Shiner egg mortality.

412 Colorado River

413 The Colorado River flows 2,330 km from its origin in Colorado to its confluence with the 414 Gulf of California in Mexico draining parts of seven U.S. and two Mexican states; our 415 assessment focused on the portion of the Colorado River that flows through the Grand Canyon in 416 Arizona (Fig 1). Water storage associated with the allocation of water in the Colorado River for 417 human consumption and hydroelectric power generation has altered the hydrologic regime (Fig S2) in the Colorado River between Glen Canyon Dam and Lake Mead. Most fish species native 418 to the Colorado River have declined in abundance and distribution while numerous non-native 419 420 species have become established [62]. The Humpback Chub (Gila cypha), a fish native to the

421 Colorado River, was listed as endangered by The Humpback Chub (*Gila cypha*), a fish native to
422 the Colorado River, was listed as endangered by the U.S. Fish and Wildlife Service in 1967 and
423 given full protection under the Endangered Species Act of 1973 (ESA). To mitigate the effects of
424 anthropogenic changes to the river on Humpback Chub, an understanding of the mechanisms by
425 which Glen Canyon Dam and non-native species affect Humpback Chub is needed.

426 A critical life-history bottleneck for Humpback Chub is recruitment into the first year 427 class-(Fig 6). Temperature, light, and seasonal high river discharge from snowmelt are thought 428 to cue spawning behavior [63].[58]. Hydropower development has dampened the range of river 429 discharges of the Lower Colorado River within Grand Canyon. Historically, river discharge 430 varied between 15 and 3400 m³/s, however discharge was greater in large flood events; current dam operations limit flows to a range of 140 to 1000 m³/s (Fig S2). Resulting changes in 431 432 turbidity and water temperature create risks to endangered Humpback Chub, and other endemic fish. For example, the quantity and quality of habitat is reduced through changes in turbidity, 433 434 biogeochemistry, and the temperature regime. Hypolimnetic water releases from Lake Powell 435 maintain cold temperatures in the Colorado River downstream from Glen Canyon Dam; currently, spawning is limited to a single tributary, the Little Colorado River. As embryos 436 437 survive into the larval stage, nursery habitats to support growth and foraging are essential [64-438 6659-61]. A secondary risk to juvenile survival post-larval stage is piscivory from predation by 439 non-native species including Rainbow Trout (Oncorhynchus mykiss) and Channel Catfish 440 (Ictalurus punctatus) [67].[62]. Temperatures for well over 100 km downstream of Lake Powell are excellent for non-native, cold water species, including a closely managed world-class 441 Rainbow Trout fishery at Lees Ferry. Rainbow and Brown Trout (Salma trutta) are currently 442 managed as an invasive species downstream of the confluence of the Colorado and Little 443

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444	Colorado River, approximately 97 km downstream of Glen Canyon Dam, to mitigate predation
445	upon native fishes including the endangered Humpback Chub.

446	The CM exercise documented a generally high understanding of the relationships of
447	anthropogenic drivers to Tier 1 EECs, cascading to multiple, less understood hypotheses about
448	how these factors would combine to affect habitats at Tier 2 (Fig 6). High confidence in the
449	linkages from Tier 2 invertebrate habitat to Tier 3 insect population is followed by a lesser
450	understanding of how insect populations are linked to larval production. The uncertainty of the
451	relations between food resources stands in contrast to high certainty that was ascribed to the
452	linkages from larval chub habitat to larval chub production, and from spawning habitat to larval
453	chub production and thence to recruitment. Since 1997, the Glen Canyon Dam Adaptive
454	Management Program has supported extensive monitoring and research across the spatial and
455	temporal landscape of the Colorado River. As a result, the information needed to characterize
456	some of the stressors is readily available (Table S2). However, the status of some existing
457	information was characterized as insufficient or not available. For the information needs
458	characterized as being insufficient or not available, we identified the spatial and temporal scales
459	at which data collection would facilitate the inferences needed to inform the management goal
460	(Fig-7)Understanding how these knowledge gaps affect Humpback Chub recruitment could
461	identify strategies that will help achieve the management goal of improving Humpback Chub
462	recruitment in the Colorado River.

Fig 6. Conceptual model describing the relationship of anthropogenic drivers to essential
ecosystem characteristics (EECs) affecting the recruitment of Humpback Chub in the Colorado

465	River between Gien Canyon Dam and Lake Mead, Arizona. Essential ecosystem characteristics
466	(EECs) are groupings of ecosystem components. Tier 1 EECs represent physical and chemical
467	effects; fundamental measures of process that are directly affected by anthropogenic and natural
468	drivers. Tier 2 EECs represent a broad habitat category that is intended to encompass the
469	physical, chemical, and biological components of the riverine habitats that influence
470	reproduction, growth, and survival of biotic communities. The Tier 3 EECs represent
471	components of the biological systems that respond to changes in the hierarchical components of
472	the conceptual model. The strength of our understanding of how natural and anthropogenic
473	drivers interact with habitats, biological systems, and fish in large rivers is represented by the
474	different types of lines in the figure. Solid blue lines depict a strong understanding of the
475	relationship, the dotted-dashed blue line represents a moderate understanding of the relationship,
476	and the black dashed line represents a weak understanding of the relationship. The different
477	types of lines also represent the strength of our understanding of within EEC-tier relationships.
1.	
478	Fig 7. The CM exercise documented a high understanding of the relationships of
479	anthropogenic drivers to Tier 1 EECs, cascading to multiple, less-understood hypotheses about
480	how these factors would combine to affect habitats at Tier 2 (Fig 6). High confidence in the
481	linkages from Tier 2 invertebrate habitat to Tier 3 insect production is followed by a lesser
482	understanding of how insect production is linked to larval production. The uncertainty of the
483	relations between food resources stands in contrast to high certainty that was ascribed to the
484	linkages from larval chub habitat to larval chub production, and from spawning habitat to larval
485	chub production and then to recruitment. Since 1997, the Glen Canyon Dam Adaptive
486	Management Program has supported extensive monitoring and research across the spatial and

487	temporal landscape of the Colorado River. As a result, the information needed to characterize
488	some of the stressors is readily available (Table S2). However, the status of some existing
489	information was characterized as insufficient or not available. For the information needs
490	characterized as being insufficient or not available, we identified the spatial and temporal scales
491	at which data collection would facilitate the inferences needed to inform the management goal
492	(Fig S3). Understanding how these knowledge gaps affect Humpback Chub recruitment could
493	identify strategies that will help achieve the management goal of improving Humpback Chub
494	recruitment in the Colorado River.
495	The spatial and temporal scales of the management goal, the scientific inferences needed to
496	inform the management goal, and that data collection needs to occur to support the inferences for
497	monitoring information needs identified as requiring additional data in the case study addressing
498	Humpback Chub recruitment in the Colorado River between Glen Canyon Dam and Lake Mead,
499	Arizona (see Table S2 for additional detail). A:Tier 2 EEC - larval Humpback Chub habitat;
500	Stressors=habitat fragmentation, turbidity and Tier 2 EEC= Humpback Chub spawning habitat;
501	Stressor-water temperature and Tier 3 EEC-insect production; Stressor-benthie
502	macroinvertebrate habitat quantity and quality and Tier 3 EECs-all; Inter tier interaction-trophic
503	level interactions; B:Tier 3 EEC-larval Humpback Chub production; Stressor-larval Humpback
504	ehub habitat quantity and quality; C: Tier 3 EEC-Humpback Chub egg quality and production;
505	Stressors- Humpback Chub spawning habitat quantity and quality; D:Larval Humpback Chub
506	production; Inter tier interaction-mortality of Humpback Chub eggs; E:Tier 3 EEC- larval
507	Humpback Chub production; Stressor-predation by invasive species and Tier 3 EEC-Primary
508	production; Stressor=nutrient flux; F:Tier 1 EEC=Biogeochemistry/thermodynamics; Inter-tier
1	

509	interaction=sediment adsorption of contaminants and nutrients; G:Her 3 EEC=Age-0 Humpback	
510	Chub recruitment; Inter-tier interaction=mortality of larval Humpback Chub	
511	Columbia River	
512	White Sturgeon is the largest freshwater fish in North America [63]. Like other sturgeon	
513	species, anthropogenic stressors have negatively affected White Sturgeon productivity. Our	
514	knowledge of factors affecting White Sturgeon productivity remain poorly understood [40].	
515	Therefore, we used the CM to identify knowledge gaps associated with the hypotheses that dam	
516	construction and operation, land-use practices, and invasive species, in some combination, affect	
517	the recruitment of age-0 White Sturgeon (Fig 7). Within the basin, development of hydroelectric	
518	and water-storage dams have changed the magnitude and seasonality of the natural river	
519	discharge (Fig S4) and thermal regimes [64], reduced the quantity and quality of spawning	
520	habitats [65, 66], and disrupted historical migration patterns [67]. Prior to hydropower	
521	development, White Sturgeon experienced a hydrograph that peaked during June-July due to	
522	snowmelt [64]. However, from 1949 to 1993 the average discharge in June decreased from	
523	14,000 m ³ /s to 6,000 m ³ /s and the maximum water temperature has increased by 1.8°C [64].	
524	White Sturgeon likely used the natural hydrograph and thermal regime as cues to seek out	
525	optimal spawning habitats and initiate spawning [65, 68].	
536		
526	Fig 7. Conceptual model describing the relationship of anthropogenic drivers to essential	Formatted: Add space between paragraphs of the same style
527	ecosystem characteristics (EECs) affecting the recruitment of White Sturgeon in the Columbia	
528	River, U.S. Essential ecosystem characteristics (EECs) are groupings of ecosystem components.	
529	Tier 1 EECs represent physical and chemical effects; fundamental measures of process that are	
1		

530	directly affected by anthropogenic and natural drivers. Tier 2 EECs represent a broad habitat
531	category that is intended to encompass the physical, chemical, and biological components of the
532	riverine habitats that influence reproduction, growth, and survival of biotic communities. The
533	Tier 3 EECs represent components of the biological systems that respond to changes in the
534	hierarchical components of the conceptual model. The strength of our understanding of how
535	natural and anthropogenic drivers interact with habitats, biological systems, and fish in large
536	rivers is represented by the different types of lines in the figure. Solid blue lines depict a strong
537	understanding of the relationship, the dotted-dashed blue line represents a moderate
538	understanding of the relationship, and the black dashed line represents a weak understanding of
539	the relationship. The different types of lines also represent the strength of our understanding of
540	within EEC-tier relationships.
541	Factors other than river discharge and water temperature may also be affecting age-0
542	White Sturgeon recruitment [40]. In areas of the Columbia and Snake Rivers with hydropower
543	development, White Sturgeon populations are functionally isolated by dams. Consequently,
544	White Sturgeon depend on conditions within restricted reaches to sustain production. In some
545	reaches, suitable rearing habitat exists, and individual growth rates are high, but spawning habitat
546	is limited and recruitment of fish is poor [65]. In other reaches, favorable spawning conditions
547	exist but growth of young fish may be density limited [69]. How the availability of food
548	resources for larval and juvenile White Sturgeon varies among reservoirs may affect age-0 White
549	Sturgeon recruitment. Research has also suggested that contaminants may affect White Sturgeon
550	reproductive biology The Columbia River is the fourth largest river by volume in the United
1	
551	States [68] and drains a basin of 671,000 km ² that includes parts of seven states, land ceded to 14

552	groups of affiliated tribes in the U.S. portion of the Columbia River Basin, land ceded to three
553	tribal groups, known as First Nations, in the Canadian portion of the basin, and one Canadian
554	province ([69]; Fig 1). The Columbia River supports anadromous fish species such as Chinook
555	(Oncorhynchus tshawytscha), Coho (O. kisutch), Chum (O. keta), and Sockeye Salmon (O.
556	nerka), including 12 populations of four species of salmon and steelhead (O. mykiss) listed as
557	threatened or endangered under the ESA. The Columbia River also contains a complement of
558	resident native and non-native fishes. Among the resident native fish species are White Sturgeon,
559	the largest freshwater fish in North America [70]. Like other sturgeon species, anthropogenie
560	stressors have negatively affected White Sturgeon productivity.
561	Our knowledge of factors affecting White Sturgeon productivity remain poorly
562	understood [38]. Therefore, we used the CM to identify knowledge gaps associated with the
563	hypotheses that dam construction and operation, land-use practices, and invasive species, in
564	some combination, affect the recruitment of age-0 White Sturgeon. Within the basin,
565	development of hydroelectric and water storage dams have changed the magnitude and
566	seasonality of the natural river discharge (Fig S3) and thermal regimes [71], reduced the quantity
567	and quality of spawning habitats [72, 73], and disrupted historical migration patterns [74]. Prior
568	to hydropower development, White Sturgeon experienced a hydrograph that peaked during June-
569	July due to snowmelt [71]. However, from 1949 to 1993 the average discharge in June decreased
570	from 14,000 m ³ /s to 6,000 m ³ /s and the maximum water temperature has increased by 1.8°C
571	[71]. White Sturgeon likely used the natural hydrograph and thermal regime as cues to seek out
572	optimal spawning habitats and initiate spawning [72, 75].
573	Factors other than river discharge and water temperature may also be affecting age 0
574	White Sturgeon recruitment [38]In areas of the Columbia and Snake Rivers with hydropower 27

575	development, White Sturgeon populations are functionally isolated by dams. Consequently,
576	White Sturgeon depend on conditions within restricted reaches to sustain production-In some
577	reaches, suitable rearing habitat exists, and individual growth rates are high, but spawning habitat
578	is limited and recruitment of fish is poor [72]. In other reaches, favorable spawning conditions
579	exist but growth of young fish may be density limited [76]. How the availability of food
580	resources for larval and juvenile White Sturgeon varies among reservoirs may affect age-0 White
581	Sturgeon recruitment. Research has also suggested that contaminants may affect White Sturgeon
582	reproductive biology [7770]. The introduction of non-native fishes has clearly affected the native
583	fish assemblage in the Columbia River [32, 7871]. Channel Catfish, Smallmouth Bass
584	Micropterus dolomieu, and Walleye (Sander vitreus) that have been introduced into the
585	Columbia River have all been shown to prey upon or compete with native fish species [78, 7971,
586	72] and may also affect White Sturgeon.
587	The CM (Fig <u>87</u>) provides structure to the multiple competing hypotheses and indicates
588	how anthropogenic drivers may be affecting Tier 1, 2, and 3 EEC's. We hypothesized pathways
589	of stress arising from anthropogenic drivers that affect age-0 White Sturgeon recruitment (Fig 8).
590	Not surprisingly, the strength of our understanding of the effects of stressors on White Sturgeon
591	was greater for relationships between life stages that are more easily sampled (e.g., adults) and
592	that rely on data that are readily accessible as part of monitoring associated with hydropower
593	development (e.g., discharge and water temperature) or metrics that are from combinations of
594	these variables (e.g., estimates of White Sturgeon spawning habitat; Table S3). However, for
595	relationships between harder to sample White Sturgeon life stages (e.g., larvae), biota that
596	require expertise and equipment atypical of traditional fisheries assessments in large rivers (e.g.,

598	modeling expertise (e.g., sediment transport dynamics), the existing information was insufficient
599	or not available. For example, we identified the need to better understand the effects of channel
600	morphology and hydraulics on benthic macroinvertebrate habitat, invertebrate production, and
601	subsequent larval White Sturgeon production. The most certain pathways connected changes in
602	hydrology, hydraulics, and temperature regimes to reduced spawning habitat in Tier 2, then to
603	decreased egg quality and production at Tier 3.
604	Fig 8Conceptual model describing the relationship of anthropogenic drivers to essential
605	ecosystem characteristics (EECs) affecting the recruitment of White Sturgeon in the Columbia
606	River, U.S. Essential ecosystem characteristics (EECs) are groupings of ecosystem components.
607	Tier 1-EECs represent physical and chemical effects; fundamental measures of process that are
608	directly affected by anthropogenic and natural drivers. Tier 2 EECs represent a broad habitat
609	eategory that is intended to encompass the physical, ehemical, and biological components of the
610	riverine habitats that influence reproduction, growth and survival of biotic communities. The
611	Tier 3 EECs represent components of the biological systems that respond to changes in the
612	hierarchical components of the conceptual model. The strength of our understanding of how
613	natural and anthropogenic drivers interact with habitats, biological systems, and fish in large
614	rivers is represented by the different types of lines in the figure. Solid blue lines depict a strong
615	understanding of the relationship, the dotted-dashed blue line represents a moderate
616	understanding of the relationship, and the black dashed line represents a weak understanding of
617	the relationship. The different types of lines also represent the strength of our understanding of
618	within EEC tier relationships.

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619	Our results suggest there are stressors that can affect the management goal of increasing
620	age-0 White Sturgeon recruitment that are poorly understood and that could confound efforts to
621	manage White Sturgeon in the Columbia River. Our characterization of the spatial and temporal
622	scales that data should be collected at could help guide future efforts to fill data gaps to support
623	the inferences needed to address the goal of improving recruitment of age-0 White Sturgeon (Fig
624	9 <u>S5</u>).

Fig 9. The spatial and temporal scales of the management goal, the scientific inferences needed 625 626 to inform the management goal, and that data collection needs to occur to support the inferences 627 for monitoring information needs identified as requiring additional data in the case study addressing White Sturgeon recruitment in the Columbia River (see Table S3 for additional 628 detail). A:Tier 1 EEC= Biogeochemistry/thermodynamic; Stressor= altered biogeochemical 629 regime and Tier 2 EEC= benthic macroinvertebrate habitat; Stressors=channel stability, sediment 630 631 deposition, fragmentation and Tier 2 EEC=Larval White Sturgeon habitat; Stressors=habitat 632 fragmentation sediment deposition, water temperature and Tier 2 EEC=White Sturgeon 633 spawning habitat; Stressors=contaminants, sediment deposition and Tier 3 EEC=White Sturgeon 634 egg quality and production; Stressor=predation by invasive species and Tier 3 EEC=larval White Sturgeon production; Stressor=predation by invasive species and Tier 3 EEC=benthic 635 macroinvertebrate production; Stressor=benthic macroinvertebrate habitat quantity and quality; 636 637 B:Tier 1 EEC=sediment transport; Stressors=altered sediment regime, altered hydraulic regime and Tier 1 EECs=channel morphology/hydraulics, sediment transport; Stressor=altered hydraulic 638 regime and Tier 3 EEC=primary production; Stressor=nutrient fluxes; C:Tier 1 EEC=Channel 639 morphology/Hydraulics, Sediment transport; Stressor=altered hydraulic regime and Tier 1 EEC 640

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641	= channel morphology/hydraulics, sediment transport; Inter-tier interaction=sediment transport
642	dynamics; D:Tier 1 EEC=Channel morphology/Hydraulics, Sediment transport; Stressor=altered
643	hydraulic regime and Tier 1 EEC=Channel morphology/Hydraulics, Sediment transport; Inter-
644	tier interaction=sediment transport dynamics; E:Tier 1 EEC=Biogeochemistry/Thermodynamics;
645	Inter-tier interaction=sediment adsorption of contaminants and nutrients; F: Tier 3 EEC=larval
646	White Sturgeon production; Stressors=larval White Sturgeon habitat quantity and quality and
647	Tier 3 EEC=larval White Sturgeon production, age 0 White Sturgeon recruitment; Inter-tier
648	interaction=mortality and White Sturgeon egg quality and production; Inter-tier
649	interaction=predation of White Sturgeon eggs by native fish and Tier 3 EECs=larval White
650	Sturgeon production; Inter-tier interaction=predation of White Sturgeon larvae by native fish and
651	Tier 3 EECs=all; Inter-tier interactions=trophic level interactions.

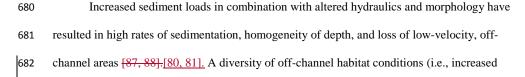
652 Upper Mississippi and Illinois Rivers

653 In the agricultural Midwest, basin-wide land uses affect the delivery of sediments, nutrients, and runoff to the Upper Mississippi and Illinois rivers [80-8373-76]. Within the 654 floodplain of these two large rivers, agriculture and residential land uses often rely upon the use 655 656 of levees to isolate productive or developed lands during seasonal high-flow events. Within the channel, these rivers support commercial navigation with locks and dams and river-training 657 structures, which have dramatically altered channel morphology and hydraulics throughout the 658 system. Together, the cumulative effects of these modifications to the basin, floodplain, and river 659 660 have implications for habitat diversity and native fish biodiversity [84]-[77]. Additionally, recent invasion and expansion of non-native species, namely Silver Carp (Hypophthalmichthys 661

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molitrix) and Bighead Carp (*H. nobilis*), have direct and indirect effects on native fishes that
likely compound or confound stress pathways on native fish biodiversity [85, 8678, 79].
Therefore, we used the CM to explore how these primary anthropogenic drivers have likely
influenced fish habitats and associated life stages (Fig 108).

Fig 108. Conceptual model of how anthropogenic drivers in the upper Mississippi and Illinois 666 Rivers influence native fish habitats and recruitment. Essential ecosystem characteristics (EECs) 667 are groupings of ecosystem components. Tier 1 EECs represent physical and chemical effects; 668 fundamental measures of process that are directly affected by anthropogenic and natural drivers. 669 670 Tier 2 EECs represent a broad habitat category that is intended to encompass the physical, 671 chemical, and biological components of the riverine habitats that influence reproduction, growth, and survival of biotic communities. The Tier 3 EECs represent components of the biological 672 systems that respond to changes in the hierarchical components of the conceptual model. The 673 674 strength of our understanding of the relationships of how natural and anthropogenic drivers 675 interact with habitats, biological systems, and fish in large rivers is represented by the different 676 types of lines in the figure. Solid blue lines depict a strong understanding of the relationship, the dotted-dashed blue line represents a moderate understanding of the relationship, and the black 677 dashed line represents a weak understanding of the relationship. The different types of lines also 678 679 represent the strength of our understanding of within EEC-tier relationships.



683 residence time, low velocity, warm temperatures, availability of food resources) support growth and development of larval and juvenile fishes [89, 9082, 83] and often provide important food 684 resources for adult fishes [91-9584-88]. Further, deep, low-velocity off-channel habitats are 685 recognized as important refugia for a wide range of fishes during high-flow events and seasonal 686 periods of low temperatures [96-9989-92]. Loss of floodplain connectivity has eliminated the 687 seasonal exchange of nutrients, organisms and organic matter between river and floodplain 688 689 environments that support biological diversity and productivity [100, 101].[93, 94]. Reduced availability of spawning, nursery, foraging, or overwintering habitat conditions can serve as 690 691 bottlenecks to fish populations through limited larval production, reduced growth, and increased 692 overwinter mortality. For example, high sedimentation rates have been filling backwaters in the Illinois River for decades, thus limiting the availability of overwintering conditions for fishes 693 694 that bioenergetically need a deep refuge with slow water velocities. Missing year-classes in this reach, represented by truncated size structure in the Largemouth Bass (Micropterus salmoides) 695 696 population are hypothesized to be a result of periodic winter mortality (Fig <u>\$4\$6</u>). 697 The application of our CM makes clear that while the general effects of anthropogenic drivers on hydrology, sediment transport, biogeochemistry and hydraulics and morphology are 698 699 well understood, there is much less known about how those effects influence the quality and 700 availability of required habitat conditions (Tier 2, Fig 108). Although there is likely overlap of 701 habitat requirements among species with similar life histories, the diversity of habitat conditions

necessary to support a native and diverse fish community has not been explored. Consequently,

the existing information needed to assess the relationship between habitat quality and quantity,

and egg production, juvenile recruitment, and adult survival of fish populations within the Upper

705 Mississippi and Illinois rivers was categorized as insufficient to not available (Tier 3 Inter-tier 33

706	interaction, Fig 108; Table S4). Addressing these knowledge gaps could improve the
707	effectiveness of habitat restoration efforts focused on maintaining a diverse native fish
708	community. The spatial and temporal scales of data collection that would support needed
709	inferences to address restoring and maintaining native fish biodiversity and habitat quantity and
710	quality are characterized in Fig 1187.

711	Fig 11. The spatial and temporal seales of the management goal, the seientific inferences needed
712	to inform the management goal, and that data collection needs to occur to support the inferences
713	for monitoring information needs identified as requiring additional data in the case study
714	addressing native fish biodiversity and habitat diversity in the Mississippi and Illinois rivers (see
715	Table S4 for additional detail). A:Tier 1 EEC=Sediment transport; Stressor=altered hydraulic
716	regime and Tier 1 EEC=Biogeochemistry/Thermodynamics; Stressor=altered biogeochemical
717	regime and Tier 1 EEC=Biogeochemistry/Thermodynamics; Inter tier interaction=sediment
718	adsorption of contaminants and nutrients; B: Tier 2 EEC=adult native fish overwintering habitat;
719	Stressors=water velocity, water temperature, dissolved oxygen, sediment deposition and Tier 2
720	EEC=Juvenile native fish habitat; Stressors=water depth, water velocity, water temperature,
721	dissolved oxygen, contaminants, sediment deposition and Tier 2 EEC=Native fish spawning
722	habitat; Stressors=water depth, water velocity, habitat fragmentation, sediment deposition, water
723	temperature, dissolved oxygen, contaminants and Tier 3 EEC=Adult and juvenile native fish
724	recruitment; Inter-tier interaction=mortality and Tier 3 EEC=all; Stressors=invasive species and
725	Tier 3 EEC=all; Inter-tier interaction=trophic level interactions; C:Tier 1 EEC=Channel
726	morphology/Hydraulics; Inter-tier interaction=channel forming processes; D:Tier 1
727	EEC=Channel morphology/Hydraulics, Sediment transport; Inter-tier interaction=sediment
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728	transport dynamics; E:Tier 3 EEC=Adult native fish recruitment; Stressor=adult native fish
729	overwintering habitat quantity and quality and Tier 3 EEC=Juvenile native fish recruitment;
730	Stressor=juvenile native fish habitat quantity and quality and Tier 3 EEC=Native fish egg quality
731	and production; Stressor=spawning habitat quantity and quality.

732 Similarities across case studies

We observed similarities in the stressors and interactions within EEC tiers across the four 733 case studies. For Tier 1, an altered hydrologic regime was identified as a stressor to the 734 735 Hydrologyhydrology EEC in all four rivers (Table 1). Presumably this is due to the ubiquitous effects of dams on the systems examined. However, in some rivers the altered hydrologic regime 736 737 originated from other anthropogenic (e.g., water use, land use, biological community alteration) 738 and natural (e.g., Climateclimate) drivers. Similarly, all four case studies listed an altered water 739 temperature regime as a stressor to the 740 Biogeochemistry/Thermodynamicsbiogeochemistry/thermodynamics EEC with linkages to several anthropogenic drivers (Fig 4, 6-8; Tables S1-S4). There were also similarities across case 741 studies with respect to the identification of interactions between Tier 1 EECs with all four case 742

studies noting interactions between Tier 1 EEC components.

744	Table 1. Stressors or inter-tier interactions affecting Tier 1 Essential Ecosystem Characteristics (EEC) identified as an information
745	need in the application of the conceptual model to case studies in the Canadian River [54], Colorado River [54], Columbia River (3),
746	and Upper Mississippi and Illinois rivers (4).(1), Colorado River (2), Columbia River (3), and Upper Mississippi and Illinois rivers
747	(4). Tier 1 EECs are measurable characteristics that describe processes that can significantly alter the morphological or chemical
748	characteristics within a river channel.

	Tier 1 EEC			
Stressor or inter-tier interaction	Hydrology	Channel Morphology/Hydraulics	Sediment Transport	Biogeochemistry/ Thermodynamics
Altered Hydrologic Regime	1, 2, 3, 4	1	1, 2	
Altered Hydraulic Regime		1, 3, 4	3	
Altered Sediment Regime			3, 4	
Altered Water Temperature Regime				1, 2, 3, 4
Altered Biogeochemical Regime				1, 3, 4
Channel forming processes		1, 2, 3, 4		
Sediment transport dynamics		1, 2, 3, 4	1, 2, 3, 4	
Sediment adsorption of contaminants and nutrients				1, 2, 3, 4

751	For Tier 2, there were similarities across case studies; however, the adaptation and
752	elaboration of the components to the management goal in the case studies was apparent (Table
753	2). The management goal associated with the case study for the Upper Mississippi and Illinois
754	Rivers resulted in Tier 2 EEC components (e.g., overwintering habitat) and stressors (e.g.,
755	dissolved oxygen) that were unique. Spawning habitat was identified as a Tier 2 EEC component
756	in all the case studies and multiple stressors were identified as affecting this component in two or
757	more of the case studies. Larval fish and invertebrate habitat were noted as Tier 2 EEC
758	components with some similarities in stressors across case studies. Habitat fragmentation,
759	sediment deposition, and water temperature were listed as stressors to Tier 2 EECs in all four

case studies. No interactions between Tier 2 EEC components were listed for the case studies.

761 Table 2. Stressors affecting Tier 2 Essential Ecosystem Characteristics (EEC) identified in the application of the conceptual model to

- case studies in the Canadian River [54], Colorado River [54],(1), Colorado River (2), Columbia River (3), and Upper Mississippi and
- 763 Illinois rivers (4). Tier 2 EECs are broadly described as physical, chemical, or biological components of "habitat" that are
- 764 hypothesized to have overall fitness consequences.

	Tier 2 EEC (habitat)					
Stressor	Spawning	Overwintering	Juvenile fish	Larval fish	Invertebrate	
Altered riparian community					1	
Channel stability					2, 3	
Contaminants	1, 3, 4		4			
Discharge					1	
Dissolved oxygen	4	4	4			
Habitat fragmentation	1, 3, 4			1, 2, 3	2, 3	
Sediment deposition	3, 4	4	4	3	1, 2, 3	
Turbidity				2		
Water Depth	4		4			
Water temperature	1, 2, 3, 4	4	4	1, 3		
Water velocity	3, 4	4	4			

765

766	
767	The adaptionadaptation of Tier 3 EECs and elaboration of the biological system related to
768	the management goal addressed by the case studies resulted in EECs that were comprised of fish
769	life stages ranging from eggs to adult fish, primary and invertebrate production, invertebrates and
770	biodiversity (Table 3). All Tier 3 EEC components, except biodiversity, were present in the four
771	case studies. Not surprisingly, habitat quantity and quality were listed as stressors to all the EECs
772	related to fish and invertebrates. Six of eight stressors or inter-tier interactions were listed as
773	affecting fish larvae and five of eight were noted as affecting egg quantity and quality. In
774	contrast to Tier 2, interactions were extensively noted between Tier 3 EECs and trophic level
775	interspecific interactions were listed in all four case studies.

776 Table 3. Stressors or inter-tier interactions affecting Tier 3 Essential Ecosystem Characteristics (EEC) identified in the application of

the conceptual model to case studies in the Canadian River [54], Colorado River [54], (1), Colorado River (2), Columbia River (3), and

778 Upper Mississippi and Illinois rivers (4). Tier 3 EECs represent components of the hypothesized biological system upon which the

cascading effects of anthropogenic and natural drivers act, and interactions occur.

				Tier 3 EEC			
Stressor or inter-tier interaction	Adult fish recruitment	Juvenile fish recruitment	Larval fish production	Fish egg quality/ production	Invertebrate production	Primary production	Biodiversity
Direct mortality			1				
Predation/competition by invasive species			1, 2, 3	3			4
Habitat quantity/quality	4	4	1, 2, 3	1, 2, 3, 4	1, 2, 3		
Nutrient flux						1, 2, 3	
Trophic level interspecific interactions	1, 2, 3, 4	1, 2, 3, 4	1, 2, 3, 4	1, 2, 3, 4	1, 2, 3, 4	1, 2, 3, 4	4
Predation			3	3			
Mortality	1, 3, 4	1, 2, 3, 4	1, 2, 3				
Fish condition				1, 3, 4			

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782 Discussion

The main objective of our exploration of CMs was to impose some structure on the complex ecosystems found in large rivers and from that structure, identify gaps in monitoring information that could inform the management of fish. Comparison across our four case studies provides some insights into large rivers and the utility of the CM to identify gaps in our understanding of factors affecting fish in large rivers.

788 Despite large differences in the physical and ecological contexts of the river systems, the 789 case studies-also demonstrated substantial commonalities in the data needed to better understand how human activities affect these systems and in the application of the CM. The general tiered 790 791 structure of drivers and cascading responses through EECs worked well with the four examples. Each of the four rivers could be placed in the tiered CM to illustrate current perceptions about 792 drivers and responses. The hierarchical CM generally increased in complexity from top to 793 794 bottom. Among all rivers, there tended to be greater understanding of links from drivers to Tier 1 795 and Tier 2 EECs, and less understanding about linkages to Tier 3.

The strength of understanding of interactions between anthropogenic and natural drivers and EECs, and between and within EECs, varied considerably among river systems, however, resulting from both variable complexity and existing knowledge. For example, linkages from drivers to Tier 1 EECs were considered strong in the case of the Humpback Chub in the Grand Canyon, but between Tier 1 and Tier 2 only moderate. This is probably indicative of the substantial research investments in examination of physical processes in this river system

1	802	[10295]. A logical next step in the process of identifying and prioritizing data collection needs,
	803	once major hypothesized stressor response pathways have been described based on expert
	804	elicitation, is to conduct an extensive literature review to support or refute those pathways and
	805	the strength and understanding of the pathways

806 Although we did not prescribe a specific approach to the CM process, the case studies 807 employed similar strategies. Our modelling exercises started with the definition of a management 808 goal. In all our case studies, the management goals pertained to a desired biological endpoint represented in Tier 3. After the definition of the management goal, we conceptualized 809 810 interactions between drivers and EECs and between EECs with a combination of top-down and 811 bottom-up approaches. A top-to-bottom approach to working with these models is generally consistent with a management perspective wherein anthropogenic drivers that are most directly 812 813 managed in a large-river system (e.g., land and water use, etc.) cascade from top to bottom 814 through fluxes to physical and chemical habitats, and then to biological responses. While this is 815 generally true for anthropogenic drivers, a notable exception to the top-to-bottom management 816 approach would be that in the U.S., there are few actions currently directed at reducing emissions 817 affecting climate [103][96] which is a natural driver in our CM. Climate was hypothesized to be 818 a stressor in the case study application of the CM to Arkansas River Shiner management in the 819 South Canadian River and is hypothesized to be affecting hydrologic regimes elsewhere [104, 820 105].[97, 98], but was not specifically mentioned in other case studies. The CMs can readily be 821 modified to incorporate other factors or pathways (e.g., climate effects) as new information or 822 perspectives become available. A bottom to top approach is equally or more valuable as it starts with the foundation of understanding about the species or community, and then seeks to identify 823 824 which stressors affect population or community responses. A bottom-up approach can readily

825 identify information gaps in linkages from ecological processes to demographic parameters
826 [106].[99].

The top-to-bottom and bottom-up approaches meet in the middle in Tier 2 in the concept 827 828 of habitat: the resources and conditions present in an area that produce occupancy [107].[100]. 829 Tier 2 is critical as it has little value if it is not defined based on biological requirements or if 830 managers lack understanding on how habitat is formed. Among our examples, the Upper 831 Mississippi River is notable for asserting strong understanding of the linkages from land-use 832 stressors to sediment regime to diminished overwintering habitat for native adult fishes. After 833 that, interactions with other processes and life stages combine to increase uncertainty about 834 whether overwintering habitat is a limiting factor in biodiversity. In contrast, the high confidence in understanding how White Sturgeon egg quality and production are linked to spawning habitat 835 836 in the Columbia River Basin, provides a strong linkage upward through Tier 1 EECs and 837 potential management actions (Fig 87). Although at times elusive, the concept of habitat is 838 critical for linking management to biotic endpoints [108101]. 839 Large rivers are typically managed for multiple objectives, including fisheries, multispecies, or ecosystem objectives. Management decisions typically require an understanding of 840 841 how management actions propagate through a river ecosystem. Although the emphasis may be 842 on a biological endpoint (among other objectives), understanding the intermediate steps and the 843 processes linking them, and potential interactions between processes or EEC components, can 844 help formulate effective management strategies; especially as multiple objectives compete. In a multi-species context, the conceptual models can help identify commonalities and differences in 845 in how stressors propagate to biota and therefore provide a basis for prioritizing monitoring 846 efforts. In the case where species or guilds have similar habitat affinities and life histories, a 847

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848	dominant anthropogenic stress pathway may be hypothesized and focus on a single or few
849	monitoring components may be justified. An example may be multiple large-river species that
850	are known to be cued to spawn by spring flow pulses. In such a case, the characteristics of the
851	annual hydrograph would be a dominant physical monitoring variable and biological monitoring
852	could focus on reproductive success of one or more of the species. In the case where multiple
853	species of concern have different reproductive strategies – for example, rheophilic species like
854	sturgeon that may require in-channel dispersion of young to flowing habitats compared to
855	invasive carp whose young thrive when they can disperse to lentic floodplain pools – pathways
856	and monitoring strategies will diverge. In the latter case, it would probably not be sufficient to
857	monitor and assess the characteristics of the annual hydrograph; instead, hydrologic metrics
858	would need to be integrated with hydraulic and geomorphic metrics to assess where and when
859	the different habitats would be available and could be targeted for young-of-the-year sampling.
860	Effectively addressing multiple species would rely on detailed knowledge of life histories and
861	how they play out on the landscape - such information is missing for many species and may
862	need to be developed for effective design of monitoring and management actions.
863	Management actions intended to benefit fish in large rivers can directly or indirectly
864	affect multiple ecosystem components. Without consideration of the effects of management on
865	non-target ecosystem components, unintended consequences may limit management efficacy.
866	Hypothesizing inter-tier interactions in the Tier 3 EEC (e.g., see Fig 7), can provide insight on
867	the potential interactions among fish species and other biological components in the context of
868	the hierarchical CM. In all our case studies, the lumping of multiple biological interactions in
869	Tier 3 resulted in a simplification of complex trophic interactions. For example, as Tier 3
870	encompasses all biological responses, it includes multiple life stages of many interacting species 44

871	at varying trophic levels. Because of this, the four CMs diverged significantly at Tier 3 as
872	components were expanded to accommodate existing understanding. Even as the Tier 3
873	components were expanded in complexity, they remained highly simplified views of the
874	ecosystem. Simplification was based, in part, on the importance of key species in management
875	goals and the experts' existing knowledge. Even though the hypothesized Tier 3 interactions in
876	our case studies conveyed a simplification of the trophic interactions, the hypothesized
877	interactions do suggest the need for information that clarifies the trophic interactions and effects
878	of Tier 3 EEC components on the biological endpoint. If desired or warranted, the Tier 3 EEC
879	could be elaborated to capture more complexity. For example, in Fig 7, the Tier 3 inter-tier
880	interaction between anadromous and resident fishes and white sturgeon larvae could expanded to
881	include interactions with specific fish species. Monitoring can help clarify the effects of
882	management actions, including on non-target ecosystem components, but only if data are
883	collected to characterize key ecosystem processes that could affect the outcome. The process of
884	considering and elucidating Tier 3 EEC interactions can help identify the non-target ecosystem
885	components that could be affected when managing for a specific biological endpoint.
886	The CMs explored here also provide a framework for considering return on science
887	investments. The CMs explored here also provide a framework for considering return on science
888	investments. Large rivers are typically managed for multiple objectives, including fisheries,
889	multi species, or ecosystem objectives. Management decisions typically require an
890	understanding of how management actions propagate through a river ecosystem. Although the
891	emphasis may be on a biological endpoint (among other objectives), understanding the
892	intermediate steps and the processes linking them, and potential interactions between processes
893	or EEC components, can help formulate effective management strategies; especially as multiple 45

894	objectives compete. For example, to increase recruitment of a fish species, it is not enough to
895	know that increased reservoir flow releases are associated with increased spawning unless we
896	understand how much flow, when, for how long, and at what temperature, are needed to optimize
897	other EECs to increase survival and recruitment.
898	The knowledge needed for effective management of large rivers can be gained by
899	monitoring intermediate endpoints along the cascade, but the type of information and costs vary
900	widely. Costs for monitoring Tier 1 EECs can be high but some programs are already in place.
901	For example, large rivers are likely to have monitoring infrastructure installed for Tier 1
902	monitoring of discharge and temperature regimes, with varying potential for monitoring
903	sediment transport and water quality. Investment at Tier 2 may emphasize physical processes and
904	habitats that can be measured at relatively low cost, assuming that habitats are adequately
905	defined based on biological criteria. In larger rivers, Tier 2 habitat assessments can be more cost
906	effective compared to smaller rivers because they can rely on automated data collection through
907	hydroacoustics and remote sensing [108].[101]. As discussed above, habitat assessments have
908	value only to the extent that they are based on well-defined biological requirements; it is notable
909	that some large-river management efforts have found that relatively simple habitat models are
910	useful to predict biological responses [109102]. At Tier 3, costs can increase substantially
911	because of structural uncertainties (i.e., which life stages, which species are most important to
912	monitor) and because of the inherent uncertainties of monitoring fish in large river systems
913	where detection probability can be low and highly variable [36, 110103]. The CMs presented
914	here can be used to assess which pathways are thought to be most important, and which
915	components of EECs would yield the best return on investment for decision making. Generally,
916	the cost of monitoring increases from Tier 1 to Tier 3 in the CM hierarchy; at the same time, the 46

917	relevance of information to decision making is typically greater for biological responses depicted
918	<u>in Tier 3 [104].</u>
919	Because both costs and information benefits increase from Tier 1 to Tier 3 in the CM
920	hierarchy, it is difficult to generalize about where the benefit:cost ratio would be optimized.
921	Indeed, as discussed by Jacobson and Berkley [30], the decision about where in the hierarchy
922	monitoring resources would get the highest return on investment may depend more directly on
923	managers' and stakeholders' perceptions about risks of acting with incomplete information. For
924	example, the details of how a fish's reproductive strategy depends on the nuances of a seasonal
925	hydrograph may not be known, but stakeholders may believe strongly that the natural
926	hydrograph was functional for the species and therefore monitoring of the flow regime will have
927	the highest return on investment and, by extension, restoration of the flow regime is likely to
928	have the most positive effects. On the other hand, in systems where stakeholders opinions are
929	divided or socio-economic values would be compromised by a return to a natural flow regime,
930	managers may be required to demonstrate more precisely how elements of the flow regime
931	propagate to species' benefits [105]. Thus, once information needs are identified and there is an
932	assessment of the availability of data identified as information needs, there needs to be a process
933	whereby the costs of collecting the information need to be placed in a socioeconomic context
934	<u>(e.g., see [30]).</u>
935	The development of the CMs described in this manuscript can be a first step in
936	application of structured decision-making (SDM) and its iterative form-adaptive management
937	(AM) processes [106, 107]. Structured decision-making is a stakeholder driven process by which
938	a problem can be defined with conceptual models and decomposed into decision components
1	

939 <u>that include the problem context, stakeholder objectives, potential management actions,</u> 47

940	consequences of those actions on the objectives, and trade-offs related to different decisions
941	(actions) [107-109]. One primary focus of SDM is the identification of uncertainties such as
942	those identified in the CMs for the case studies in this paper [110]. Quantification of the
943	influence of decision relevant uncertainties can be modeled using sensitivity analysis and other
944	techniques and ranked [107, 109, 111]. In addition, the quantitative techniques available to assist
945	in solving complex ecological problems are robust and range in complexity from consequences
946	tables to Bayesian models to dynamic optimization models [107-109, 112, 113]. The SDM
947	process is often used as the set-up phase for adaptive management which includes monitoring
948	over time to reduce uncertainty related to how management will influence important outcomes
949	(e.g. fish population status; [109, 112].
950	The CM may also help to identify which processes or components are amenable to a field
951	monitoring effort and which are more aptly addressed through laboratory or mesocosm
952	experiments. For example, if it is hypothesized that the condition of age 1+ Arkansas Shiners is a
953	critical determining factor in egg quality or production (Fig 54), it could be determined that the
954	best approach to developing a quantitative relation between condition and eggs is through a
955	controlled laboratory experiment rather than field-based monitoring. The CM helps to visualize
956	where different types of information may be applied within a decision-making framework.
957	A large-river CM may also serve as a precursor to computational ecological or population
958	models [30][30]. Similar questions about how monitoring and other science efforts should be
959	distributed among EECs and processes can be addressed iteratively by carrying out sensitivity
960	analyses in a modeling frameworkIndeed, given substantial uncertainties associated with
961	monitoring data, computation modeling can be considered a necessary component of large-river
962	monitoring and evaluation systems [106, 111<u>99, 114</u>]. 48

963 Conclusions

964	We found the process of conceptualizing the relationships between and within EECs
965	fostered a critical assessment of what we know about factors affecting the management endpoint
966	being considered. By visualizing how EEC drivers directly and indirectly affect management
967	endpoints, our CM identified critical information gaps and uncertainties that, if resolved, could
968	improve our understanding of how to best meet management objectives. The process of
969	conceptualizing the EEC relationships affecting fish in large rivers could help to structure, or
970	restructure, monitoring programs around scientifically sound monitoring questions, promote the
971	selection of relevant ecological indicators that characterize resource condition or management
972	outcomes, and facilitate communication and information sharing within and between
973	organizations managing or researching management endpoints. Ultimately, understanding the
974	mechanisms by which EECs influence large-river fishes will improve the effectiveness of
975	restoration and management actions.
976	As shown with our case studies, our CM is flexible and applicable to a wide range of
977	river systems with different anthropogenic drivers and management objectives. We feel our CM
978	provides a generic structure that scientists can adapt to their management goals and needs. By
979	not being overly prescriptive, for example, with respect to the components of the Tier 2 and 3
980	EEC components, scientists can adapt the CM to different biological communities and
981	management endpoints. By doing so, we feel that users have the flexibility to place their
982	management questions in the context of EECs that are specific to their large-river system.
983	Although the case studies addressed management issues that were river or basin specific,
984	there were similarities relative to information needs and data availability. For example, in most

systems information on river discharge and water temperature were needed and available.
Conversely, information regarding trophic relationships and the habitat requirements of larval
fishes were generally lacking. This result suggests that there may be a common need for a better
understanding of certain factors across large-river systems.

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1377

1379 Supporting Information Captions

- 1380 Fig S1. Discharge (m³/s) patterns over time in the Canadian River, Oklahoma near Canadian, TX
- 1381 (USGS 2019; USGS Gage 07228000).
- 1382 Fig S2. Hydrograph showing pre- and post-Glen Canyon Dam closure in 1964 (dashed line)
- 1383 mean monthly discharge (m^3/s) , which transitions from seasonally stochastic to a more
- 1384 homogeneous regime focusing on anthropogenic interests.
- 1385 Fig S3. The spatial and temporal scales of the management goal, the scientific inferences needed
- 1386 to inform the management goal, and that data collection needs to occur to support the inferences
- 1387 <u>for monitoring information needs identified as requiring additional data in the case study</u>
- 1388 addressing Humpback Chub recruitment in the Colorado River between Glen Canyon Dam and
- 1389 <u>Lake Mead, Arizona (see Table S2 for additional detail). A:Tier 2 EEC= larval Humpback Chub</u>
- 1390 <u>habitat; Stressors=habitat fragmentation, turbidity and Tier 2 EEC= Humpback Chub spawning</u>
- 1391 <u>habitat; Stressor= water temperature and Tier 3 EEC=insect production; Stressor=benthic</u>
- 1392 <u>macroinvertebrate habitat quantity and quality and Tier 3 EECs=all; Inter-tier interaction=trophic</u>
- 1393 level interactions; B:Tier 3 EEC=larval Humpback Chub production; Stressor= larval Humpback
- 1394 <u>chub habitat quantity and quality; C: Tier 3 EEC=Humpback Chub egg quality and production;</u>
- 1395 <u>Stressors= Humpback Chub spawning habitat quantity and quality; D:Larval Humpback Chub</u>
- 1396 production; Inter-tier interaction=mortality of Humpback Chub eggs; E:Tier 3 EEC= larval
- 1397 <u>Humpback Chub production; Stressor=predation by invasive species and Tier 3 EEC=Primary</u>
- 1398 production; Stressor=nutrient flux; F:Tier 1 EEC= biogeochemistry/thermodynamics; Inter-tier
- 1399 interaction=sediment adsorption of contaminants and nutrients; G:Tier 3 EEC=Humpback Chub
- 1400 <u>age-0 recruitment; Inter-tier interaction=mortality of larval Humpback Chub.</u>
 - 60

1401	Fig S4. Proportion of total annual Columbia River discharge at The Dalles, OR occurring in the	
1402	month of June from 1879 to 2015.	
1403	Fig S5. The spatial and temporal scales of the management goal, the scientific inferences needed	
1404	to inform the management goal, and that data collection needs to occur to support the inferences	
1405	for monitoring information needs identified as requiring additional data in the case study	
1406	addressing White Sturgeon recruitment in the Columbia River (see Table S3 for additional	
1407	detail). <u>S4A:Tier 1 EEC= biogeochemistry/thermodynamic; Stressor= altered biogeochemical</u>	
1408	regime and Tier 2 EEC= benthic macroinvertebrate habitat; Stressors=channel stability, sediment	
1409	deposition, fragmentation and Tier 2 EEC=Larval White Sturgeon habitat; Stressors=habitat	
1410	fragmentation sediment deposition, water temperature and Tier 2 EEC=White Sturgeon	
1411	spawning habitat; Stressors=contaminants, sediment deposition and Tier 3 EEC=White Sturgeon	
1412	egg quality and production; Stressor=predation by invasive species and Tier 3 EEC=White	
1413	Sturgeon larvae production; Stressor=predation by invasive species and Tier 3 EEC=benthic	
1414	macroinvertebrate production; Stressor=benthic macroinvertebrate habitat quantity and quality;	
1415	B:Tier 1 EEC=sediment transport; Stressors=altered sediment regime, altered hydraulic regime	
1416	and Tier 1 EECs=channel morphology/hydraulics, sediment transport; Stressor=altered hydraulic	
1417	regime and Tier 3 EEC=primary production; Stressor=nutrient fluxes; C:Tier 1 EEC=channel	
1418	morphology/hydraulics, sediment transport; Stressor=altered hydraulic regime and Tier 1 EEC =	
1419	channel morphology/hydraulics, sediment transport; Inter-tier interaction=sediment transport	
1420	dynamics; D:Tier 1 EEC=channel morphology/hydraulics, sediment transport; Stressor=altered	
1421	hydraulic regime and Tier 1 EEC=channel morphology/hydraulics, sediment transport; Inter-tier	
1422	interaction=sediment transport dynamics; E:Tier 1 EEC=biogeochemistry/thermodynamics;	
1423	Inter-tier interaction=sediment adsorption of contaminants and nutrients; F: Tier 3 EEC=White 61	

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1424	Sturgeon larvae production; Stressors=larval White Sturgeon habitat quantity and quality and	
1425	Tier 3 EEC=White Sturgeon larvae production, age-0 White Sturgeon recruitment; Inter-tier	
1426	interaction=mortality and White Sturgeon egg quality and production; Inter-tier	
1427	interaction=predation of White Sturgeon eggs by native fish and Tier 3 EECs=White Sturgeon	
1428	larvae production; Inter-tier interaction=predation of White Sturgeon larvae by native fish and	
1429	Tier 3 EECs=all; Inter-tier interactions=trophic level interactions.	
1430	Fig S6. Largemouth bass (Micropterus salmoides) data from the Pool 13 of the Upper	
 1431	Mississippi River (A, B) and the La Grange Pool of the Illinois River (C, D). The two river	
1432	reaches are roughly the same latitude, but the La Grange Pool is more limited in overwintering	
1433	habitat. Population abundance is presented in panels A and C where each point is an individual	
1434	fish cumulatively caught with standardized day time electrofishing (Ratcliff et al. 2014) in a	
1435	specific year. The dashed triangle highlights 'missing' >400 mm size classes since 2000 in the	
1436	La Grange Pool. Population size structure is indexed by proportional stock density (PSD) is	
1437	presented in panels B and D with the dashed line showing trends in the largest size classes over	
1438	time. Data and methodology were downloaded from the publicly available databases via the	
1439	Upper Mississippi River Restoration's Long Term Resource Monitoring Graphical Fish Browser	
1440	[<u>112</u>]15].	Field Code Changed
1441	Fig S7. The spatial and temporal scales of the management goal, the scientific inferences needed	
1442	to inform the management goal, and that data collection needs to occur to support the inferences	
1443	for monitoring information needs identified as requiring additional data in the case study	
1444	addressing native fish biodiversity and habitat diversity in the Mississippi and Illinois rivers (see	Formatted: Font color: Black
1445	Table S4 for additional detail). A:Tier 1 EEC=sediment transport; Stressor=altered hydraulic	
1446	regime and Tier 1 EEC=biogeochemistry/thermodynamics; Stressor=altered biogeochemical 62	

1447	regime and Tier 1 EEC=biogeochemistry/thermodynamics; Inter-tier interaction=sediment
1448	adsorption of contaminants and nutrients; B: Tier 2 EEC=adult native fish overwintering habitat;
1449	Stressors=water velocity, water temperature, dissolved oxygen, sediment deposition and Tier 2
1450	EEC=juvenile native fish habitat; Stressors=water depth, water velocity, water temperature,
1451	dissolved oxygen, contaminants, sediment deposition and Tier 2 EEC=native fish spawning
1452	habitat; Stressors=water depth, water velocity, habitat fragmentation, sediment deposition, water
1453	temperature, dissolved oxygen, contaminants and Tier 3 EEC=adult and juvenile native fish
1454	recruitment; Inter-tier interaction=mortality and Tier 3 EEC=all; Stressors=invasive species and
1455	Tier 3 EEC=all; Inter-tier interaction=trophic level interactions; C:Tier 1 EEC=channel
1456	morphology/hydraulics; Inter-tier interaction=channel forming processes; D:Tier 1 EEC=channel
1457	morphology/hydraulics, sediment transport; Inter-tier interaction=sediment transport dynamics;
1458	E:Tier 3 EEC=adult native fish recruitment; Stressor=adult native fish overwintering habitat
1459	quantity and quality and Tier 3 EEC=juvenile native fish recruitment; Stressor=juvenile native
1460	fish habitat quantity and quality and Tier 3 EEC=native fish egg quality and production;
1461	Stressor=spawning habitat quantity and quality.
1462	Table S1. Summary of information needs identified in the Conceptual Model describing factors
1463	affecting the recruitment of the Arkansas River Shiner in the South Canadian River, OK (Fig 4;
1464	this publication) by Essential Ecosystem Characteristic (EEC) Tier, EEC, and stressor or within
1465	Tier interactions and an assessment of the status of existing information that could be used to
1466	address the information needs.
1467	Table S2. Summary of information needs identified in the Conceptual Model describing factors
1468	affecting the recruitment of the Humpback Chub in the Colorado River, Arizona (Fig 6; this

publication) by Essential Ecosystem Characteristic (EEC) Tier, EEC, and stressor or inter-tier63

1470	interactions and an assessment of the status of existing information that could be used to address	
1471	the information needs.	
1472	Table S3. Summary of information needs identified in the Conceptual Model describing factors	
1473	affecting the recruitment of age-0 White Sturgeon in the Columbia River (Fig <u>87</u> ; this	
1474	publication), by Essential Ecosystem Characteristic (EEC) Tier, EEC, and stressor or inter-tier	
1475	interactions and an assessment of the status of existing information that could be used to address	
1476	the information needs.	
1477	Table S4. Summary of information needs identified in the Conceptual Model describing factors	
1478	affecting the restoration and maintenance of native fish biodiversity and habitat quantity and	
1479	quality in the Upper Mississippi and Illinois rivers (Fig 108; this publication), by Essential	
1480	Ecosystem Characteristic (EEC) Tier, EEC, and stressor or inter-tier interactions and an	
1481	assessment of the status of existing information that could be used to address the information	
1482	needs.	
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Reviewer 1 wonders about the utility of such conceptual models in the more realistic case of multiple species contexts. How could the model be expanded in this regard, besides the case of simple richness measures? In addition, how could the model help in actually prioritising or ranking the variables or interactions identified? Finally, reviewer 1 also noticed that an important stressor related to fragmentation and connectivity is only marginally discussed and included in the model. This is critical for meta-population dynamics and should be given more emphasis.

Reviewer 2, similarly raises the critical issue of the multi-species context, and how the needs of different species could be simultaneously identified. Therefore, I advice to expand the Discussion in this regard, eventually acknowledging limitations and suggesting future research needs. Reviewer 2 also wonders how such models could effectively guide restoration and decision-making beyond monitoring needs; if the relative importance of data gaps and interactions is not quantified (e.g. via a cost-benefit analysis), how could it help prioritise the focus of monitoring and action?

Besides carefully responding to each reviewers' comments and modify the manuscript accordingly, I also suggest to simplify the manuscript, which feels rather long. Perhaps some of the background information from each case study could be included as supplementary or shortened. Also, caption from Fig.5 (and sister-figures) is rather hard to digest for the reader. I wonder if this could be simplified as well.

Response: Thank you for the opportunity to revise our manuscript. We have tried to address comments of Reviewer 1 and 2 below and in the revised manuscript. To shorten and simplify the manuscript, we relocated some of the contextual information from the case studies and moved it to an Appendix in Supplemental Information. With respect to Fig 5 and sister figures, we have discussed trying to simplify the figure captions but have not come up with a good solution. There is a lot of information contained within the figures and feel that further generalizations would not be clarifying. We do, however, acknowledge that the figure caption format is awkwardly long. What we propose is

that we retain Fig 5 in the main body of the text as an example, and then move subsequent sister figures to Supplemental Materials. Please let us know if this satisfies your and the reviewer's comments to reduce the length of the manuscript.

Comments to the Author

5. Review Comments to the Author

Reviewer #1: In this study a conceptual model is used to aid the development of best practices of large river monitoring programs. The model was developed based on former scientific works and during scientist's workshop negotiations. Case study applications prove that the application of this complex conceptual model can be useful to identify critical information gaps, which can then be used to develop management and monitoring objectives.

I like the approach of developing such conceptual models, which can reveal information gaps, and think that the model in general can be useful to adopt across large river systems with some refinements and local adaptations. Consequently, I believe showing such an approach can provide useful information for the readers.

 What I lack is to show more convincingly how such complex models can be used for multispecies systems, where not only the requirements of a single species is evaluated, which in fact the more realistic situation. How can individual species level models be put together to provide meaningful information for management? It would be useful to discuss this in more detail in the Discussion section.

Response: Thank you for this comment. We have added text in the discussion that describes how the CM could contribute to our understanding of the need for a multi-species context in management activities. Please see: L645-691

2) Also a critical issue which should be briefly discussed is how the identified critical target variables should be prioritized, especially in a multispecies systems, where several variables will appear. Development of this section could convince the reader and could clearly show the applicability of such conceptual models by management. Response: Thank you for this insight. We have provided language in the discussion that describes how the CM could be used to identify critical target variables in a multispecies context. Please see: L650-668

3) Although channel morphology/hydraulics may contain fragmentation/connectivity issues this should be made more clear in the material, because this is one of the most critical issue, which determine fish (meta)population or metacommunity dynamics. In fact fragmentation is often used as one of the most critical variable of anthropogenic drivers and as such is a critically important target to mitigate by management. However, it does not appear either in Fig 2 or Fig 3, but only on the case study figures belonging to morphology/hydraulics TIER1 components.

> Response: Thank you. We agree with your assessment. We have further emphasized the importance of fragmentation by specifically mentioning it in the manuscript section describing the CM form. Please see: L223-234. Also, in Table 2, habitat fragmentation is emphasized as affecting multiple facets of the CM in multiple river systems and is mentioned in the text describing Table 2. This result indicates the need to better study the effects of habitat fragmentation on multiple biotic components.

Reviewer #2: Dear Editor,

This study demonstrates how a conceptual model can be used to identifying knowledge gaps in the mechanisms by which Essential Ecosystem Characteristics influence large-river fish species in the USA. These gaps should then be filled to improve the effectiveness of restoration and management.

I agree on the value of these conceptual models to identify knowledge gaps and inform decisions on what to monitor to fill them and allow a better understanding of the system and, therefore, enhance our capacity to manage them adequately. However, I disagree with some of the arguments:

 The conceptual model represents potential interactions across different structural element of the river system but does not allow quantitative evaluations of strength of those interaction. As such, the value of is conceptual model is limited to identifying knowledge gaps and cannot be used to evaluate the relative importance of each interaction. Therefore, this conceptual model should only be used for identifying knowledge gaps and not for decision-making, as argued (see L771-773), beyond monitoring. Response: This is an excellent point and we have removed the statement in L771-773 and elaborated on the considerations that need to be accounted for, and the difficulties with, assessing benefit:cost ratios. Please see: L692-726. Also, you are correct that the CM does not provide quantitative evaluations of the strength of the relations. We do acknowledge this and suggest that the CM could provide a basis for developing quantitative assessments in L749-754 and have added language that describes how the CMs could be the basis for Structural Decision Making and Adaptive Management processes (see L727-741).

2) The conceptual model lacks a cost analysis to evaluate the most efficient way of filling knowledge gaps. Some of the gaps might be more difficult/ costly or even feasible to fill. Without such analysis we can only identify the gaps but cannot prioritise where to focus monitoring on a cost-effective way and just confirm where gaps exist.

> Response: Thank you for this comment. In addition to addressing the benefit:cost issue above, we have provided language in the discussion that describes how the CM could be used to identify critical target variables in a multispecies context but that there are critical uncertainties that need to be considered. See: L650-668. We agree with you about prioritizing based on cost effectiveness but respectfully suggest that the CM could provide information that would suggest where to focus monitoring effort.

3) Three of the case studies present conceptual models for individual species. While I see the value of developing these conceptual models for charismatic endangered species, I wonder how feasible/ useful it would be this method when facing management needs for many species simultaneously. One of the case studies does present a conceptual model for the full fish community, but focused on diversity, rather than individual species, so no information of particular species issues are addressed. Would it be feasible to elaborate a conceptual model that addressed all individual species needs/ issues simultaneously? This would allow identifying knowledge gaps common to multiple species simultaneously.

Response: Thank you for this comment. We have added text in the discussion that describes how the CM could contribute to our understanding of the need for a multi-species context in management activities. Please see: Please see: L645-691 4) Minor comments:

- L304 & 369. What does ATV stand for?

Response: Thank you for bringing this to our attention. ATV stands for all-terrain vehicle. We have removed the acronym from the revision.

- The manuscript is quite long, especially because of the description of each case study. It would be good to present the information of these case studies in a more synthetic way (maybe on a table?).

Response: Thank you for the comment. Per your and the Associate Editor's recommendation we have pulled out some of the contextual information from the case studies and moved the information to an Appendix in Supplemental Information. We have also moved three figures and associated captions to the supplemental information section.